

EVALUATION OF MARINE MICROBIOTA WITH BIOREMEDIATION POTENTIAL FOR HYDROCARBONS

Evaluación de la microbiota marina con potencial biorremediador por hidrocarburos

Aiden ZÚÑIGA VILLALBA, Sonia ROHRA BENÍTEZ,
Olivia PÉREZ GÓMEZ and Silvana Teresa TAPIA PANIAGUA*

Departamento de Microbiología, Facultad de Ciencias, Universidad de Málaga, Blvr. Louis Pasteur, 31, Málaga, 29010, España.

*Author for correspondence: stapia@uma.es

(Received: December 2024; accepted: May 2025)

Keywords: bioremediation, pentane, benzene, marine pollution, bacteria, marine biodiversity.

ABSTRACT

The release of hydrocarbons resulting from human activities is one of the main causes of global marine pollution. The increasing tourism in Mediterranean regions such as the Costa del Sol (Málaga, Spain) poses a risk to the environmental health of these ecosystems due to the intensified use of motorboats. This situation calls for bioremediation enhanced by bioaugmentation and biostimulation. In this study, cultivable bacteria were isolated from water samples contaminated with hydrocarbons at Muelle Uno (Málaga, Spain). Subsequently, biostimulation processes were applied under different experimental conditions in minimal and complex media, with and without the presence of benzene or pentane, to evaluate their influence on bacterial abundances. The results showed that hydrocarbons extended the bacterial growth even in complex media and highlighted the essential role of nitrogen in bacterial proliferation. Molecular identification revealed the presence of *Pseudoalteromonas* sp. and *Vibrio* sp. in the seawater samples. However, benzene favored the growth of *Bacillus* sp., while pentane facilitated the proliferation of *Vibrio* sp. Massive sequencing results confirmed the dominance of *Vibrio* sp. under pentane conditions, whereas *Pseudomonas* sp. was predominant in benzene conditions. In conclusion, this study demonstrates the significant impact of hydrocarbon pollution on marine microbial communities, particularly affecting key bacterial genera such as *Vibrio* sp., *Bacillus* sp., and *Pseudomonas* sp. These findings provide a robust foundation for future research aimed at optimizing the use of these bacteria in bioremediation processes for hydrocarbon-contaminated marine environments.

Palabras clave: biorremediación, pentano, benceno, contaminación marina, bacteria, biodiversidad marina.

RESUMEN

La liberación de hidrocarburos derivada de actividades humanas es una de las principales causas de contaminación marina a nivel global. El auge turístico en regiones mediterráneas como la Costa del Sol (Málaga, España) pone en riesgo la salud medioambiental de estos ecosistemas debido al mayor uso de embarcaciones de motor. Por ello es necesario llevar a cabo procesos de biorremediación, y la mejora de ésta, mediante

técnicas de bioaumentación y bioestimulación. En este estudio se aislaron bacterias cultivables procedentes de muestras de agua contaminadas por carburantes en el Muelle Uno (Málaga, España). Posteriormente, tras procesos de bioestimulación en diferentes condiciones experimentales en medios mínimos y complejos, con y sin la presencia de benceno o pentano, se evaluó la influencia en las abundancias bacterianas. Los resultados mostraron que los hidrocarburos prolongaron el periodo de crecimiento bacteriano, incluso en medios complejos, y que la presencia de nitrógeno fue esencial para la proliferación bacteriana. La identificación molecular mostró la presencia de *Pseudoalteromonas* sp. y *Vibrio* sp. en las muestras de agua de mar. No obstante, la presencia de benceno promovió la aparición de *Bacillus* sp., y la de pentano, la de *Vibrio* sp. Los resultados obtenidos mediante secuenciación masiva mostraron nuevamente la presencia de *Vibrio* sp. en las condiciones con pentano, mientras que en las condiciones con benceno predominó *Pseudomonas* sp. En conclusión, este estudio resalta cómo la contaminación por hidrocarburos afecta significativamente a las comunidades microbianas marinas, impactando a géneros bacterianos clave como *Vibrio* sp., *Bacillus* sp. y *Pseudomonas* sp. Estos hallazgos proporcionan unas especies candidatas prometedoras para estudios futuros sobre degradación de hidrocarburos.

INTRODUCTION

The aquatic environment represents about 71% of the Earth's surface (Seymour 2014). Life in these environments is conditioned by different factors such as temperature, salinity, the presence of sunlight, or the amount of oxygen dissolved in the water that determine the establishment of different biological communities. In the case of microorganisms, these represent around 70% of marine biomass (Bar-On and Milo 2019) and have a great metabolic diversity, especially in prokaryotic organisms, whose participation in biogeochemical cycles is essential for the correct functioning of the planet (Delong 2009, Seymour 2014, Deutschmann et al. 2023). Recently, the analysis of marine microbiota has been promoted to better understand its role in ecological processes and to investigate its possible biotechnological applications, such as the production of new bioactive compounds or the remediation of contaminated environments (Dash et al. 2013, Acosta-González and Marqués 2016, Hamdan et al. 2019, 2023). However, it is still less characterized than the terrestrial microbiota, especially that present on the ocean floor (Sanz-Sáez et al. 2020). Oil spills, as well as their derivatives, are one of the main causes of pollution in these environments, such as the destruction of the Deepwater Horizon oil platform in the Gulf of Mexico (2010), or the Prestige catastrophe in Galicia, Spain in 2002, in which thousands of tons of crude oil were released into the sea (Acosta-González and Marqués 2016). Similar environmental disasters

have also been recorded in the Mediterranean Sea, such as the release of large quantities of fuel off the coast of Jiyeb (Lebanon) in 2006 (Little et al. 2021, Hamdan et al. 2023).

The Mediterranean Sea is a continental sea that connects to the Atlantic Ocean through the Strait of Gibraltar and to the Indian Ocean through the Suez Canal. Surrounded by the continents of Europe, Africa, and Asia, this sea, with an area of approximately 2.5 million km² and a length of 3860 km, is the second-largest inland sea on the planet, surpassed only by the Caribbean Sea. Its pollution is not only due to occasional accidents, but also to commercial and tourist activity. In the case of coastal areas of the Mediterranean Sea, such as Costa del Sol (Málaga and Granada, Spain), a large part of this pollution is due to the increase in tourist and recreational activities, due to the use of motor boats (cruises, yachts, sport fishing boats, etc.). Malaga is one of the main tourist destinations on the Iberian Peninsula, welcoming approximately 2.6 million travelers between June and August 2023, with tourism increasing by 4.8% since 2022 (CSM 2023, DPM 2024). In addition, in relation to tourism revenue, an increase of 10.1% was estimated from 2022. Currently, the province of Malaga has 11 marinas along its coast (**Fig. 1**).

One of the best known is Muelle Uno (Málaga, Spain), located in the heart of the city, which stands out for its high presence of motor-boats for tourist and recreational purposes (PM 2024) As of 2021, the economic development of this city encouraged the construction of port infrastructures capable of



Fig. 1. Location of the marinas in the province of Malaga.

accommodating the docking of luxury mega yachts and commercial cruise ships on this promenade. This area recorded an average temperature of 24.8 °C on the Malaga coast in August 2023, which favors the proliferation of microorganisms facilitated by the presence of hydrocarbons.

This situation is similar in other parts of the Mediterranean Sea, where biological diversity is highly threatened by these fuels, since their chemical composition, which consists mainly of a mixture of aliphatic and aromatic hydrocarbons, can have harmful effects (Chandra et al. 2013, Varjani 2017). Despite being a minority, monocyclic aromatic hydrocarbons (MAHs) such as benzene, toluene, ethylbenzene, and xylene (BTEX), and polycyclic aromatic hydrocarbons (PAHs) are the main causes of these environmental catastrophes (Varjani and Upasani 2017, Honda and Suzuki 2020, Goveas et al. 2022). Acute or chronic exposure to these discharges significantly alters ecosystems and poses a risk to living organisms (White et al. 2017, Arockiaraj and Kankara 2018, Cariello et al. 2020).

For this reason, the scientific community is investing great human and economic efforts in solving these problems. One of these strategies is based on remediation processes, which are addressed through physical, chemical, or biological methods (Varjani 2017). Among all of them, the biological method, commonly known as bioremediation, has proven to

be the best alternative to eliminate hydrocarbons in these media, due to its low cost, low environmental impact, and high efficiency (Hamdan et al. 2019, Gao et al. 2023). Furthermore, the application of techniques such as biostimulation, bioaugmentation and the use of biosurfactants has further improved the efficiency of decontamination in these habitats (Fig. 2).

Biostimulation increases biodegradation rates by adding essential nutrients, such as nitrogen and phosphorus, to the medium (Luo et al. 2024), while bioaugmentation leverages the degradative capacities of different microorganisms to eliminate all compounds of the pollutant in question (Luo et al. 2024). Furthermore, the use of bacterial biosurfactants increases the availability of hydrocarbons in the marine environment as a potential energy and carbon source (Varjani and Upasani 2017, Goveas et al. 2022).

Hydrocarbon-oxidizing bacteria, or hydrocarbonoclastic bacteria, are able to metabolize these compounds via aerobic or anaerobic pathways. The degradation of aliphatic and aromatic hydrocarbons follows different metabolic pathways; therefore, different bacterial genera have genes encoding the enzymes required for each pathway. Some strains of the genera *Alcanivorax* sp. and *Vibrio* sp. have been shown to be capable of oxidizing aliphatic hydrocarbons (Imron et al. 2019, Sinha et al. 2021), while other strains of the genera *Acidovorax* sp.,

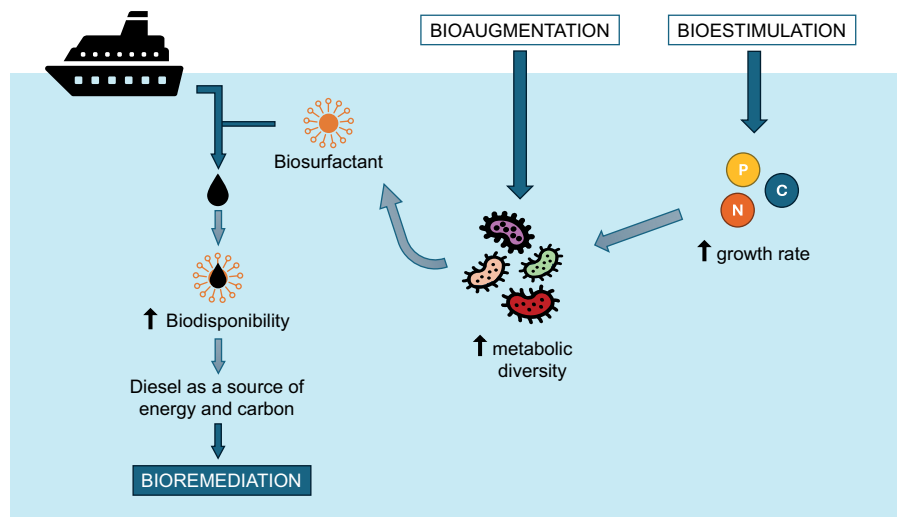


Fig. 2. Techniques used to improve the efficiency of bioremediation in marine environments.

Bacillus sp., *Burkholderia* sp., and *Sphingomonas* sp. are capable of degrading aromatic hydrocarbons present in petroleum (Seo et al. 2009).

Other studies also highlight the degradative potential of genera such as *Petrimonas* sp. and *Desulfotribrio* sp., which appear to be able to metabolize PAHs via anaerobic pathways on the ocean floor (Qian et al. 2020).

For this reason, the marine microbiota of waters contaminated by hydrocarbons from motor boats should be studied, as it may harbor hydrocarbonoclastic bacteria that could enhance the degradation of the pollutants present. In addition, biostimulation can promote the growth of specific cultivable isolates with the capacity to metabolize hydrocarbons such as benzene and pentane.

Therefore, in this work we will make an approach by characterizing the marine microbiota present in waters contaminated by hydrocarbons in Muelle Uno (Málaga, Spain), using molecular and culture techniques. Subsequently, we will evaluate the effect of biostimulation on the growth and composition of the bacterial microbiota under controlled experimental conditions with and without the presence of hydrocarbons (benzene and pentane).

MATERIALS AND METHODS

Marine samples

The sampled water comes from an area of Muelle Uno that contained a fuel slick, close to the docking area for recreational motorboats (coordinates

36.717713, -4.413229) (Fig. 3). The water temperature was 15 °C (ST 2024). A total of 1 L of water was sampled from three different points. They were then transported to the laboratory at a cold temperature for processing.

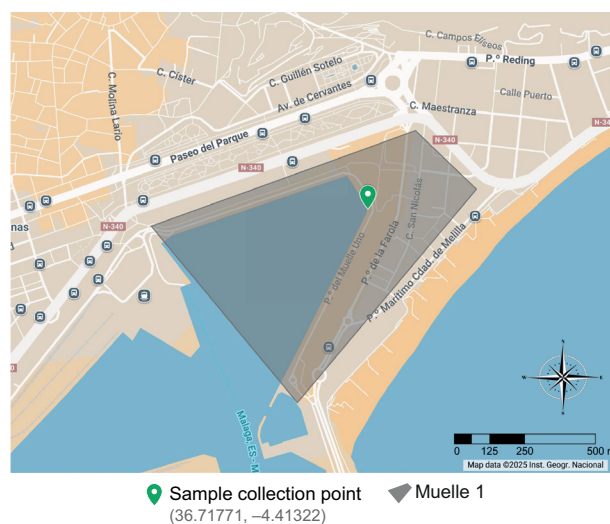


Fig. 3. Sample collection point. The delimited area corresponds to the entire Muelle Uno complex.

Microbiological characterization of water and isolation of bacteria with potential biotechnological interest

In order to promote bacterial growth in the sampled water, and to select bacteria of biotechnological interest, the following bacterial culture

media were prepared: nutrient-rich medium, tryptone soy broth (TSB, Oxoid, Hampshire, UK), and minimum nutrients medium (M9, Sigma-Aldrich, San Luis Missouri, USA), both supplemented with 15% NaCl (Honeywell, North Carolina, USA) to select marine bacteria. All culture media had a volume of 0.5 L, and different experimental conditions were performed (**Table I**). The hydrocarbons used in this study were benzene and pentane (ITW Reagents, Monza, Italy). The corresponding cultures had concentrations of 25 mL of benzene or pentane in 0.5 L of medium to exert greater pressure and select bacteria resistant to these hydrocarbons. To each condition, 50 mL of water from the samples mentioned in the previous section were added. In addition, the media in which there was no bacterial growth were supplemented with 5 mL of sterile 20% casamino acids (BD, New Jersey, USA) and/or 5 mL of sterile 50% D-glucose (PanReac AppliChem, Darmstadt, Germany). All media were tested in duplicate.

TABLE I. EXPERIMENTAL CONDITIONS USED IN THIS STUDY INCLUDED TRYPTONE SOY BROTH (TSB) AND MINIMAL MEDIA (M9)* AS CONTROLS (C_TSB AND C_M9), AS WELL AS VERSIONS OF EACH SUPPLEMENTED WITH EITHER BENZENE (TSB_B, M9_B) OR PENTANE (TSB_P, M9_P). ALL MEDIA WERE SUPPLEMENTED WITH 15% NaCl.

Condition	Culture medium
C_TSB	Tryptone soy broth
TSB_B	Tryptone soy broth + benzene
TSB_P	Tryptone soy broth + pentane
C_M9	Minimal media (M9)
M9_B	M9 + benzene
M9_P	M9 + pentane

*TSB and M9 are described in detail in section 2.

All media were incubated at 20 °C throughout the experiment. The incubation period for each condition ranged from the addition of water until the appearance of turbidity, or up to a maximum of 26 days. In order to recover and isolate culturable bacteria, different tryptone soy agar plates (TSA, Oxoid) and supplemented with 15% NaCl, were streaked from two different types of samples: seawater samples and samples from media (TSB or M9) with added benzene (B) or pentane (P), designated as TSB_B, M9_B, TSB_P, M9_P. One hundred microliters of each sample at different

dilutions were spread on plates and incubated for up to 72 h at 20 °C.

After the incubation period ended, visually distinct colonies were isolated in pure culture. The strains were then stored in TSB + 15% NaCl + 20% glycerol at –80 °C until analysis. For identification, DNA was extracted from each colony using the GeneJet DNA Purification Kit (Thermo Fisher Scientific, Massachusetts, USA) following the manufacturer's protocol. The 16S rRNA region was amplified by PCR (Muyzer et al. 1998) and subsequently sent to Macrogen for sequencing. These sequences were identified by comparing them with the sequences available in the Genbank database (NCBI).

Taxonomic characterization after biostimulation

The appearance of turbidity in the media marked the end of its incubation period. From this moment, 1 mL of the surface of each medium was collected for the extraction of its total genomic DNA as described in the previous section. The quantification of the extracted DNA was performed using fluorometric methods, following the protocol of the commercial Qubit dsDNA HS Assay Kit (Thermo Fisher Scientific, Waltham, Massachusetts, USA).

All extracted DNA was stored at –20 °C for further processing. DNA samples from each experimental condition were sent for massive sequencing using new-generation sequencing (NGS) techniques. Total DNA obtained from the media samples was sent for sequencing using the V3-V4 fragment of bacterial 16S rRNA using the Illumina MiSeq technology (Illumina, San Diego, California, USA), using the paired-end sequencing technique (2 × 300 pb). This procedure was carried out by Novogene (Cambridge, United Kingdom) using the primers 5'CCTAYGGGRBGCASCAG3' and 5'GGACTACNNGGGTATCTAAT3'. The analysis was carried out by the Ultrasequencing Service of the Bioinnovation Center of the University of Malaga (SCBI, UMA), using the technology of the Picasso supercomputer to carry out a sample quality control process with the FastQC software. In addition, for the purification of the obtained sequences, a workflow based on the MOTHR software (v. 1.13.0) was used, which allowed the alignment of the paired reads to form the amplicon. The removal of adapters and chimeras from the sequencing was performed using the vsearch software (v. 2.16.0).

From the filtered data, the taxonomic assignment of the representative sequences of each amplicon sequence variant (ASV) was performed to obtain information about the taxa and their relative abundance.

This process was carried out by comparing sequences with the SILVA database (v. 138.1) under 97% similarity and a threshold defined at 0.005%. The taxonomic analysis of the microbiota present in these samples was performed through graphical representations generated using the RStudio software (v. 4.4.0).

The study of bacterial communities in each sample was carried out using the Chao1, Shannon and Simpson indices, which described taxonomic richness, diversity and dominance, respectively. The graphical representation of these results allowed comparisons to be made between the mean α -diversity values of different experimental conditions. Variations in the composition of microbial communities between different samples were measured through β -diversity.

RESULTS

Isolation and characterization of bacterial strains in the presence of pentane and benzene

The incubation period for each condition ranged from the addition of the collected water to the presence of turbidity in the medium, indicative of microbial growth. Not all media developed this appearance, so a maximum incubation period of 26 days was established. In the other media, the incubation time varied according to the experimental condition (**Table II**).

The media corresponding to conditions C_TSB and TSB_B were the first to show turbidity after 24 h of incubation, whereas TSB_P required 48 h to develop turbidity. Conditions C_M9 and M9_P developed turbidity 24 h after supplementation with 20% casamino acids. However, in the M9_B condition, no growth was observed after the addition of

casamino acids or after the subsequent addition of glucose (**Table II**).

From the media that presented turbidity, four colonies corresponding to the media containing benzene (TSB_B and M9_B) (14.29%) and three colonies from the media containing pentane (TSB_P and M9_P) (10.71%) were isolated. In addition, 21 colonies were obtained from the sampled seawater (75%). All of them were identified. Those colonies isolated from seawater belonged to the Proteobacteria phylum and the Gammaproteobacteria class. In all cases, the genera identified were *Vibrio* sp. and *Pseudoalteromonas* sp., with the isolation of a strain of *Vibrio alginolyticus* being noteworthy (**Table III**).

The colonies isolated from benzene conditions (TSB_B and M9_B) belonged entirely to the genus *Bacillus* sp., while those in the pentane groups (TSB_P and M9_P) corresponded mostly to the genus *Vibrio* sp. (**Table III**).

Post-bioestimulation taxonomic analysis of the microbiota

An abundance table was obtained from the sequences obtained. Graphical representations described the abundances of the phyla, families and genera present under the different experimental conditions. In all groups, the estimated Good's coverage was 100%, so that all sequencing was representative of the bacterial groups present in each group.

All M9 media were mainly represented by sequences belonging to the Proteobacteria phylum (> 95%), whereas the TSB media supported the growth of different phyla depending on the experimental condition.

In the TSB control (C_TSB), a similar abundance was described for the phyla Proteobacteria and

TABLE II. INCUBATION PERIOD (DAYS) FROM THE ADDITION OF WATER TO THE MEDIUM UNTIL THE DEVELOPMENT OF TURBIDITY. THE ADDITION OF CASAMINO ACIDS AND/OR GLUCOSE IS INDICATED WITH THE SYMBOL (+), AND THEIR ABSENCE WITH THE SYMBOL (-). TSB AND M9 AS CONTROLS (C_TSB, C_M9), AND EACH MEDIA SUPPLEMENTED WITH BENZENE (TSB_B, M9_B) OR PENTANE (TSB_P, M9_P).

Condition	Incubation time (days)	Casamino acids	Glucose
Tryptone soy broth (C_TSB)	1	-	-
Tryptone soy broth and benzene	1	-	-
Tryptone soy broth and pentene (TSB_P)	2	-	-
Minimal media (C_M9)	6	+	-
Minimal media and benzene (M9_B)	NC	+	+
Minimal media and pentene (M9_P)	3	+	-

TSB: tryptone soy broth; M9: minimal media; NC: absence of growth during the incubation period. TSB and M9 are described in detail in section 2.

TABLE III. COLONIES ISOLATED ON TRYPTONE SOYA AGAR PLATES FROM SEAWATER SAMPLES AND HYDROCARBON CONDITIONS: TSB AND MINIMAL MEDIUM (M9) MEDIA SUPPLEMENTED WITH BENZENE (TSB_B, M9_B) OR PENTANE (TSB_P, M9_P).

Sample	Genus	% of genus per sample
Sea water	<i>Pseudoalteromonas</i> sp.	28.57%
	<i>Vibrio</i> sp.	71.43%
Triptone soy broth + bezene (TSB_B)	<i>Bacillus</i> sp.	100%
Minimal media + benzene (M9_B)	<i>Bacillus</i> sp.	100%
Triptone soy broth + pentane (TSB_P)	<i>Vibrio/Photobacterium</i> sp.	50%
	<i>Vibrio</i> sp.	50%
Minimal media + pentane (M9_P)	<i>Vibrio</i> sp.	100%

Fusobacteriota (> 40%), as well as the presence of the phylum Bacteroidota in a lower proportion (< 5%). However, the addition of benzene in TSB_B caused a considerable increase of the Proteobacteria phylum (> 95%), causing the displacement of Fusobacteriota below 10%, as well as the absence of Bacteroidota in this condition. The TSB media with pentane (TSB_P) were also represented almost entirely by the Proteobacteria phylum (> 95%), displacing the remaining phyla in its control (Fig. 4).

In figure 5, the M9 control (C_M9) is mainly represented by the Vibrionaceae family (> 75%), along with others such as Pseudomonadaceae (> 5%) and Pseudoalteromonadaceae (> 5%) to a lesser extent. In the TSB control (C_TSB), the presence of the families Fusobacteriaceae (> 35%), Pseudoalteromonadaceae (> 25%), and Vibrionaceae (< 20%) was observed. Benzene limited the growth of the Vibrionaceae family in both media

M9_B and TSB_B. Moreover, the presence of benzene increased microbial diversity, with a greater number of bacterial families being represented. Among them, the Pseudomonadaceae family was dominant in both M9_B and TSB_B. This hydrocarbon also promoted the growth of the Sphingomonadaceae and Burkholderiaceae families to a lesser extent, whose relative abundances were not influenced by the media's nutritional composition. The families Rhodobacteraceae (< 10%) and Xanthomonadaceae (< 10%) appeared exclusively in M9_B, whereas Comamonadaceae grew only in TSB_B (< 10%). The presence of pentane increased the abundance of the Vibrionaceae family by over 90%, both in the M9 (M9_P) and TSB (TSB_P) media.

The C_M9 control was mainly represented by the genera *Vibrio* sp. (> 40%) and *Aliivibrio* sp. (> 30%), both belonging to the Vibrionaceae family, together with *Pseudomonas* sp. (< 15%) and

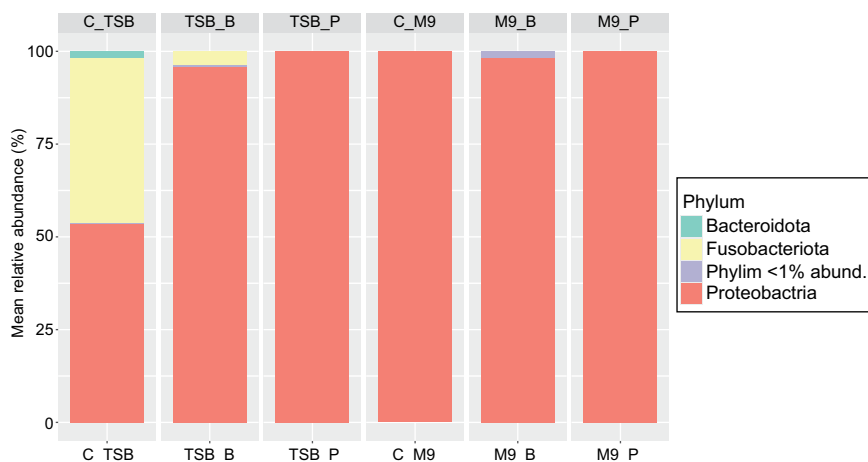


Fig. 4. Mean relative abundance values (%) of the resulting phyla in each experimental condition. Triptone soy broth (TSB) and minimal media (M9) as controls (C_TSB, C_M9), and each media supplemented with benzene (TSB_B, M9_B) or pentane (TSB_P, M9_P).

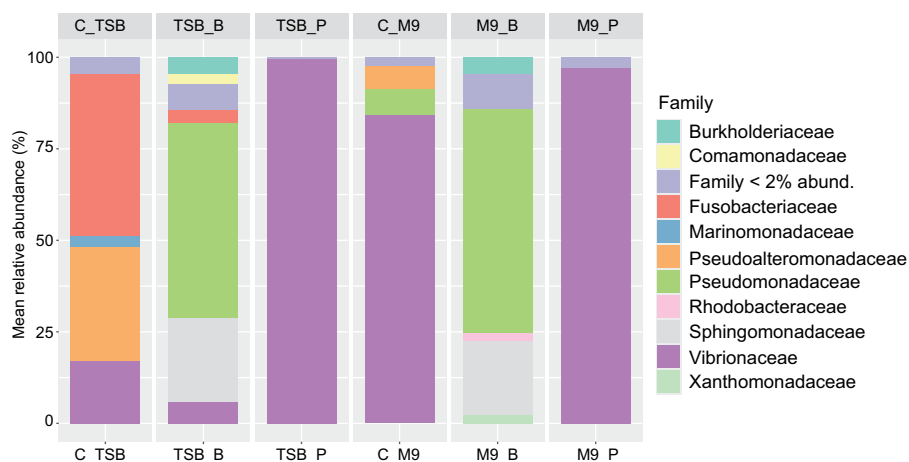


Fig. 5. Average relative abundance values (%) of the resulting families in each experimental condition. Tryptone soy broth (TSB) and minimal media (M9) as controls (C_TSB, C_M9), and each media supplemented with benzene (TSB_B, M9_B) or pentane (TSB_P, M9_P).

Pseudoalteromonas sp. (< 15%) to a lesser extent. The abundance of the genera corresponding to this family decreased below 15% in the TSB control, in favor of the genera *Fusobacterium* sp. (< 40%) and *Pseudoalteromonas* sp. (> 30%). The addition of benzene inhibits the growth of *Aliivibrio* sp. and decreases the relative abundance of *Vibrio* sp. and *Fusobacterium* sp. in TSB_B to below 10%. In minimal media, benzene prevents the development of all genera of the *Vibrionaceae* and *Fusobacteriaceae* families. Benzene conditions are mainly represented by the genus *Pseudomonas* sp. (> 40%), as well as others such as *Blastomonas* sp., *Burkholderia*

sp. and *Sphingopyxis* sp. to a lesser extent. This hydrocarbon also promotes the minor appearance of *Acidovorax* sp. in TSB media, as well as that of *Pseudoxanthomonas* sp. in M9 media. M9 media with pentane (M9_P) were mostly represented by *Vibrio* sp. (>90%), while the presence of this hydrocarbon in TSB media (TSB_P) also allowed an abundance greater than 5% of the genus *Aliivibrio* sp. (**Fig. 6**)

Table IV shows the average values of the number of ASVs sampled, as well as the Chao1, Shannon, and Simpson indices for the alpha-diversity analysis by experimental condition. One of the samples cor-

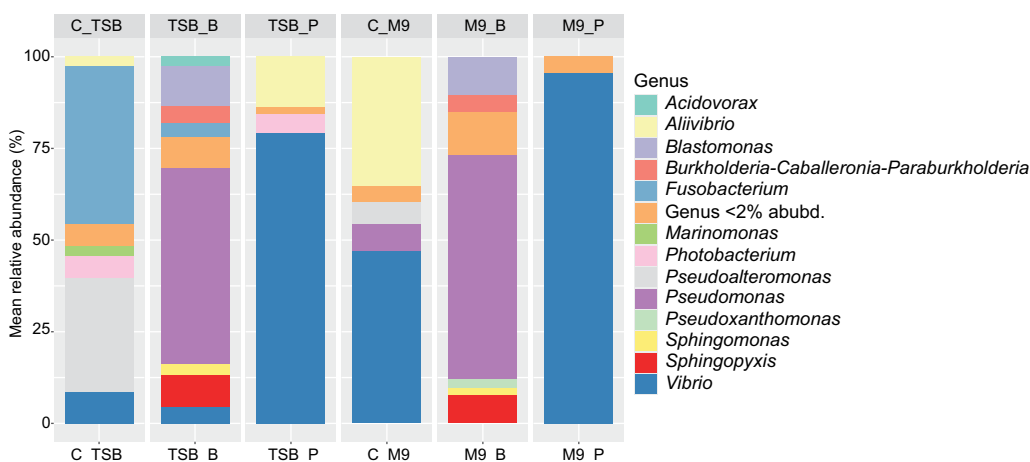


Fig. 6. Average relative abundance values (%) of the resulting genera in each experimental condition. Tryptone soy broth (TSB) and minimal media (M9) as controls (C_TSB, C_M9), and each media supplemented with benzene (TSB_B, M9_B) or pentane (TSB_P, M9_P).

TABLE IV. AVERAGE ALPHA-DIVERSITY VALUES BY EXPERIMENTAL CONDITIONS, BASED ON AMPLICON SEQUENCE VARIANTS (ASVs), CHAO1, SHANNON, AND SIMPSON INDICES. TRIPTONE SOY BROTH (TSB) AND MINIMAL MEDIA (M9) AS CONTROLS (C_TSB, C_M9), AND EACH MEDIA SUPPLEMENTED WITH BENZENE (TSB_B, M9_B) OR PENTANE (TSB_P, M9_P). MEAN VALUES \pm STANDARD DEVIATION ARE REPRESENTED.

Condition	ASVs	Chao1	Shannon	Simpson
C_TSB	113.50 \pm 12.02	113.50 \pm 12.02	2.28 \pm 0.10	0.76 \pm 0.01
TSB_P	96.00 \pm 35.36	96.00 \pm 35.36	2.72 \pm 0.46	0.87 \pm 0.05
C_M9	136.50 \pm 4.95	136.50 \pm 4.95	3.02 \pm 0.09	0.92 \pm 0.00
M9_B	200.00 \pm 8.49	200.00 \pm 8.49	2.45 \pm 0.13	0.85 \pm 0.02
M9_P	60.00 \pm 7.07	60.00 \pm 7.07	1.93 \pm 0.10	0.80 \pm 0.01

responding to the TSB_B condition failed the quality control, so the average value and deviation of this treatment could not be represented.

between C_M9 and M9_P, as well as greater distance between these communities and that of TSB_P (Fig. 7).

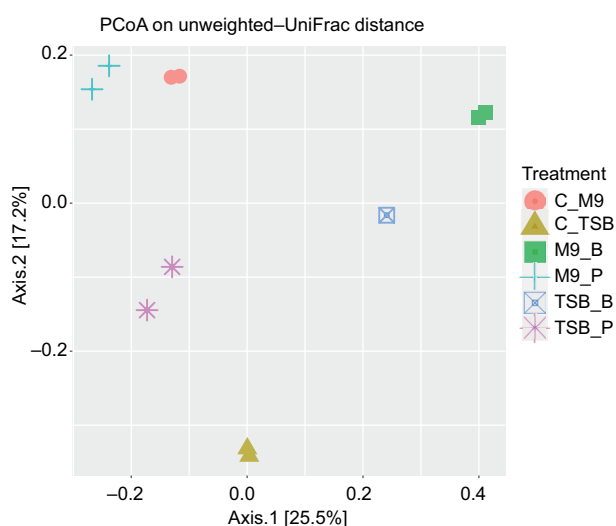


Fig. 7. Principal coordinate analysis (PCoA) between experimental conditions. Each symbol represents a sample of a given treatment. The legend relates each symbol to its corresponding treatment: triptone soy broth (TSB) and minimal media (M9) as controls (C_TSB, C_M9), and each media supplemented with benzene (TSB_B, M9_B) or pentane (TSB_P, M9_P).

Principal coordinate analysis (PCoA) allowed us to study how distinct the bacterial communities were between the different experimental groups, as well as possible differences within samples of the same experimental condition (Fig. 7).

The control groups (C-M9 vs C-TSB) differed greatly from each other. However, benzene produced more similar bacterial communities between TSB_B and M9_B, while pentane led to greater similarity

DISCUSSION

The presence of hydrocarbons in the marine environment is a significant environmental problem that alters microbial communities and other resident organisms. The environmental impact of these organic compounds in aquatic environments has been extensively documented in previous research (White et al. 2017, Arockiaraj and Kankara 2018, Pereira et al. 2018). This study reveals how hydrocarbon pollution affects the growth and composition of marine microorganisms in an experimental setting, highlighting the importance of nutrients such as nitrogen for the proliferation of bacteria with bioremediation potential. Limitation of essential nutrients in contaminated environments prolonged incubation times under all conditions, this effect being particularly pronounced in the presence of benzene and pentane, where slower microbial growth was observed compared to controls. This is in line with previous studies that also highlight nitrogen and phosphorus as key elements in the biostimulation of contaminated waters (Varjani and Upasani 2016a, Luo et al. 2024).

In the case of water samples, members of the genera *Vibrio* sp. and *Pseudoalteromonas* sp. were isolated, with *V. alginolyticus* standing out, a species previously identified as a potential diesel degrader (Imron et al. 2019, Muralidharan et al. 2023). Other species of biotechnological interest within this genus are *V. natrigiens* (Thoma and Blombach 2021, Walton and Buchan 2024) and *V. cyclotrophicus* (Hedlund and Staley 2001). Regarding the genus *Pseudoalteromonas*, some strains stand out, such as *Pseudoalteromonas* sp. JSTW, capable of degrading aliphatic and aromatic

compounds (Zan et al. 2021), and *P. agarivorans* SDRB-Py1, which is able to produce effective biosurfactants for the emulsification of hydrocarbons such as n-hexane and benzene, increasing their bioavailability (Lee et al. 2018, Ganesan et al. 2022).

Regarding the colonies isolated from the media supplemented with benzene, all the strains belonged to the genus *Bacillus* sp., which also encompasses some hydrocarbonoclastic species such as *B. algicola* 003-Phe1, capable of completely degrading aliphatic compounds of nine to 14 carbons, and others of higher molecular weight (Lee et al. 2018). *P. agarivorans* and *B. subtilis* have the ability to produce biosurfactants effective in the emulsification of benzene and n-hexane (Lee et al. 2018). *B. subtilis* is capable of producing lipopeptides and rhamnolipids as biosurfactants (Parthipan et al. 2017, Lee et al. 2018) and also presents genes encoding enzymes involved in the alkane degradation pathway, such as alkane hydrolase enzymes and alcohol dehydrogenases (Parthipan et al. 2017). The strains *B. amyloliquefaciens* VBS03 and *B. pumilus* VBS37 use hexadecane molecules as the sole carbon source in artificial marine culture media, degrading 99.6 and 86.9% of this compound over 28 days, respectively (Ferrari et al. 2019). The formation of a bacterial consortium consisting of strains of *B. pumillus*, *P. stutzeri*, and *A. calcoaceticus* was able to catabolize PAHs such as pyrene and anthracene (Marzuki et al. 2021).

Pentane present in the culture media selected or promoted the growth of *Vibrio* sp. again, suggesting that these organisms are particularly resilient to the presence of low molecular weight aliphatic hydrocarbons (Hedlund and Staley 2001, Fernandes et al. 2020).

Metagenomic analysis of the bacterial groups present in each culture medium again confirms the dominance of the Proteobacteria phylum in all media with benzene and pentane. The absence of the Bacteroidota and Fusobacteriota phyla in minimal media indicates that their nutritional requirements are more demanding than those of Proteobacteria, which mainly require a nitrogen source to grow. Furthermore, benzene inhibits the growth of Bacteroidota and limits Fusobacteriota, decreasing its abundance compared to the control. Pentane also shows a negative influence on both phyla, benefiting Proteobacteria. However, a similar study carried out on the coasts of Lebanon positioned Bacteroidota as the second most abundant phylum behind Proteobacteria in the presence of oil spills (Hamdan et al. 2023). In addition, the class Bacteroidia has been identified in marine sediments contaminated by hydrocarbons (Dell'Anno et al. 2021).

It should be noted that the presence of the Fusobacteriota phylum in the control is unusual compared to similar studies, where it is not even a representative group (Hamdan et al. 2023). Gutiérrez et al. (2016) identified the genus *Fusobacterium* sp., which belongs to this phylum, on the ocean surface of the Gulf of Mexico after the Deepwater Horizon environmental accident. The identification of this genus in surface layers of the marine environment, where oxygen availability is high, is atypical due to its obligatory anaerobic metabolism, which allows it to be commonly found in marine sediments (Gutiérrez et al. 2016). These results lead to future research to explain the development of this genus under aerobic conditions.

Regarding the identified genera, the presence of *Vibrio* sp. increases significantly in media supplemented with pentane. The abundance of this genus has already been described in marine environments contaminated with tar balls (Fernandes et al. 2020). Meanwhile, media with benzene increased the abundance of the genus *Pseudomonas* sp. This fact does not correspond with the results obtained in similar studies (Hamdan et al. 2023). However, its degradative potential was studied in different species previously mentioned (Mittal and Singh 2009, Rocha et al. 2011, Sajna et al. 2015, Varjani and Upasani 2016a, b).

This study also shows how benzene promotes the growth of other genera, such as *Burkholderia* sp., *Sphingomonas* sp., *Blastomonas* sp., *Sphingopyxis* sp., and *Acidovorax* sp., which were previously classified as PAH degraders according to this study (Seo et al. 2009). The genomic study of different strains of the genus *Sphingopyxis* sp. reveals the appearance of genes related to the degradation of aromatic hydrocarbons through benzoate metabolism (Sharma et al. 2021, Yang et al. 2020). The genus *Blastomonas* sp. is also representative only in treatments with benzene, however, no previous studies have been found that study the influence of benzene on the growth of this genus. On the other hand, media enriched with pentane reflected the abundance of the genus *Vibrio* sp.

However, environmental factors such as temperature, oxygen availability, and the presence of additional contaminants can modulate microbial metabolic pathways and, consequently, affect their hydrocarbon degradation efficiency. Although this study did not expose the bacterial communities to temperatures above the Mediterranean average, it is important to consider that climate change may alter microbial dynamics and functionality. These shifts could either enhance or impair the effectiveness of bioremediation processes. Therefore, future research should evaluate the metabolic stability and

performance of these candidate strains under variable environmental conditions that simulate projected climate scenarios.

Although this study focused on microbial characterization in the presence of benzene and pentane, it is important to consider the ecological implications of future field applications. Benzene degradation can generate intermediate compounds such as catechols or organic acids, which may retain some toxicity if not fully metabolized. Pentane degradation mainly produces short-chain alcohols, aldehydes, and carboxylic acids, which are less toxic but could still disrupt microbial balance or water quality at high concentrations. Additionally, the bioestimulation approach used, adding nutrients like casamino acids and glucose, could cause ecological imbalances if applied in natural environments without proper control. Therefore, nutrient selection and dosing must be carefully evaluated in future studies to minimize potential environmental impacts.

Finally, this study also highlights how the combined use of molecular and culture techniques can provide a more accurate view of microbial communities in contaminated environments. Alpha-diversity data and differences in beta-diversity show a clear response of the marine microbiota to different experimental conditions and types of hydrocarbons. These results highlight the importance of a comprehensive approach in the study of marine microbiota and its role in bioremediation, showing that each hydrocarbon induces specific variations in the composition and diversity of bacterial communities (Seymour 2014, White et al. 2017).

CONCLUSIONS

The addition of hydrocarbons such as benzene and pentane proved to be a limiting factor, prolonging incubation times and reducing bacterial growth in all experimental media, even those enriched with carbon and nitrogen sources. However, these contaminants favored the development of bacterial genera such as *Pseudomonas* sp., *Bacillus* sp., and *Vibrio* sp., which have been previously identified as hydrocarbon degraders. Specifically, benzene promoted the abundance of *Pseudomonas* sp. and facilitated the isolation of *Bacillus* sp., while pentane significantly increased the presence of *Vibrio* sp. These discoveries highlight the importance of these bacteria in the bioremediation of hydrocarbons in contaminated marine ecosystems, pointing to future research on their specific degradative capabilities in the presence of different types of hydrocarbons.

ACKNOWLEDGMENTS

The authors gratefully thank the University of Málaga for its institutional support throughout the development of this work.

REFERENCES

- Acosta-González A. and Marqués S. (2016). Bacterial diversity in oil-polluted marine coastal sediments. *Current Opinion in Biotechnology* 38, 24-32. <https://doi.org/10.1016/j.copbio.2015.12.010>
- Arockiaraj S. and Kankara R.S. (2019) Assessment of potential oil spill risk along Vishakhapatnam coast, India: Integrated approach for coastal management. *Coastal Management: Global Challenges and Innovations* (2019), 449-463. <https://doi.org/10.1016/B978-0-12-810473-6.00021-2>
- Bar-On Y.M. and Milo R. (2019). The biomass composition of the oceans: A blueprint of our blue planet. *Cell* 179 (7), 1451-1454. <https://doi.org/10.1016/j.cell.2019.11.018>
- Cariello Delunardo F.A., Paulino M.G., Campos Medeiros L.C, Narciso Fernandes M., Scherer R. and Chippari-Gomes A.R. (2020) Morphological and histopathological changes in seahorse (*Hippocampus reidi*) gills after exposure to the water-accommodated fraction of diesel oil. *Marine Pollution Bulletin* 150, 110769. <https://doi.org/10.1016/j.marpolbul.2019.110769>
- Chandra S., Sharma R., Singh K. and Sharma A. (2013) Application of bioremediation technology in the environment contaminated with petroleum hydrocarbon. *Annals of Microbiology* 63, 417-431. <https://doi.org/10.1007/s13213-012-0543-3>
- CSM (2023). Balance turístico verano 2023. Costa del Sol Málaga [online]. https://www.costadelsolmalaga.org/5773/com1_tc-0/com1_gs-293/com1_mn-0/descarga-de-documentos 09/06/24
- Dash H.R., Mangwani N., Chakraborty J., Kumari S. and Das S. (2013) Marine bacteria: Potential candidates for enhanced bioremediation. *Applied Microbiology and Biotechnology* 97 (2), 561-571. <https://doi.org/10.1007/s00253-012-4584-0>
- Dell'Anno F., Rastelli E., Tangherlini M., Corinaldesi C., Sansone C., Brunet C., Balzano S., Ianora A., Musco L., Montereali M. and Dell'Anno A. (2021). Highly contaminated marine sediments can host rare bacterial taxa potentially useful for bioremediation. *Frontiers in Microbiology* 12, 1-13. <https://doi.org/10.3389/fmicb.2021.584850>
- Delong E.F. (2009). The microbial ocean from genomes to biomes. *Nature* 459, 200-206. <https://doi.org/10.1038/nature08059>

- Deutschmann I.M., Krabberød A.K., Latorre F., Delage E., Marrasé C., Balagué V., Gasol J.M., Massana R., Eveillard D., Chaffron S. and Logares R. (2023). Disentangling temporal associations in marine microbial networks. *Microbiome* 11 (83), 1-16. <https://doi.org/10.1186/s40168-023-01523-z>
- DPM (2024). Información y ocio. Diputación Provincial de Málaga [online]. https://www.malaga.es/es/laprovincia/informacionyocio/tp-45/lis_pg-3/09/06/24
- Fernandes C., Khandeparker R.D.S. and Shenoy B.D. (2020). High abundance of *Vibrio* in tarball-contaminated seawater from Vagator beach, Goa, India. *Marine Pollution Bulletin* 150, 110773. <https://doi.org/10.1016/j.marpolbul.2019.110773>
- Ferrari V.B., Cesar A., Cayô R., Choueri R.B., Okamoto D.N., Freitas J.G., Favero M., Gales A.C., Niero C.V., Saia F.T. and de Vasconcellos S.P. (2019). Hexadecane biodegradation of high efficiency by bacterial isolates from Santos Basin sediments. *Marine Pollution Bulletin* 142, 309-314. <https://doi.org/10.1016/j.marpolbul.2019.03.050>
- Ganesan M., Mani R., Sai S., Kasivelu G., Awasthi M.K., Rajagopal R., Wan Azelee N.I., Selvi P.K., Chang S.W. and Ravindran B. (2022). Bioremediation by oil degrading marine bacteria: An overview of supplements and pathways in key processes. *Chemosphere* 303 (Part 1), 134956. <https://doi.org/10.1016/j.chemosphere.2022.134956>
- Gao Y., Cai M., Shi K., Sun R., Liu S., Li Q., Wang X., Hua W., Qiao Y., Xue J. and Xiao X. (2023). Bioaugmentation enhance the bioremediation of marine crude oil pollution: Microbial communities and metabolic pathways. *Water Science and Technology* 87 (1), 228-238. <https://doi.org/10.2166/wst.2022.406>
- Goveas L.C., Nayak S. and Selvaraj R. (2022). Concise review on bacterial degradation of petroleum hydrocarbons: Emphasis on Indian marine environment. *Bioresource Technology Reports* 19, 101136. <https://doi.org/10.1016/j.biteb.2022.101136>
- Gutiérrez T., Berry D., Teske A. and Aitken M.D. (2016). Enrichment of *Fusobacteria* in sea surface oil slicks from the Deepwater Horizon oil spill. *Microorganisms* 4 (3), 24. <https://doi.org/10.3390/microorganisms4030024>
- Hamdan H.Z., Salam D.A. and Saikaly P.E. (2019). Characterization of the microbial community diversity and composition of the coast of Lebanon: Potential for petroleum oil biodegradation. *Marine Pollution Bulletin* 149, 110508. <https://doi.org/10.1016/j.marpolbul.2019.110508>
- Hamdan H.Z., Ahmad F.A., Zayyat R.M. and Salam D.A. (2023). Spatio-temporal variation of the microbial community of the coast of Lebanon in response to petroleum hydrocarbon pollution. *Marine Pollution Bulletin* 192, 115037. <https://doi.org/10.1016/j.marpolbul.2023.115037>
- Hedlund B.P. and Staley J.T. (2001). *Vibrio cyclotrophicus* sp. nov., a polycyclic aromatic hydrocarbon (PAH)-degrading marine bacterium. *International Journal of Systematic and Evolutionary Microbiology* 51 (1), 61-66. <https://doi.org/10.1099/00207713-51-1-61>
- Honda M. and Suzuki N. (2020). Toxicities of polycyclic aromatic hydrocarbons for aquatic animals. *International Journal of Environmental Research and Public Health* 17 (4), 1363. <https://doi.org/10.3390/ijerph17041363>
- Imron M.F., Kurniawan S.B. and Titah H.S. (2019). Potential of bacteria isolated from diesel-contaminated seawater in diesel biodegradation. *Environmental Technology and Innovation* 14, 100368. <https://doi.org/10.1016/j.eti.2019.100368>
- Lee D.W., Lee H., Kwon B.O., Khim J.S., Yim U.H., Kim B.S. and Kim J.J. (2018). Biosurfactant-assisted bioremediation of crude oil by indigenous bacteria isolated from Taean beach sediment. *Environmental Pollution* 241, 254-264. <https://doi.org/10.1016/j.envpol.2018.05.070>
- Little D.I., Sheppard S.R.J. and Hulme D. (2021). A perspective on oil spills: What we should have learned about global warming. *Ocean and Coastal Management* 202, 105509. <https://doi.org/10.1016/j.ocecoaman.2020.105509>
- Luo Q., Wang Y., Zhu B., Chen Q. and Sun W. (2024). Bioremediation of diesel oil polluted seawater by a hydrocarbon degrading bacterial consortium with oleophilic nutrients. *Regional Studies in Marine Science* 71, 103412. <https://doi.org/10.1016/j.rsma.2024.103412>
- Marzuki I., Asaf R., Paena M., Athirah A., Nisaa K., Ahmad R. and Kamaruddin M. (2021). Anthracene and pyrene biodegradation performance of marine sponge symbiont bacteria consortium. *Molecules* 26 (22), 6851. <https://doi.org/10.3390/molecules26226851>
- Mittal A. and Singh P. (2009). Isolation of hydrocarbon degrading bacteria from soils contaminated with crude oil spills. *Indian Journal of Experimental Biology* 47 (9), 760-765.
- Muralidharan M., Gayathri K.V., Kumar P.S., Preethi D.S., Kavitha R., Rajagopal R. and Rangasamy G. (2023). Mixed polyaromatic hydrocarbon degradation by halotolerant bacterial strains from marine environment and its metabolic pathway. *Environmental Research* 216, 114464. <https://doi.org/10.1016/j.envres.2022.114464>
- Muyzer G. (1998). Structure, function and dynamics of microbial communities: The molecular biological approach. In: *Advances in molecular ecology* (Carvalho G.R., Ed.). IOS Press, Amsterdam, Netherlands, 87-117.

- Parthipan P., Preetham E., Machuca L.L., Rahman P.K.S.M., Murugan K. and Rajasekar A. (2017). Bio-surfactant and degradative enzymes mediated crude oil degradation by bacterium *Bacillus subtilis* A1. *Frontiers in Microbiology* 8, 193. <https://doi.org/10.3389/fmicb.2017.00193>
- PM (2024). Muelle Uno. Puerto de Málaga [online]. <https://www.muelleuno.com/centro-comercial-malaga/09/06/24>
- Pereira T.M., Merçon J., Passos L.S., Coppo G.C., Lopes T.O.M., Cabral D.S., Scherer R. and Chippari-Gomes A.R. (2018). Effects of the water-soluble fraction of diesel oil (WSD) on the fertilization and development of a sea urchin (*Echinometra lucunter*). *Ecotoxicology and Environmental Safety* 162, 59-62. <https://doi.org/10.1016/j.ecoenv.2018.06.040>
- Qian Y., Xu M., Deng T., Hu W., He Z., Yang X., Wang B., Song D., Chen L., Huang Y. and Sun G. (2020). Synergistic interactions of *Desulfovibrio* and *Petrimonas* for sulfate-reduction coupling polycyclic aromatic hydrocarbon degradation. *Journal of Hazardous Materials* 407, 124385. <https://doi.org/10.1016/j.jhazmat.2020.124385>
- Rocha C.A., Pedregosa A.M. and Laborda F. (2011). Biosurfactant-mediated biodegradation of straight and methyl-branched alkanes by *Pseudomonas aeruginosa* ATCC 55925. *AMB Express* 1, 9. <https://doi.org/10.1186/2191-0855-1-9>
- Sajna K.V., Sukumaran R.K., Gottumukkala L.D. and Pandey A. (2015). Crude oil biodegradation aided by biosurfactants from *Pseudozyma* sp. NII 08165 or its culture broth. *Bioresource Technology* 191, 133-139. <https://doi.org/10.1016/j.biortech.2015.04.126>
- Sanz-Sáez I., Salazar G., Sánchez P., Lara E., Royo-Llonch M., Sà E., Lucena T., Pujalte M., Vaqué D., Duarte C., Gasol J., Pedrós-Alió C., Sánchez O. and Acinas S. (2020). Diversity and distribution of marine heterotrophic bacteria from a large culture collection. *BMC Microbiology* 20, 207. <https://doi.org/10.1186/s12866-020-01884-7>
- Seo J.S., Keum Y.S. and Li Q.X. (2009). Bacterial degradation of aromatic compounds. *International Journal of Environmental Research and Public Health* 6 (1), 278-309. <https://doi.org/10.3390%2Fijerph6010278>
- Seymour J.R. (2014). A sea of microbes: The diversity and activity of marine microorganisms. *Microbiology Australia* 35 (4), 183. <https://doi.org/10.1071/ma14060>
- Sharma M., Khurana H., Singh D.N. and Negi R.K. (2021). The genus *Sphingopyxis*: Systematics, ecology, and bioremediation potential—A review. *Journal of Environmental Management* 280, 111744. <https://doi.org/10.1016/j.jenvman.2020.111744>
- Sinha R.K., Krishnan K.P. and Kurian P.J. (2021). Complete genome sequence and comparative genome analysis of *Alcanivorax* sp. IO_7, a marine alkane-degrading bacterium isolated from hydrothermally-influenced deep seawater of southwest Indian Ridge. *Genomics* 113, 884-891. <https://doi.org/10.1016/j.ygeno.2020.10.020>
- ST (2024). Sea surface temperatures on the coast [online]. <https://seatemperature.info/es/agosto/malaga-temperatura-del-agua-del-mar.html> 10/06/24
- Thoma F. and Blombach B. (2021) Metabolic engineering of *Vibrio natriegens*. *Essays in Biochemistry* 65 (2), 381-392. <https://doi.org/10.1042/EBC20200135>
- Varjani S.J. and Upasani V.N. (2016a). Carbon spectrum utilization by an indigenous strain of *Pseudomonas aeruginosa* NCIM 5514: Production, characterization and surface active properties of biosurfactant. *Bioresource Technology* 221, 510-516. <https://doi.org/10.1016/j.biortech.2016.09.080>
- Varjani S.J. and Upasani V.N. (2016b). A core flood study for enhanced oil recovery through ex-situ bioaugmentation with thermo- and halo-tolerant rhamnolipid produced by *Pseudomonas aeruginosa* NCIM 5514. *Bioresource Technology* 220, 175-182. <https://doi.org/10.1016/j.biortech.2016.08.060>
- Varjani S.J. (2017). Microbial degradation of petroleum hydrocarbons. *Bioresource Technology* 223, 277-286. <https://doi.org/10.1016/j.biortech.2016.10.037>
- Varjani S.J. and Upasani V.N. (2017). A new look on factors affecting microbial degradation of petroleum hydrocarbon pollutants. *International Biodeterioration and Biodegradation* 120, 71-83. <https://doi.org/10.1016/j.ibiod.2017.02.006>
- Walton J.L. and Buchan A. (2024). Evidence for novel polycyclic aromatic hydrocarbon degradation pathways in culturable marine isolates. *Microbiology Spectrum* 12 (1), e0340923. <https://doi.org/10.1128/spectrum.03409-23>
- White N.D., Godard-Codding C., Webb S.J., Bossart G.D. and Fair P.A. (2017). Immunotoxic effects of in vitro exposure of dolphin lymphocytes to Louisiana sweet crude oil and Corexit™. *Journal of Applied Toxicology* 37 (6), 676-682. <https://doi.org/10.1002/jat.3414>
- Yang F., Feng H., Massey I.Y., Huang F., Guo J. and Zhang X. (2020). Genome-wide analysis reveals genetic potential for aromatic compounds biodegradation of *Sphingopyxis*. *BioMed Research International* (1), 5849123. <https://doi.org/10.1155/2020/5849123>
- Zan S., Lv J., Li Z., Cai Y., Wang Z. and Wang J. (2021). Genomic insights into *Pseudoalteromonas* sp. JSTW coping with petroleum-heavy metals combined pollution. *Journal of Basic Microbiology* 61 (10), 947-957. <https://doi.org/10.1002/jobm.202100156>