

CONCENTRATION OF TOXIC ELEMENTS IN TOPSOILS OF THE METROPOLITAN AREA OF MEXICO CITY: A SPATIAL ANALYSIS USING ORDINARY KRIGING AND INDICATOR KRIGING

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Key words: Pb, V, Ni, Cu, contamination factor, pollution load index

ABSTRACT

In order to generate relevant information for decision-makers to protect the inhabitants of the Metropolitan Area of Mexico City, 89 samples of topsoil were collected in urban zones to assess the concentration of toxic elements. For this purpose, the concentration of Cr, Cu, Ni, Pb, V and Zn were quantified with X-ray fluorescence. To evaluate the pollution, we considered the maximum permissible limits, the contamination factor and the pollution load index. The spatial distribution was analyzed by geostatistical methods. Ordinary and indicator Kriging were applied to estimate the values at unmeasured sites and the proportion of the values exceeding the critical concentration for each element within a region. The study revealed that the Metropolitan Area of Mexico City has excessive environmental problems related to pollution. The concentrations of Cr and V are apparently not anthropic, while the high amounts of Cu, Zn and Pb are largely due to human activities. The pollution of Ni exhibits two single spots, while Cu and Zn contents are distributed from the city center towards the northern suburbs. The presence of Pb is spread evenly through the entire urbanized area. The probability maps clearly identify the most contaminated areas, which requires immediate action by local decision-makers.

Palabras clave: Pb, V, Ni, Cu, factor de contaminación, índice de carga contaminante

RESUMEN

Con la finalidad de generar información para la toma de decisiones y proteger a los habitantes del área metropolitana de la ciudad de México, se tomaron 89 muestras de

suelo superficial urbano, con las que se evaluó la concentración de elementos tóxicos. Se midió la concentración de Cr, Cu, Ni, Pb, V y Zn utilizando un equipo de fluorescencia de rayos X. Para valorar la contaminación del suelo se utilizaron los límites máximos permisibles, el factor de contaminación y el índice de carga contaminante. Se realizó un análisis de la distribución espacial de los elementos medidos por los métodos geoestadísticos como Kriging ordinario y el indicador Kriging. El análisis reveló que la zona metropolitana de la Ciudad de México tiene problemas ambientales excesivos. Las concentraciones de Cr y V al parecer no tienen causas antrópicas, mientras que las altas cargas de Cu, Zn y Pb son en gran medida de origen humano. La contaminación por Ni presenta dos puntos esenciales, el Cu y Zn se distribuyen desde el centro hacia los suburbios del norte. El Pb está distribuido uniformemente sobre el área urbanizada. Los mapas probabilísticos permiten identificar claramente las zonas más contaminadas en las cuales los tomadores de decisiones locales deben tomar acciones inmediatas de remediación.

INTRODUCTION

The metropolitan area of Mexico City (MAMC) has more than 20 million inhabitants and a still-growing population. The MAMC has more than 40 000 small and medium-sized industries, and vehicles use daily more than 40 million L of petroleum fuels producing thousands of t of pollutants (Molina *et al.* 2010, UN 2012). However, diagnoses of urban topsoil contamination with toxic elements (TE) are still scarce (Morton-Bermea *et al.* 2009, Rodríguez-Salazar *et al.* 2011, Aguilar *et al.* 2012).

This MAMC is the third-largest urban region in the world (UN 2012). It lies on an intermountain basin, with an altitude of approximately 2200 m above the sea level and is surrounded on three sides by recent volcanoes. This physical location, together with local phenomenon such as intense emissions from private and public transport and high concentrations of industries in the affluent suburbs, frequently causes smog along the city. The winds blow over the mountains, but there is little circulation to transport the heavier particles of smog and dust out of the basin. Therefore, emissions accumulate throughout the morning and afternoon. During the evening, wind rises and disperses a lot of the smog, but settles it again over the basin during the cooler nights. This pattern is repeated on subsequent days. Furthermore the air quality is generally worse in winter, when there is practically no rain and thermal inversions are more frequent (Molina *et al.* 2010).

Over the last decades the term “heavy metals” is widely used in environmental studies for their toxicity, although there are still considerable uncertainties about their definition. In specific literature, the density criteria varies from above 3.5 g/cm to above 7 g/cm (Duffus 2002). In addition, we have also transition

and post-transition metals as toxic elements or trace elements with considerable toxicity. Consequently we use for this study the term “toxic elements” (TE) instead of “heavy metals” as commonly used by the United States Environmental Protection Agency (US-EPA 2012).

There are different anthropic sources of TE in urban soils (Sutherland 2000, Aguilar *et al.* 2011, 2012, Alloway 2012, Aguilar *et al.* 2013a,b), their loads in and around human settlements are a global problem. They enter the soils by different pathways: 1) aerial deposition (industries, vehicles and volcanoes), 2) paints, 3) pesticide and fertilizer application, 4) waste utilization, 5) disposal of dredged sediments and 6) river and irrigation waters (Kabata-Pendias 1993, McClintock 2012, Aguilar *et al.* 2013b). They pass through the air and food into the human organism, where they are usually not degraded, and hence they accumulate and may cause cancer, neuropathy and other serious diseases (Fubini and Otero 1999, Muhle and Mangelsdorf 2003).

Geostatistic tools are now incorporated in most geographic information systems. Among them, ordinary kriging (OK) interpolates the values of a random field at unobserved locations by using values measured nearby sites. This approach is based on the “first law of geography” that “everything is related to everything else, but adjacent things are more related to each other” (Tobler 1970). However, soil pollution shows a highly skewed nature, and Juang *et al.* (2004) note that it is inappropriate to use certain geostatistical methods, such as OK to characterize their spatial distributions by a Gaussian assumption.

In contrast, indicator kriging (IK) is a non-parametric geostatistical method, which includes no assumption of normality and uses a binary transformation (0-1 indicators) of data to make the predictor

less sensitive to outliers. IK builds the cumulative distribution function at each point on the basis of the pattern and correlation structure of indicator-transformed data points in the neighborhood. IK can be used to estimate the proportion of values that exceed critical concentrations of TE by incorporating the uncertainty of the value of variables at unobserved locations. After the cumulative distribution function is built, it must be post-processed to produce a probability map (Goovaerts 1997, 1999, Antunes and Albuquerque 2013, Aguilar *et al.* 2013b).

The aim of this study was to analyze the spatial distribution of toxic elements (Pb, Zn, Ni, Cu, Cr and V) in urban topsoil's from MAMC, using the IK and OK geostatistical techniques.

Diagnostic studies of soil pollution by TE should refer to threshold values established in terms of effects on human health. For this purpose, we used: (1) the maximum permissible limits for the Official

Mexican Standards (Norma Oficial Mexicana "NOM"), (2) the contamination factor (CF), and (3) the pollution load index (PLI).

MATERIALS AND METHODS

Study area, sampling and chemical analysis

The study area covers almost the entire urbanized area of the MAMC (Fig. 1). Soils were collected using a systematic method. For this purpose, the study area of 1600 km² was divided into a grid of 10 rows and 10 columns, so that each cell had an edge length of 4 km and an area of 16 km². The high degree to which the soil was covered made necessary to change some sampling sites within the grid, depending on local conditions. For some sparsely populated peripheral areas by the surroundings of Texcoco and Cuernavaca cities, sampling was omitted for financial reasons.

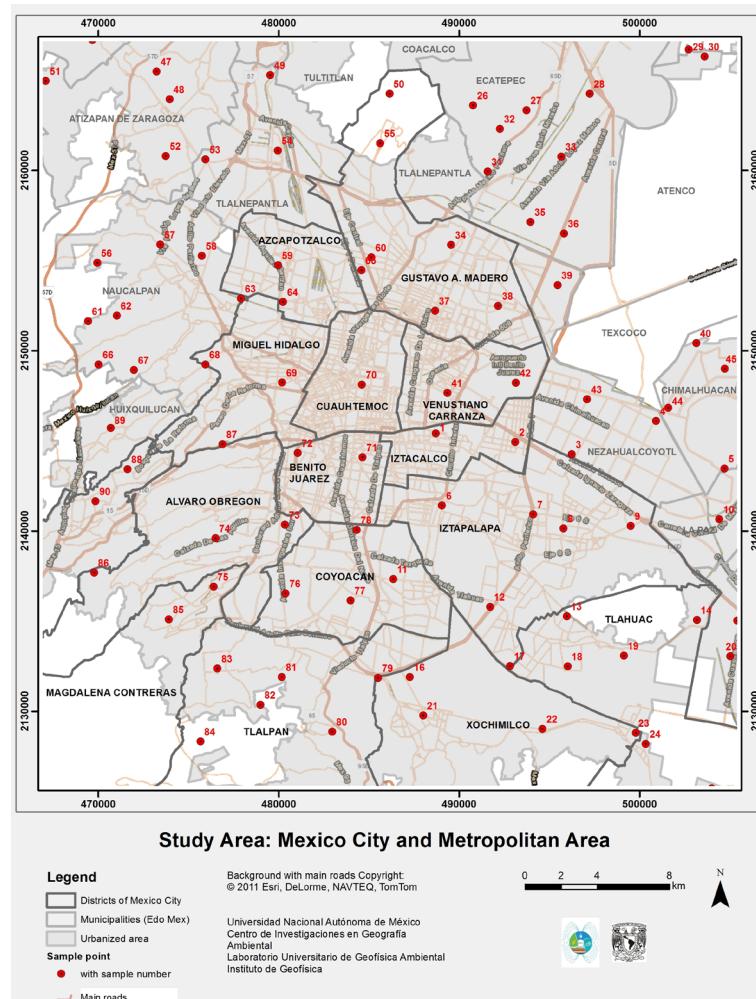


Fig. 1 Location of sampling sites in Mexico City and metropolitan area

For analysis, samples from 89 sites were collected at a depth of 2 inches using a polyvinyl chloride (PVC) cylinder of 2.5 inches in diameter. The amount of sample sites depends much less on a fixed minimum number than other facts, including the kriging prediction variance and the accuracy which the variogram can estimate. As a rule of thumb is often mentioned a number between 30 and 100 samples (Webster and Oliver 1990). In practice there are only few studies that use more than 100 sampling data.

Each sample was dried for two weeks under the shade and then pulverized in a 2-mm mesh sieve. The total concentrations of Pb, Zn, Ni, Cu, Cr and V were analyzed by energy dispersive X-ray fluorescence (ED-XRF) in the laboratory at the Advanced Research and Studies Center, called Centro de Investigación y de Estudios Avanzados (CINVESTAV), in Mérida, Yucatán. This technique is precise quantitative, fast and non-destructive and can measure a wide range of elements concentrations (C (6)-Fm (100); Khodeir *et al.* 2012, Ramírez *et al.* 2012), with a detection limit of 1 mg/kg even in complex matrices as soils. Only in the case of Ni, the minimum value was below the detection limit.

The powdered soils were pressed into pellets of 1 cm diameter without any chemical treatment or binding agents and inserted into a plastic sample holder covered with a polyester film. ED-XRF analyses were performed with a Jordan Valley EX-6600 spectrometer, equipped with a Si (Li) detector with a 20 mm² active area and 140 eV resolutions at 5.9 keV, performed at a maximum of 54 kV and 4800 mA. TE were acquired in an air atmosphere using a changeable secondary target. The typical measurement time was 300 s. Each measurement was replicated five times in order to obtain the average concentration. Quantitative data was calculated with a calibration curve of intensity versus concentration of several geological standard reference materials (Lozano and Bernal 2005, Khodeir *et al.* 2012).

Data analyses

The data obtained was analyzed in four stages: a) descriptive statistics were generated, including arithmetic means, standard deviations, minimum and maximum values for TE, b) data was compared with the maximum permissible values, c) pollution indices were applied and d) the geostatistic analysis was conducted.

Permissible limits

The Official Mexican Standards “NOM-147-SEMARNAT” of Soil Contamination was

published by the Ministry of Environment and Natural Resources (Secretaría de Medio Ambiente y Recursos Naturales; SEMARNAT) and include only the permissible limits of Pb (400 mg/kg), V (78 mg/kg), Ni (1600 mg/kg) and Cr (280 mg/kg), but exclude Cu and Zn (SEMARNAT 2007). The Official Mexican Standards are largely based on those of the US Environment Protection Agency (US-EPA 2012).

Pollution indices

Pollution indices are used to quantify TE in soils and sediments and exist as a single index for one metal, as well as an integrated index as a combination of the six elements. In the past, a large number of indices were derived, which include CF, the ecological risk factor, the enrichment factor, the index of geo-accumulation, and the PLI. Unfortunately, there is not always congruence between the different indices and publications (Sutherland 2000, Buccolieri *et al.* 2006). Therefore, the indices used in this study were defined as follows:

Contamination factor (CF)

The level of soil contamination by TE is a ratio calculated as follows:

$$CF = \frac{C_m \text{ Sample}}{C_m \text{ Background}} \quad (1)$$

Where, *CF* is the TE contamination factor, *C_m Sample* is the metal concentration of the sample and *C_m Background* is the natural metal content.

The sample was normalized with a background that reflects the natural metal content of the parent material. The background can be determined by various methods. However, it is often problematic. Certainly, the classical method takes a sample outside of anthropic influence, or at least where it is as low as possible. Estimated minimum distances are given from the study area or the sample is collected in a natural protected area; in this context, literature gives some reference on the concentrations. Hakanson (1980) provides concentrations before the industrialization of Cu 50 mg/kg, Pb 70 mg/kg, Cr 90 mg/kg and Zn 175 mg/kg. It remains questionable any transmission to different conditions (soils, substrate, source rocks). Fortunately, Morton-Bermea *et al.* (2009) have measured the backgrounds for the MAMC (**Table I**). These natural concentrations of TE are used for the CF.

CF indicates pollution according to the following intervals: a value below 1 indicates insignificant

TABLE I. TOXIC ELEMENT CONCENTRATIONS REPORTED FOR MEXICO CITY

Elements	Min	Max	Mean (mg /kg)	Background	Reference
Cr	41	138	84	105	This study
	50	265	116		Morton-Bermea <i>et al.</i> 2009.
	65	559	135		Rodríguez-Salazar <i>et al.</i> 2011.
		280			Recommended by SEMARNAT (2007).
Ni	dl ^a	186	45	56	This study.
	20	146	39		Morton-Bermea <i>et al.</i> 2009.
	29	151	49		Rodríguez-Salazar <i>et al.</i> 2011.
		1600			Recommended by SEMARNAT (2007).
Cu	7	550	92	32	This study.
	15	398	54		Morton-Bermea <i>et al.</i> 2009
	26	461	93		Rodríguez-Salazar <i>et al.</i> 2011.
		3100			Recommended by EPA (region 9).
Zn	82	933	287	76	This study.
	36	1641	219		Morton-Bermea <i>et al.</i> 2009.
	95	1890	447		Rodríguez-Salazar <i>et al.</i> 2011.
		32000			Recommended by EPA (region 9).
Pb	20	654	163	19	This study.
	5	452	82		Morton-Bermea <i>et al.</i> 2009.
	15	693	116		Rodríguez-Salazar <i>et al.</i> 2011.
		400			Recommended by SEMARNAT (2007).
V	57	111	83	87	This study.
	50	179	97		Morton-Bermea <i>et al.</i> 2009.
	60	229	186		Rodríguez-Salazar <i>et al.</i> 2011.
		78			Recommended by SEMARNAT (2007).

^a below detection limit (dl)

pollution of the site, a value of 1-3 moderate, 4-6 considerable and >6 very high pollution.

Pollution load index (PLI)

All sites were also analyzed by a synopsis of six elements, which indicates the degree of contamination of the site. The “pollution load index” (PLI) is derived from the unique CF for each site and is calculated as:

$$PLI = \sqrt[n]{CF_1 * CF_2 * CF_3 * \dots * CF_n} \quad (2)$$

Where, CF is the TE contamination factor, n represents the number of metals analyzed.

The PLI is simple to interpret: a value of zero indicates excellent quality, a value of 1 the presence of only baseline levels of pollutants and values above 1 is a progressive deterioration of site quality (Tomlinson *et al.* 1980).

Geostatistic analysis

A georeferenced data matrix based on TE

concentration and indices was constructed. Subsequently a geostatistical analysis was performed with the Gamma Design Software (Robertson 2008), following the sequence: 1) Exploratory data analysis and 2) Spatial autocorrelation. We calculated the experimental semivariograms with the equation:

$$\gamma(h) = \frac{1}{2n} \sum_{i=1}^n [Z(x_i) - Z(x_i + h)]^2 \quad (3)$$

Where, $\gamma(h)$ is the experimental semivariance value for all pairs at a lag distance h , $Z(x_i)$ is the TE concentration and the indices values at the point i , $Z(x_i + h)$ is the TE concentration and the indices values of other points separated from x_i by a discrete distance h , x_i is the georeferenced position where the $Z(x_i)$ values were measured, n is the number of pairs of observations separated by a distance h (Isaaks and Srivastava 1989, Hernández-Stefanoni and Ponce-Hernández 2006, Delgado *et al.* 2010).

The experimental semivariance was adjusted to the theoretical model of spatial distribution of TE, concentration and TE indices.

OK was used as the interpolation method to obtain point data to plot Cr and PLI maps. IK was used as the interpolation method to obtain point data to plot TE probability maps. The variable measured on a continuous scale is converted to several indicator variables. Each variable takes a value of 1 or 0 at each sampling site and the value of each variable is estimated elsewhere in the study area, with < 1 indicating a value below the threshold level (Cetin and Kirda 2003, Longley *et al.* 2005).

The TE are continuous variables $Z(x_i)$, the indicator variable $I(x_i; z_k)$ where z_k is the desired threshold limit is defined as follows (Goovaerts 1997, 1999):

$$I(x_i; z_k) = \begin{cases} 1 & \text{if } Z(x_i) \leq z_k, \\ 0 & \text{otherwise} \end{cases}, \quad k = 1, \dots, K \quad (4)$$

Where K is the number of cutoff. The experimental indicator semivariogram, $gI^*(h)$, is then defined for every set of indicators at each cutoff z_k as:

$$\gamma_I^*(h) = \frac{1}{2N(h)} \sum_{i=1}^{N(h)} [I(x_i; z_k) - I(x_i + h; z_k)]^2 \quad (5)$$

Where $N(h)$ is the number of indicator pairs that transform $I(x_i; z_k)$ and $I(x_i + h; z_k)$ separated by vector h . The conditional cumulative distribution function at location x_0 is:

$$F(x_0; z_k | (n)) = I^*(x_0; z_k) = \sum_{i=1}^n \lambda_i I(x_i; z_k) \quad (6)$$

Where $I^*(x_0; z_k)$ is the estimated indicator level at “unsampled location” (all other locations within a raster, where no samples were collected) x_0 and λ_i is the weight assigned to the known indicator value $I(x_i; z_k)$.

In this case, permissible limits for Pb (400 mg/kg) and V (78 mg/kg) were used as threshold level, for Cu and Zn there are no defined limits (SEMARNAT 2007). A threshold value of CF = 3 was applied to Ni, Cu, Zn, Pb and PLI = 2.

The ArcGIS 9.0 (ESRI 2004) software was used for mapping. For a study of this scale level we used the UTM projection, zone 14, horizontal datum ellipsoid and the World Geodetic System 84 (WGS84).

RESULTS AND DISCUSSION

Permissible limits and indices

Mean and ranges of concentrations are comparable to others previously recorded for Mexico City (**Table I**). In all these studies, Cu, Zn and Pb show a wide range of values, whereas the concentrations of V, Cr and Ni are closer to their means (**Fig. 2**).

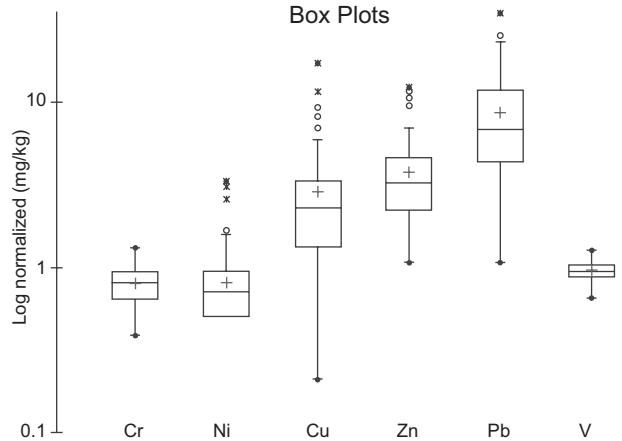


Fig. 2. Boxplot of contamination factor for toxic elements in urban soils of Mexico City. (Log normalized mg/kg. Boxes: 25th percentile and 75th percentile. Line between boxes: median. Symbol +, mean. Whisker: range. Circles: outliers. Stars: the extreme values)

Alloway (2012) specified general background values according to their different parent material. The background concentrations outside parenthesis indicate values for igneous rocks (basic). While those in round brackets represents sedimentary rocks (silt and clay): Cr 200 (90) mg/kg, Ni 150 (68) mg/kg, Cu 90 (39) mg/kg, Zn 100 (120) mg/kg, Pb 3 (23) mg/kg and V 250 (130) mg/kg.

The bedrock of MAMC is divided in two major units: basic igneous rocks, such as basalt, andesite and tuff in the south and west and lacustrine sediments in the east, from the former lake of Texcoco. The backgrounds mentioned (**Table I**) for the Valley of Mexico City by Morton-Bermea *et al.* (2009) are close to those of lacustrine sediments in Alloway (2012), with the exception of Zn. The mean concentration of V, Ni and Cr in the present study is similar to their background values published in Morton-Bermea *et al.* (2009). Accordingly, we conclude that concentrations of V, Ni and Cr are determined apparently by the parent material (soils and pollution), whereas elevated concentrations of Pb, Cu and Zn are from anthropic origin.

Skewness values indicate that only Cr (0.36) and V (0.17) are approximately normally distributed. All other metals show a positive skew towards the lower concentrations. This reinforces the assumption that the concentrations of Cr and V had a natural origin as principal component. Localized anthropic sources would act against a normal distribution.

There are some positive correlations between Cu and Zn (0.78), Cu and Pb (0.64), Zn and Ni (0.62) and Pb and Zn (0.58; **Table II**).

TABLE II. PEARSON'S CORRELATIONS MATRIX FOR TOXIC ELEMENTS

	Cr	Ni	Cu	Zn	Pb	V
Cr	1					
Ni	-0.11	1				
Cu	0.31	0.01	1			
Zn	0.21	0.62	0.78	1		
Pb	0.32	0.15	0.64	0.58	1	
V	0.51	-0.29	0.05	-0.12	0.04	1

The permissible limits established by the Ministry of Environment and Natural Resources (SEMARNAT 2007) are legal references to be applied in studies and reports, yet the limit for Pb was exceeded at four sites (4.5 %) and for V at 63 sites (70.8 %).

The limit for Cr or Ni was not exceeded at any analyzed site. Since no limits for Cu or Zn are specified in the Official Mexican Standards, the values recommended by the US EPA were considered, however the reading for Cu or Zn did not exceed the limit at any site.

These concentrations of TE are slightly lower than those reported in previous studies (**Table I**). Only 4.5 % of the samples were beyond the permissible limit of

Pb, although in 70.8 % of samples, V exceeded the Official Mexican Standards for TE concentration in soils (SEMARNAT 2007). The Official Mexican Standard (SEMARNAT 2007) is largely based on those of the US Environment Protection Agency (US-EPA 2012). The limits set is very high. Although a high load is expected, only a few samples were above the permissible limit. If we take the background into consideration, it becomes clear that the high loads of V are due to the parent material and not to anthropic pollution. Either way, the V is dangerous. However, more research is needed to elucidate the sources of potentially toxic elements.

The background values and the skewness differentiated two groups of metals: a) Cr, Ni and V with a concentration very close to the background and a normal Gaussian distribution and b) Cu, Zn and Pb with a load significantly above the background and a clear skewness. Therefore, the parent rocks cause the concentrations of Cr, Ni and V, while the concentrations of Cu, Zn and Pb are largely of human origin.

Results based on Official Mexican Standards are difficult to interpret in relation to the real load of pollution in urban soils because countries differ in their standards. One of the strictest is the Netherlands where, the standard for Ni is 615 times the Mexican Norm.

Accumulated CF means that although Cr, Ni and V are under the baseline of 1, which is the critical value between insignificant and moderate pollution, Cu (2.87), Zn (3.78) and Pb (8.59) are above it. Pb, Zn and Cu are found in high concentrations in several parts of the city (**Table III**).

With the obtained results, sites are classified considering the pollutant load, those with a CF below 1 are insignificantly polluted. Polluted sites are divided into three categories: moderate (CF 1-3), high (4-6)

TABLE III. CONTAMINATED SITES BASED ON CONTAMINATION FACTOR AND CLASSIFIED BY POLLUTION

Element	Type of data	Insignificant	Moderate	High	Very high	Total
Cr	Absolute	77	12			89
	Relative	87%	13%			100%
Ni	Absolute	71	18	2		89
	Relative	80%	20%	2%		100%
Cu	Absolute	17	72	30	6	89
	Relative	19%	81%	34%	7%	100%
Zn	Absolute	0	89	50	10	89
	Relative	0%	100%	56%	11%	100%
Pb	Absolute	0	89	77	55	89
	Relative	0%	100%	87%	62%	100%
V	Absolute	60	29			89
	Relative	67%	33%			100%

and very highly polluted (> 6). Moreover they are (inversely) cumulated, which means samples that were classified as very high are also included in the considerable and moderate categories, in order that only the two “insignificant” and “moderate” records together add up to 100 %.

Most of the samples with Cr (87 %), Ni (80 %) and V (67 %) are unpolluted, while not a single sample with Zn and Pb was classified as insignificant. This means that 100 % of the samples were contaminated with Zn and Pb. In this study, Cu shows a similar pattern as Zn and Pb, although not as immoderate.

There is no general agreement on permissible limits and therefore different thresholds have been suggested (Dosskey and Adriano 1991, Tyler 1992, De Vries and Bakker 1996). For this reason, we also used the contamination factor (CF) and the pollution load index (PLI), which are indicators that have been applied elsewhere worldwide (Caeiro *et al.* 2005, Mmolawa *et al.* 2011, Özkan 2012). The application of these two indices to this study area show a considerable TE loads but no consistent pattern.

Spatial analysis

The normal probability plot for Cr and V is sufficiently closed to linear, thus indicating that their concentrations in Mexico City are approximately normally distributed.

The nugget variance, structural variance and the coefficient of determination (r^2 ; fitness) between experimental and theoretical semivariograms were used to evaluate the model simulations. The TE concentrations and CF were represented by exponential

and spherical semivariograms (**Table IV**) with r^2 oscillating from 0.70 to 0.98. The structural variance calculated by $[C/(Co + C)] \times 100$ had values oscillating from 73 to 99.8 %. In general, the nugget variance values were low in relation to the structural variance, indicating that the spatial distribution pattern for TE and indices were well identified. Only CF of Zn had nugget values higher than 20 % (**Fig. 3**).

Figures 4 and **5** show the probability based on IK, that the permissible limit of 400 mg/kg for Pb and 78 mg/kg for V will be exceeded. Pb concentrations highlight a few hotspots in the city center and the west area. For $V > 70$ % of the sampling sites are above the permissible limit, with a gradient increasing from north to south.

Until now different CF values were considered. Insofar as the maps, however, a fixed uniform limit must be chosen. For the maps started from **figures 6** to **9** a CF of 3 is set. For this reason, in this section we can only present maps for Ni, Cu, Zn and Pb (compare **Table IV**). Cr and V have no single value in excess of 3.

The map for Ni shows bright less polluted areas, with the exception of two distinct

hotspots in the northern suburbs. Thus it becomes clear why there is a difference between Ni from both Cr and V. Cr and Ni have only a slight impact on the study area and are significantly different from the group of elements with very high anthropic loads (Cu, Zn, Pb). The maps for Cu (**Fig. 7**), Zn (**Fig. 8**) and Pb (**Fig. 9**) show dark, highly contaminated areas, in this increasing order.

A synthesis of all six of the TE shows high probability of moderate pollution in the districts (Benito

TABLE IV. CHARACTERISTICS OF SEMIVARIOGRAM MODELS OF TOXIC ELEMENTS

Parameters/ Threshold	Model	Nugget variance	Total variance	Range	Structural variance (%)	Model r^2
Pb 400 mg/kg	Exponential	3.79	20.47	15720	81.5	0.98
V 78 mg/kg	Exponential	0.200	117	5640	99.8	0.91
Ni $CF=3$	Spherical	0.0001	0.056	3970	99.8	0.70
Cu $CF=3$	Exponential	0.018	0.300	14490	94	0.96
Zn $CF=3$	Exponential	0.080	0.30	12090	73	0.95
Pb $CF=3$	Exponential	0.042	0.51	12120	91.8	0.94
PLI $CF=2$	Exponential	0.0077	0.0591	13320	87	0.91

CF = Contamination Factor; PLI = Pollution Load Index

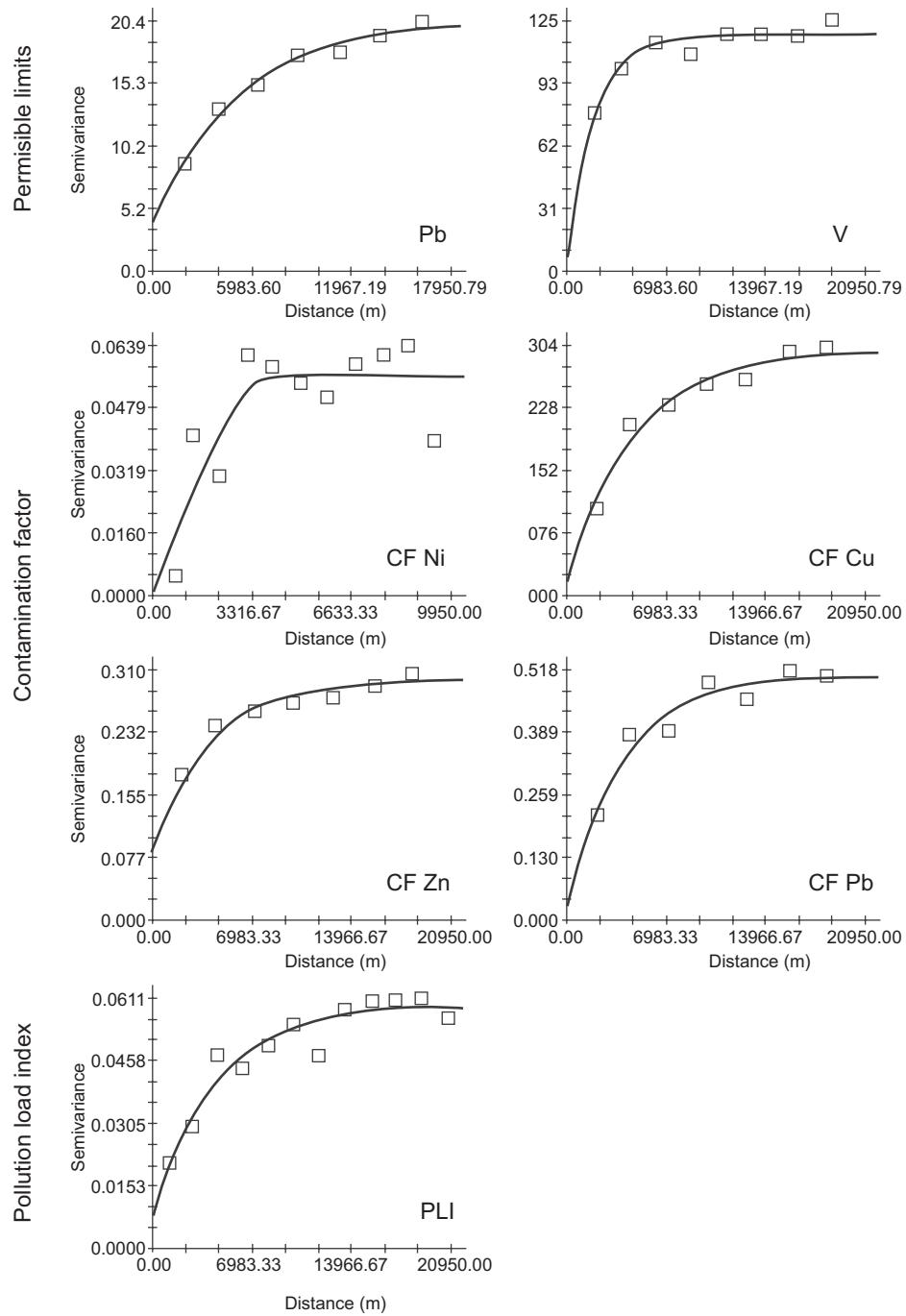


Fig. 3. Variogram models of toxic elements and indices

Juárez, Iztapalapa, Iztacalco, Venustiano Carranza, Cuauhtémoc, Alvaro Obregón and Gustavo A. Madero precincts) and the northern municipalities (Ecatepec and Tlalnepantla), while the spatially marginal areas of the metropolitan zone have lower probabilities. The map with a PLI = 2 was chosen to identify areas of urgent remediation action (Fig. 10).

Cartographic figures have a practical utility for illustration; however, selecting specific thresholds (CF, PLI) can cause visual impressions to serve a political agenda.

After ascertaining the most polluted areas, it is necessary to identify the chemical compounds of TE in the soil by chemical fractionation (Bautista 1999,

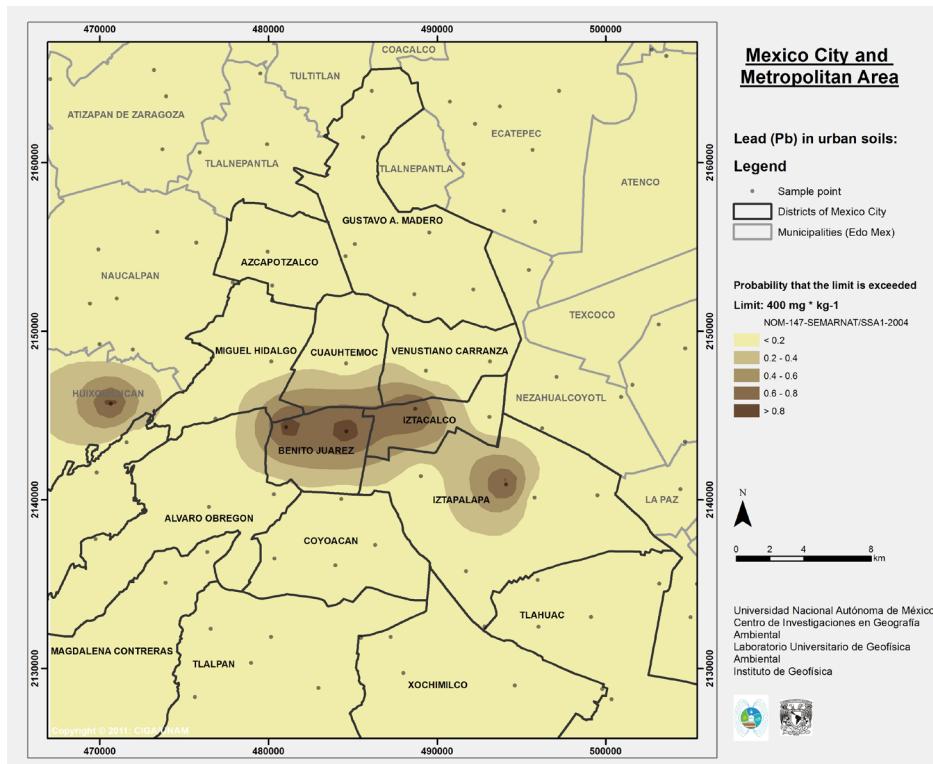


Fig. 4. Map with permissible limit of Pb (400 mg/kg)

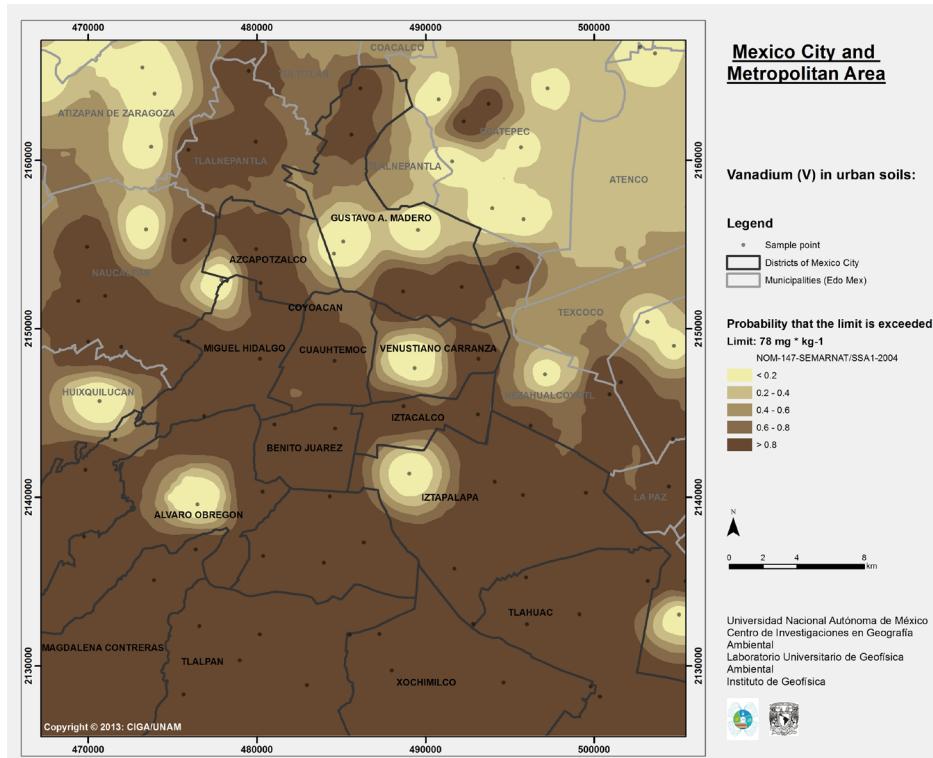


Fig. 5. Map with permissible limit of V (78 mg/kg)

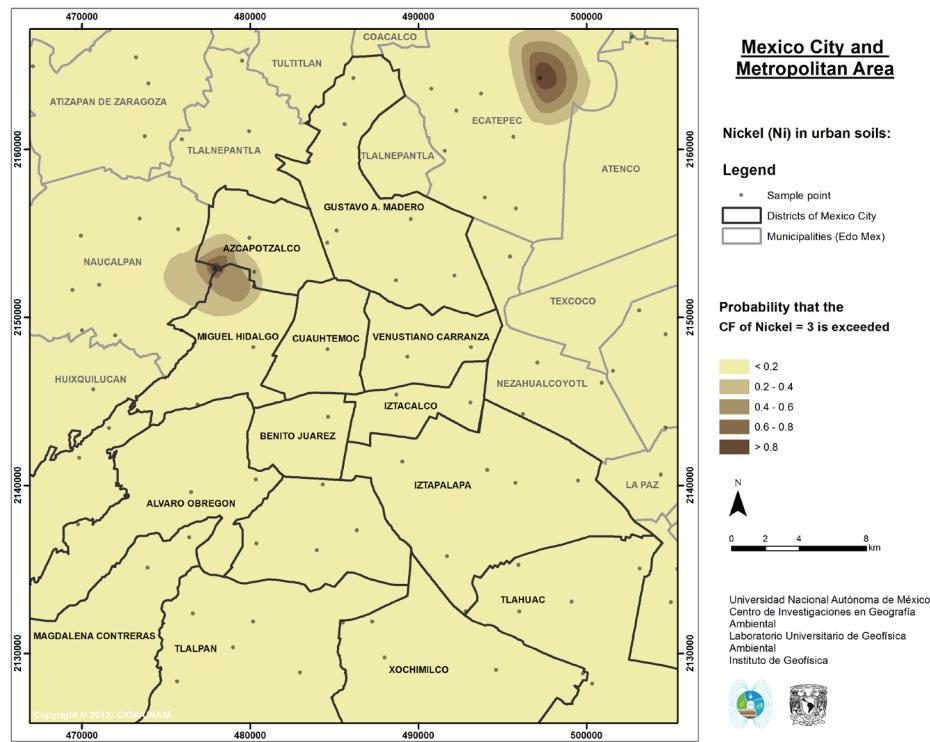


Fig. 6. Map of Ni with contamination factor of 3

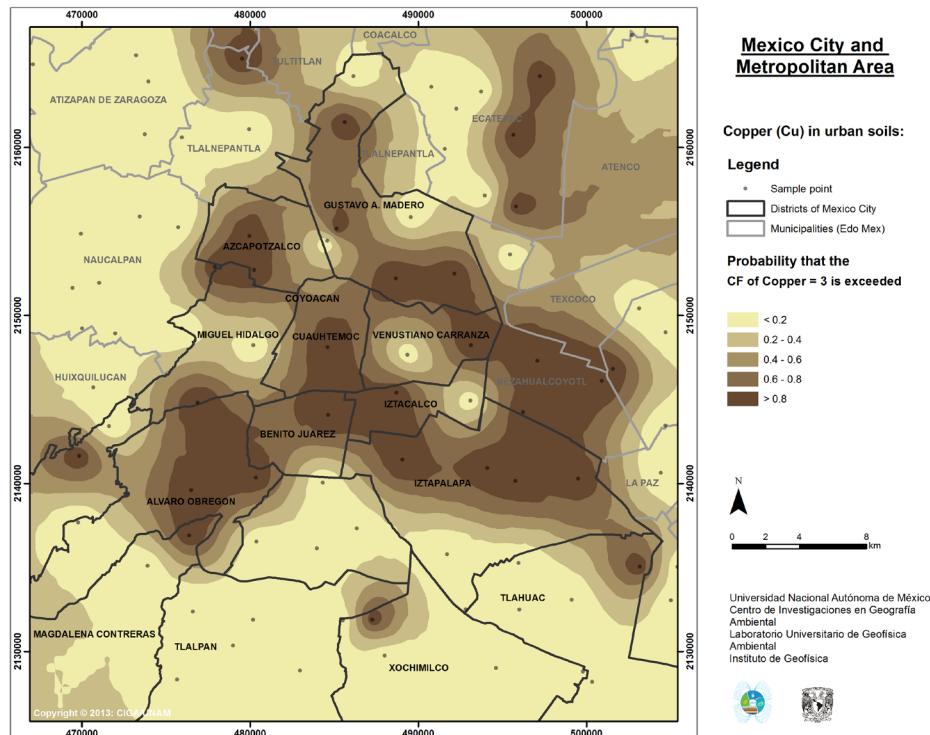


Fig. 7. Map of Cu with contamination factor of 3

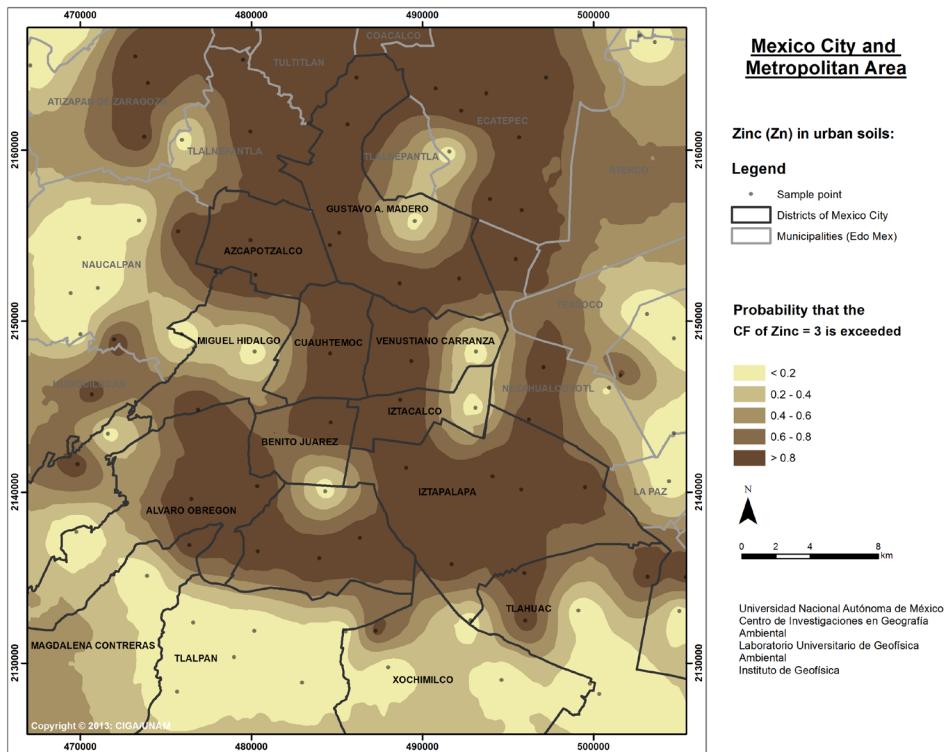


Fig. 8. Map of Zn with contamination factor of 3

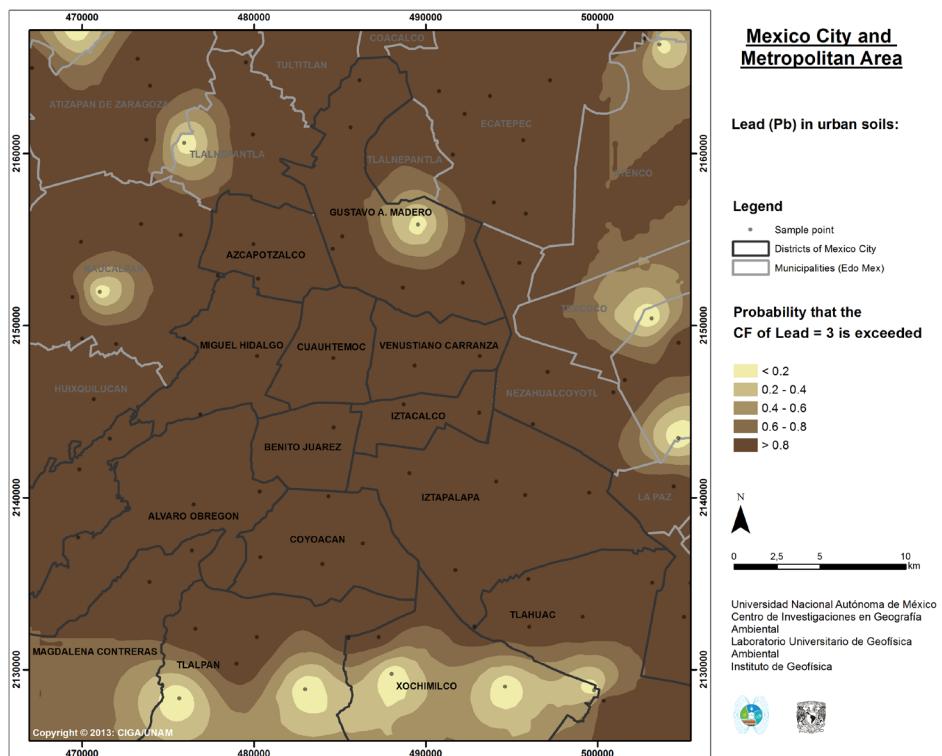


Fig. 9. Map of Pb with contamination factor of 3

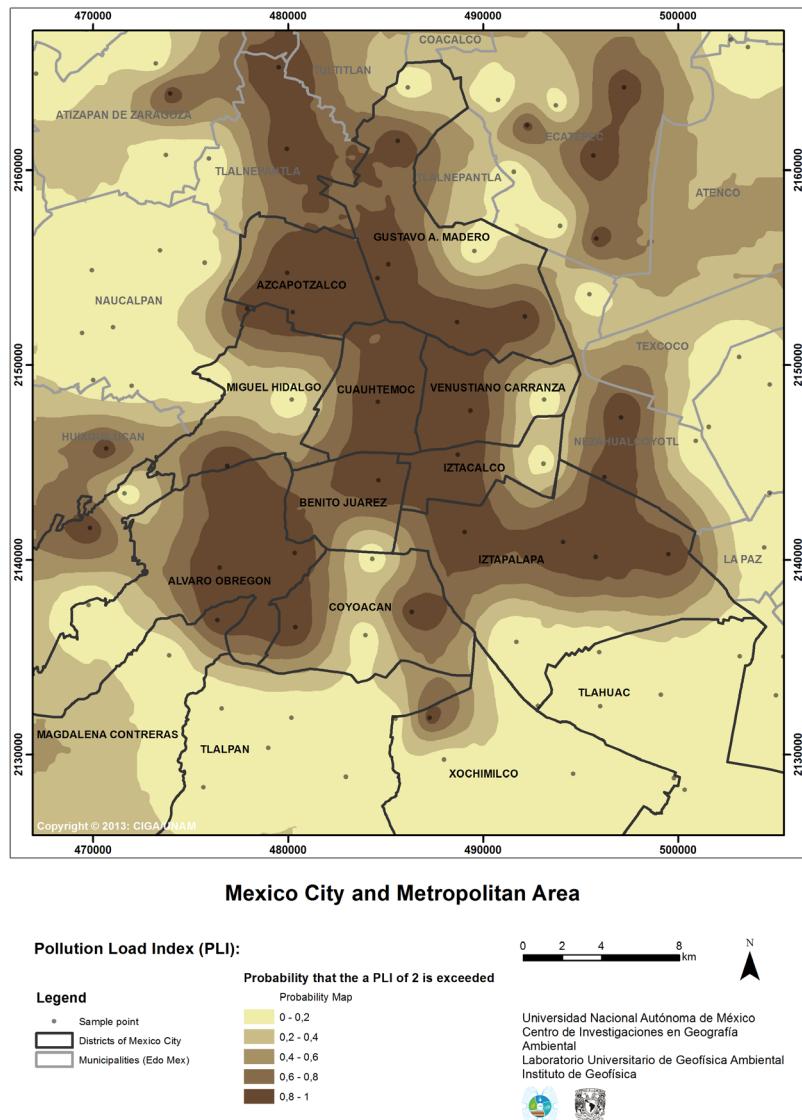


Fig. 10. Areas of soil pollution by toxic elements; map of probability that a pollution load index of 2 will be exceeded

Fitamo *et al.* 2007). This will allow the construction of a mitigation strategy in order to address the problem of soil contamination.

One of the main contributions of this study is the generation of reliable maps of an area of 1600 km² with 89 sampling points, using geostatistical tools and indicators of pollution.

CONCLUSIONS

After applying interpolation and geostatistical models according to various data management

options, the most significant findings of the study are:

The mexican maximum permissible limits are not completely reported for all TE which is concerning, as evidenced by examples of V and Pb.

High loads of Cr and V can be put into another perspective by using the CF with “backgrounds”: The high values are originated primarily from the parent rock. However, the danger of Cr and V does not depend on its origin.

The observed high loads of Cu, Zn and Pb are caused mainly by anthropic sources and human activities.

Probability maps as a result of the indicator kriging demonstrate very clearly the spatial distribution

of toxic elements all over the urbanized area and identify easily hotspots.

Ni depicts two extreme spots of pollution close to the city limits of Mexico City and Naucalpan. The second one in the municipality of Ecatepec. Based on the map, potential pollutant sources can be fixed spatially.

The TE of Cu and Zn are widespread over the central and northern parts of the urbanized area, especially in the districts of Mexico City: Benito Juárez, Iztapalapa, Iztacalco, Venustiano Carranza, Cuauhtémoc, Alvaro Obregón and Gustavo A. Madero, and the municipalities of Ecatepec and Tlalnepantla.

The high load of Pb is uniformly distributed in high concentration over the entire urbanized area. It represents an important environmental problem.

Results suggest an urgent mitigation strategy for the MAMC. The pollution load index map is an easy and clear way to illustrate the spatial relationships and demonstrate this urgency to the decision-making authorities and local stakeholders as well as to the general community.

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