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Research article

Vegetation life forms at treeline a high mountain forest in Mexico

Formas de vida de la vegetación en el límite superior de un bosque de alta montaña en México

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Abstract

The objective of the present study was to characterize the diversity of life forms and plant community structure along an altitudinal gradient of the *Pinus hartwegii* treeline ecotone on *Nevado de Toluca*, Mexico. Plant species in the tree boundary ecotone were sampled, identified and quantified based on plant morphology under adverse conditions (life forms). The diversity of life forms in the ecotone was calculated using Shannon's diversity index and the structure was evaluated in terms of density and abundance of plant species. A total of 43 plant species belonging to 30 genera and 18 families were identified. Chamaefites were the most abundant life form (>50 %); as altitude increased, life form diversity increased, while abundance decreased. Species diversity was proportionally and significantly (p<0.05) related to increasing altitude along the treeline ecotone (diffuse form), which was characterized by more pronounced abiotic stress conditions at higher altitudes due to the altitudinal range sampled and because this ecotone represents the distribution limit of forest species and the beginning of grassland distribution. The diversity and distribution of life forms present in the *P. hartwegii* treeline ecotone may modify microenvironmental conditions or processes and, therefore, they regulate the response of other species associated with climate change.

Key words: Alpine ecosystems, altitudinal gradient, altitudinal migration, Climate change, facilitation, *Nevado de Toluca*.

Resumen

El objetivo del presente estudio fue caracterizar la diversidad de las formas de vida y la estructura de la comunidad vegetal a lo largo de un gradiente altitudinal del ecotono del límite arbóreo de *Pinus hartwegii* en el Nevado de Toluca, México. Se muestrearon, identificaron y cuantificaron las especies vegetales con base en la morfología de la planta en condiciones adversas (formas de vida). La diversidad de formas de vida se calculó con el Índice de Diversidad de *Shannon*, y la estructura se evaluó en función de la densidad y abundancia de los taxa de plantas. Se identificaron 43 taxones pertenecientes a 30 géneros y 18 familias botánicas. Los caméfitos

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fueron la forma de vida más abundante (>50 %). La diversidad de las formas de vida fue mayor a medida que aumentó la altitud, pero su abundancia disminuyó. La diversidad de especies se relacionó proporcional y significativamente (*p*<0.05) con una altitud creciente a lo largo del ecotono del límite arbóreo. La altura y la densidad de los árboles de *P. hartwegii* decreció gradualmente a lo largo del gradiente muestreado (forma difusa), el cual se caracterizó por sus condiciones de más estrés abiótico a mayores altitudes, ya que este ecotono representa el límite de distribución de las especies forestales y el inicio de los pastizales. La diversidad y distribución de las formas de vida en el ecotono del límite arbóreo de *P. hartwegii* pueden modificar las condiciones o los procesos microambientales; por lo tanto, regulan la respuesta al cambio climático de otras especies asociadas.

Palabras clave: Ecosistemas alpinos, gradiente altitudinal, migración altitudinal, Cambio climático, facilitación, Nevado de Toluca.

Introduction

High-altitude Mountain ecosystems located above 2 800 m (Challenger and Soberón, 2008) contain numerous transition zones between plant communities (knows as ecotones), which are highly dynamic (Kark, 2017). Treeline, which represents the transition zone between forested areas and other types of vegetation, marks the uppermost extent of forested patches on slopes with the same aspect (Richardson and Friedland, 2009; Kark, 2017). There is not clarity about the global mechanisms that control tree growth beyond treeline. However, there is no doubt about the influence of temperature over the ecological and physiological processes of treeline in the high mountains, which gives them a high dynamic useful in the early detection of shifts induced by global warming (Holtmeier, 2009; Treml and Chuman, 2015; Kark, 2017). Nevertheless, both vegetation structure and plant composition of treeline are key for understanding the response of mountain ecosystems to changes in environmental conditions (Richardson and Friedland, 2009; Harsch and Bader, 2011). Moreover, they can promote plant growth and resilience in the face of extreme conditions, potentially extending the longevity of plant life (Wielgolaski *et al.*, 2017).

As part of an evolutionary adaptation process, many plant species in high mountain ecosystems have developed notable structural, phenotypical, and physiological modifications to mitigate the abiotic stress faced at high altitudes (Parmesan, 2006; Alberto *et al.*, 2013). These adaptations have generated a large variety of plant life forms and endemism in high mountain ecosystems (Cavieres and Badano, 2009; Körner, 2017). The diversity of life forms in these ecosystems is a crucial factor in the establishment and survival of plants across their distribution areas (Callaway *et al.*, 2002), at new sites (Cavieres and Badano, 2009; Badano *et al.*, 2015) and influences the latitudinal or altitudinal plant migrations in response to changing environmental conditions. As a result, many alpine plants develop stress-tolerance strategies such as small size or slow growth rate, buffering effects on the microenvironment, will not change much with time (Anthelme *et al.*, 2014). In addition, the altitudinal migration of a forest could be promoted or limited as a function of the plant composition in the ecotone (Alberto *et al.*, 2013; Badano *et al.*, 2015).

The upper altitudinal limit of Mexican pine (*Pinus hartwegii* Lindl.) is the highest in the world (Calderón de Rzedowski and Rzedowski, 2005). This species has adapted to extremely adverse aboveground (extremely low temperatures, freezing and abrasive winds, and high solar radiation) (Körner *et al.*, 2021) and belowground environmental conditions (shallow and unconsolidated soils, low nutrient availability and freezing soils) (Kozlowski and Pallardy, 2002; Körner *et al.*, 2021). These adaptations, however, make *Pinus hartwegii* extremely sensitive to environmental changes.

Several studies have analyzed changes in its upper altitudinal limit at the *Nevado de Toluca* and *Iztaccíhuatl-Popocatépetl* (both volcanoes in central Mexico) stemming from an average local temperature increase of 1.5 °C in the past 50 years (Pérez-Suárez *et al.*, 2022). This phenomenon is especially relevant considering that the temperature of central Mexico is expected to increase 2 °C by 2080 (Arriaga and Gómez, 2004; Alfaro-Ramírez *et al.*, 2020). Furthermore, little knowledge has been

complied on the composition, structure and functioning of the species that compose the treeline ecotone of *Pinus hartwegii* or on the presence of facilitative and competitive relationships between different life forms in this ecotone, especially in the case of annual life forms, which respond more quickly to environmental changes compared to longer living life forms such as trees.

In this context, this study aimed to know what is the composition of the plant community found along the altitudinal gradient of the treeline ecotone of *Pinus hartwegii*? What proportion of life forms accompany *Pinus hartwegii*? As well as if, is the proportion of existing life forms across the altitudinal gradient related to a reduction in *Pinus hartwegii* trees at increasing altitudes? To answer these questions, the present study determined the species diversity and the proportion of life forms of the plant community along the altitudinal gradient of the *Pinus hartwegii* treeline ecotone in the *Nevado de Toluca*, central Mexico. The resulting data could serve as a baseline for future comparisons and form the basis for more in-depth analyses of the competitive and facilitative relationships that either promote or limit the upward migration of the *Pinus hartwegii* forest. Upon understanding these relationships, more precise models could be created to potentially predict the altitudinal migration of *Pinus hartwegii* based on plant community composition of its treeline ecotone.

Materials and Methods

Study area

The study area was located in the State of Mexico in central Mexico between *Toluca* and *Tenango* valleys within the *Nevado de Toluca* Flora and Fauna Protection Area (18°51'-19°19' N and 99°38'-100°09' W; Figure 1). This area encompasses a range of altitude from 3 000 to 4 680 m (Vargas, 1984). Semicold, subhumid climate (C[E]wig) and cold climate (E[T]Hwig) are dominant (García, 1990), with average temperatures below 10 °C and low rainfall, as well as snowfall during the winter season (Challenger and Soberón, 2008). Main soil group is andosol, as well as feozem, regosol, cambisol, and litosol soils in lesser proportions (Vargas, 1984). Temperate coniferous forests are abundant and dominated by species of *Abies* Mill., *Pinus* L., and *Quercus* L. genera between 3 000 and 4 100 m, while above 3 950 and up to 4 500 m high mountain grasses are present and dominated by *Festuca* L. and *Calamagrostis* Adans. (Calderón de Rzedowski and Rzedowski, 2005).



Figure 1. Location of the *Nevado de Toluca* Flora and Fauna Protection Area in central Mexico.

Sampling design

A stratified systematic sampling design was used to subdivide the study area according to altitudinal level and to subsequently identify any patterns in the spatial variation of plant communities (Matteucci and Colma, 1982).

Five perpendicular transects of 220 m in length along the treeline ecotone were established to cover an altitudinal gradient. Transects that were accessible and that had a similar state of conservation at the treeline and a similar gradient were selected. Concerning the volcanic crater, two transects had a northwesterly orientation, and three had an easterly-southeasterly exposure.

Six 20×20 m plots were established along each transect based on the experimental model of Camarero and Fortin (2006). Three plots were located below the forest line (forested area), which were called from lowest to highest altitudes as Bos1, Bos2 and Bos3, while the other three were located above the forest line (grassy area) being called Pas4, Pas5 and Pas6 idem. Central points of each plot were separated by 40 m. Each plot was further divided into four quadrants of 10×10 m numbered clockwise from 1 to 4. In each plot, two opposite quadrants were alternately selected (1 and 3 or 2 and 4); in these quadrants, two plots of 3×3 m were randomly established to assess the herbaceous strata. Therefore, data were generated per plot and altitudinal level considering plant cover on both sides of the forest line and sampling sites orientation.

Characterization of life forms

The diversity of life forms at treeline ecotone of *Pinus hartwegii* was determined through collecting and identifying all plant species from August to October 2018. The number of individuals of each species was recorded in each plot to subsequently calculate relative abundance of each species and specific diversity was calculated of each altitudinal level according to the Shannon diversity index (H') (Buckland *et al.*,

2005). The required sampling effort was based on rarefaction and extrapolation curves generated in iNEXT Online version (Chao *et al.*, 2022) using Hill numbers (Chao *et al.*, 2014; Cox *et al.*, 2017). Hill numbers were used since it that incorporates species richness and relative abundance and allows to explicitly choose how sensitive the diversity metric is to rare species (Chao *et al.*, 2014; Cox *et al.*, 2017). We used a Hill number with q=1 (*i. e.*, Hill Shannon diversity, which emphasizes neither rare nor common species to allow characterizing gradients in biodiversity) (Chao *et al.*, 2014).

Sample robustness was measured with sample-size-based diversity accumulation curves, in which the expected Hill Shannon diversity is plotted as a function of sample size using the number of individuals (Chao *et al.*, 2014; Cox *et al.*, 2017). Coverage-based diversity curves were used as well to plot expected diversity as a function of interpolated (observed) and extrapolated (expected) coverage.

All plant species found at upper treeline ecotone of *Pinus hartwegii* were then categorized as phanerophytes, chamaephytes, hemicryptophytes, cryptophytes or therophytes life forms according to the classification of Raunkiaer (1934). This classification is based on height of plants' buds during seasons with adverse conditions. Life forms proportion at each altitudinal level, and the percentage cover (species coverage) occupied by each life form were calculated. Collected specimens were identified in the Hortorio ChAPA Herbarium of the *Colegio de Postgraduados* in *Montecillo* campus.

Structure of the alpine treeline ecotone of *Pinus hartwegii*

Each plot was georeferenced (Garmin GPSmap[®] 64s) to calculate the average altitude (altitudinal level) and to establish the treeline ecotone's limits. In all plots, tree height, normal diameter (*ND*) and tree density were determined through the moving split-window method adapted from Camarero and Fortin (2006). This method consists of calculating the difference using Euclidean distances squared, between two halves of a moving window that runs along the transect.

At following, altitudinal distribution model of the upper treeline ecotone of *Pinus* hartwegii was obtained in the Idrisi Selva software (Eastman, 2012). Altitudinal dimensions and limits of the treeline ecotone of Pinus hartwegii were defined via a supervised classification based on georeferenced points, which were also used to confirm the generated model. Finally, by using the data generated by the elevation distribution model, the form of treeline ecotone was characterized according to Harsch and Bader (2011), who recognized four possible forms: (I) Diffuse: characterized by a gradual decrease in height and tree density along the treeline ecotone; (II) Abrupt: characterized by a continuous forest (>3 m tall) directly bordering low alpine vegetation. Tree height and density thus changes rapidly; although trees may be present above the continuous forest, but their presence is infrequent; (III) Island: characterized by clumped patches or linear strips of Krummholz or trees above the continuous forest limit; or (IV) Krummholz: characterized by severely stunted or deformed multi-stemmed trees. This growth form can occur in clumped patches above the upright forest, in which case we class the treeline as an island treeline, or it can occur as a dispersed or contiguous band above the upright forest.

Statistical analysis

The Shapiro-Wilk test and the homoscedasticity by the Levene test checked the normality of the data given the number of samples (Rahman and Govindarajulu, 1997). A one-way analysis of variance (ANOVA) was carried out with tree density and species diversity (*H'*) as the main factors and with altitudinal level as the classification factor and was followed by Tukey's multiple comparison test (Zar, 1998). A linear regression was applied for the associations of the altitude-height and altitude-*ND* variables. Life forms diversity was evaluated through calculating their proportional distribution per transect and per altitude strata. All statistical analyses were performed in JMP8 (SAS, 2008) at a 95 % confidence level.

Results

Life forms diversity and richness

A total of 43 plant species belonging to 30 genera and 18 botanical families were identified *Pinus hartwegii* treeline ecotone in *Nevado de Toluca*. Asteraceae (12 species) and Poaceae (8 species) families had the highest diversity and abundance (Table 1). Dominant species were *Calamagrostis tolucensis* (Kunth) Trin. *ex* Steud., *Agrostis tolucensis* Kunth, *Festuca tolucensis* Kunth, among other (Table 1). Diversity accumulation curves were saturated for the two altitudinal levels (Figure

2a-b). Sample coverage reached an asymptote with the sample effort employed in all altitudinal levels (Figure 2c-d). Sample coverage on Hill Shannon diversity was on average \geq 98 % for all altitudinal level (Figure 2e-f).

Table 1. Families and species present in the alpine treeline ecotone of *Pinushartwegii* Lindl. at *Nevado de Toluca* Flora and Fauna Protection Area.

Family		RA (
Species	Bos1	Bos2	Bos3	Pas4	Pas5	Pas6	LF
Alliaceae							
Allium L. sp.	0	2.23	0.90	0	0.09	0	Chamaephytes
Apiaceae							
Eryngium monocephalum Cav.	3.00	2.23	5.28	6.75	9.84	10.71	Hemicryptophytes
Asteraceae							
<i>Conyza coronopifolia</i> Kunth	1.22	4.93	6.79	1.44	0.68	0.67	Chamaephytes
Conyza schiedeana (Less.) Cronquist	0.13	0.58	1.20	0.24	0	0	Chamaephytes
Erigeron galeottii (A. Gray) Greene	0.13	0.23	0	0.24	0	0	Hemicryptophytes
Eupatorium prunellifolium Kunth	0	0.23	0	0.36	0.38	0	Hemicryptophytes
Gnaphalium lavandulaceum DC.	0	0	0.60	2.77	3.11	1.25	Chamaephytes
Gnaphalium liebmannii Sch. Bip. ex Klatt	0	0	0	0.60	1.75	1.83	Hemicryptophytes
Gnaphalium oxyphyllum DC.	0.27	0.82	3.02	1.44	1.85	2.12	Hemicryptophytes
Gnaphalium sarmentosum Klatt	0.13	1.52	1.66	2.89	1.26	2.60	Hemicryptophytes
Hieracium dysonymum S. F. Blake	2.18	1.52	2.87	2.53	4.38	3.57	Hemicryptophytes
Hieracium pringlei A. Gray	0.81	0.35	1.35	0.72	1.65	2.02	Hemicryptophytes
Senecio mairetianus DC.	0.68	1.29	1.05	1.32	0.38	0.48	Phanerophytes
Senecio roseus Sch. Bip.	0	0.35	0	0	3.89	0.38	Hemicryptophytes
Asparagaceae							
Echeandia durangensis (Greenm.) Cruden	0	0.23	0	0.12	0	0	Chamaephytes
Brassicaceae							
<i>Draba</i> L. sp.	0.95	0.58	0.45	0.12	1.85	1.15	Chamaephytes
<i>Draba jorullensis</i> Kunth	0.13	1.41	1.05	1.20	3.99	1.73	Chamaephytes
Draba tolucensis Kunth	0	0.82	0.15	0.84	2.14	0.77	Chamaephytes
Caryophyllaceae							
Arenaria lanuginosa (Michx.) Rohrb.	0	0.23	0.60	1.56	1.94	1.25	Hemicryptophytes
Arenaria bryoides Willd. ex D. F. K. Schltdl.	0.81	2.23	2.11	6.99	3.70	6.08	Chamaephytes

Euphorbiaceae							
Euphorbia L. sp.	10.38	22.56	6.04	6.27	2.14	6.56	Chamaephytes
Fabaceae							
Lupinus aschenbornii S. Schauer	2.45	6.81	4.38	7.84	9.25	9.26	Terophytes
Lupinus montanus Kunth	1.63	3.29	1.51	4.82	3.60	5.21	Terophytes
<i>Lupinus</i> L. sp.	0.13	0.70	0.15	0	0	0	Terophytes
Gentianaceae							
<i>Halenia plantaginea</i> (Kunth) G. Don	0	0.35	0	0.96	0.19	2.60	Hemicryptophytes
Geranium L. sp.	0	0.11	0.15	0	0	0	Chamaephytes
Juncaceae							
<i>Luzula racemosa</i> Desv.	0	0.35	0.90	3.61	2.82	3.76	Chamaephytes
Orobanchaceae	0 13	0	0.30	0.36	0	0.00	Chamaonhytos
	0.15	0	0.50	0.50	0	0.09	Chanaephytes
Oxalia alpina (Rose) Rose ex R. Knuth	24.31	11.75	8.30	13.51	15.20	13.80	Cryptophytes
Plantaginaceae							
Penstemon gentianoides (Kunth) Poir.	0.54	3.52	0.45	0.24	0	0	Phanerophytes
<i>Plantago tolucensis</i> Pilg.	0.54	0	0.60	0	0.58	0.19	Chamaephytes
Poaceae							
Agrostis tolucensis Kunth	2.04	1.99	6.79	1.80	2.24	1.93	Chamaephytes
Blepharoneuron tricholepis (Torr.) Nash	1.22	0.23	1.20	0.12	0	0.57	Chamaephytes
<i>Calamagrostis orizabae</i> (Rupr. <i>ex</i> E. Fourn.) Beal	0.81	0.58	1.51	0.48	0	0.19	Chamaephytes
<i>Calamagrostis tolucensis</i> (Kunth) Trin. <i>ex</i> Steud.	23.08	16.45	22.20	17.73	13.45	12.93	Chamaephytes
Festuca L. sp.	4.50	0.23	4.07	1.44	0.29	0.67	Chamaephytes
Festuca tolucensis Kunth	7.65	5.05	3.02	4.10	5.94	5.11	Chamaephytes
Poa annua L.	0.13	0.11	0.15	0.12	0	0.09	Chamaephytes
Trisetum spicatum (L.) K. Richt.	3.00	1.29	2.56	2.53	0.87	0.96	Chamaephytes
Rosaceae							
Alchemilla vulcanica Schltdl. & Cham.	0	1.99	4.83	1.56	0.09	0.09	Chamaephytes
<i>Potentilla richardii</i> Lehm.	6.83	0.35	1.05	0	0.09	0.09	Chamaephytes
Solanaceae							
<i>Physalis</i> L. sp.	0	0.35	0.60	0.24	0.19	0.09	Chamaephytes
Pinaceae							
<i>Pinus hartwegii</i> Lindl.	19.90	13.40	15.80	4.20	1.30	0.50	Phanerophytes

Relative abundance (*RA*, %) by altitudinal level and life form (LF) of registered species. Bold numbers indicate the highest abundance values.



a-b = Hill Shannon diversity as a function of number of individuals (*i. e.*, sample-sized-based); c-d = Sample coverage for rarefied samples as a function of number of individuals; e-f = Hill Shannon diversity as a function of sample coverage (*i. e.*, coverage-based for each altitudinal level). Continuous lines represent interpolation (*i. e.*, observed) and dashed lines represent extrapolation (*i. e.*, expected) of species diversity.

Figure 2. Diversity accumulation curves for each altitudinal level.

Altitudinal gradients and life forms composition

Bunch grasses had highest cover throughout the ecotone, and *Calamagrostis tolucensis* (20-60 % cover), *Festuca tolucensis* (10-30 %), *Luzula racemosa* Desv. (10-35 %), and *Agrostis tolucensis* (2-15 %) were the most dominant in all plots. Chamaephytes were the dominant life form throughout *Pinus hartwegii* treeline ecotone (56 % of total recorded species), followed by hemicryptophytes (26 %), phanerophytes (16 %), and cryptophytes (2 %). Cover of life forms changed as increasing altitude, phanerophyte cover decreased by 91 %, while chamaephyte cover increased by 54 % (Figure 3a). However, the proportion of phanerophyte cover reduced by 94 %, and chamaephyte remained stable along the altitudinal gradient (Figure 3b).



a = Dominant life forms per altitudinal level in forest (Bos1, Bos2, Bos3) and grassland (Pas4, Pas5, Pas6); b = Species belonging to different life forms in *Pinus hartwegii* Lind. treeline ecotone at *Nevado de Toluca*.

Figure 3. Relative covers.

Highest diversity (H'=2.13) was recorded at altitudinal level Pas6 (4 096 m), while the lowest diversity (H'=1.5) was recorded at altitudinal level Bos1 (4 043 m; Figure 4). Shannon diversity index was positively related with increasing altitude at plots with a northwesterly aspect (F=15.22, P=0.0022) and at plots with an easterly-southeasterly exposure (F=8.47, P=0.0012; Figure 5). Diversity of sample sites also differed significantly (F=3.67, P=0.017); sites 1, 2, and 3, which had an easterly-southeasterly exposure, had greatest diversity (Figure 5).



- a = Easterly-southeasterly aspect; b = Northwesterly aspect. Means with common letter are not significantly different ($P \le 0.05$).
- **Figure 4.** Shannon's diversity index (*H'*) of plant life forms by altitudinal level, in forest (Bos1, Bos2, Bos3) and grassland (Pas4, Pas5, Pas6) covers at sampling sites.



S1, S2, and S3 = Easterly-southeasterly aspect; S4 and S5 = Northwesterly aspect. Means with common letter are not significantly different ($P \le 0.05$).

Figure 5. Shannon's diversity index (H') of life forms at sampling sites

Characterization of ecotone structure

The lower limit of *Pinus hartwegii* treeline ecotone of *Nevado de Toluca* was located at 3 980 to 4 090 m and its upper limit at 4 030 to 4 130 m. Tree density decreased significantly along the altitudinal gradient from 70 % at the lower limit to 8 % at the upper limit, indicating that *Pinus hartwegii* conforms a diffuse ecotone where in the tree density gradually decrease as altitude increases. Forested area of altitudinal level Bos1 (4 043 m) had the highest average tree density (925 trees ha⁻¹), while altitudinal level Pas6 (4 096 m) had the lowest average tree density (25 trees ha⁻¹).

Tree height and *ND* were also significantly and negatively related with increasing altitude (r^2 =0.21, P=0.0001; r^2 =0.12, P=0.0001, respectively). Tree height reduced by 48 % from the lowest to the highest altitudinal range, and the *ND* also reduced by 20 %. Along the altitudinal gradient of the ecotone, tree height ranged from 0.3 to 22.0 m, and average tree height was 3.88 m. Both shortest and tallest trees were recorded at 4 043 m (Bos1). Overall, *ND* of trees ranged from 1.0 to 72.0 cm, and average *ND* was 11.82 cm. Largest trees ranged from 50 to 72 cm in diameter and were recorded at the Bos1 (4 043 m) and Bos2 (4 050 m) altitudinal levels.

Diffuse structure of the *Pinus hartwegii* ecotone was characterized by an increasing diversity and cover of high-mountain grasses species, such as Eryngium monocephalum Cav., Lupinus aschenbornii S. Schauer, Arenaria bryoides Willd. ex D. F. K. Schltdl., Oxalis alpina (Rose) Rose ex R. Knuth, Draba tolucensis Kunth, and Poa annua L., among others, with increasing altitude (Table 1). These species increased in abundance by up to 80 % over the altitudinal gradient, and their cover also increased by up to 90 %. Species accompanying Pinus hartwegii at lower altitude levels, including *Festuca tolucensis* and *Trisetum spicatum* (L.) K. Richt. gradually reduced their cover by 10 to 40 % with increasing altitude. Meanwhile, species such as Calamagrostis tolucensis and Gnaphalium sarmentosum Klatt gradually increased their cover by 5 to 30 % with increasing altitude. Similarly, the abundance of species such as Poa annua and Draba tolucensis above the forest line, which were recorded in plots Pas4, Pas5, and Pas6 (4 072 to 4 096 m), gradually increased by 10 to 35 % as altitude increase. Other species such as Senecio mairetianus DC. and Castilleja *lithospermoides* Kunth had a relatively stable relative abundance and cover along the altitudinal gradient of the treeline ecotone of *Pinus hartwegii*.

Discussion

Pinus hartwegii, native to the mountains of Mexico (Calderón de Rzedowski and Rzedowski, 2005), is therefore an ideal model for studying the mechanisms that regulate the survival of tree life forms at high altitudes (Körner and Paulsen, 2004; Pérez-Suárez *et al.*, 2022) and the impacts of climate change on species distribution. At the *Nevado de Toluca*, *Pinus hartwegii* forms a diffuse ecotone that is characterized by the convergence of tree life forms, or therophytes, and herbaceous life forms, or chamaephytes, such as *Calamagrostis tolucensis*, which dominates high-mountain grasses (Calderón de Rzedowski and Rzedowski, 2005). The capacity of this transition zone to respond to changes in environmental conditions is closely related with the composition and proportion of life forms along the altitudinal gradient.

The study of plant life forms makes it possible to identify the interactions between species and environment (Treml and Chuman, 2015; Kark, 2017; Körner, 2017), as well as their morphological and physiological adaptations (Heslop-Harrison, 2017; dos Santos *et al.*, 2022), which allows to observe the selection and adaptation processes that could regulate the survival of plant species at increasing temperatures (Körner and Paulsen, 2004; Brooker *et al.*, 2008). In the present study, therophytic life forms such as *Lupinus montanus* Kunth and *L. aschenbornii* were found to survive under unfavorable conditions in seed form; these species then subsequently emerge or germinate when environmental conditions are more favorable (Parmesan, 2006). The effects of increasing temperatures on the survival of distinct species, including those associated with *Pinus hartwegii*, are important to understand and to assess the potential response of species to climate changes (Ramírez-Contreras and Rodríguez-Trejo, 2009).

Notably, herbaceous life forms have developed adaptations that allow them to germinate, establish, and grow in shallow soils with scarce nutrients and freezing temperatures (Kozlowski and Pallardy, 2002; Körner *et al.*, 2021). These life forms are also able of modifying the microsites where they establish (Badano *et al.*, 2015), thereby directly or indirectly favoring the establishment of other life forms (Brooker *et al.*, 2008; Anthelme *et al.*, 2014). Herbaceous plants may create viable microsites for other species and facilitate the colonization process of more complex life forms such as *Pinus hartwegii* (Ramírez-Contreras and Rodríguez-Trejo, 2009). Thus, life forms proportion along an ecotone can potentially change the structure of an ecosystem and its processes (Badano *et al.*, 2015; Körner, 2017) and influence altitudinal migration of species (Aitken *et al.*, 2008).

Density and diversity of plant life forms observed in our study within of *Pinus hartwegii* treeline ecotone at *Nevado de Toluca* evidence the diffuse form of this ecotone. This ecotone is characterized by a gradual reduction in density of trees and a gradual increase in plants of smaller dimensions such as herbaceous plants and bunch grasses (Harsch and Bader, 2011; Alfaro-Ramírez *et al.*, 2017). Diffuse ecotones are the most common around the world (Harsch and Bader, 2011), and a frequent altitudinal variation have been reported within these ecotones (Bader *et al.*, 2007), which suggest an increase in facilitation relationships between herbaceous and forest species (Callaway *et al.*, 2002; Brooker *et al.*, 2008; Badano *et al.*, 2015), as well as changes in composition and life forms proportion in the altitudinal gradient (Callaway, 2007; Arzac *et al.*, 2011).

Along the treeline ecotone of *Nevado de Toluca*, chamaephytic life forms are dominant. Chamaephytes can preserve their buds under a layer of dry leaves and can more quickly emerge during favorable climate conditions (Kozlowski and Pallardy, 2002) than plants found in seed form (Parmesan, 2006; Maestre *et al.*, 2009). In the study area, greatest number of recorded species belonged to the Asteraceae family, species that are chamaephytic in life form.

Species diversity was greatest above the forest line in areas with predominant grass cover. This is partially due to the transition between ecosystems and the convergence of species at their distributional limits, which led to high species richness (Cavieres and Badano, 2009; Körner, 2017). However, some species as *Muhlenbergia hintonii* Swallen, *M. vaginata* Swallen and *Agrostis calderoniae* Acosta Cast. associated with *Pinus hartwegii* in the forest nucleus (Calderón de Rzedowski and Rzedowski, 2005), were not found within the treeline ecotone. In addition, some species abundant at the lower portion of the ecotone, such as *Festuca tolucensis* and *Trisetum spicatum*, were also found at higher altitudes, although their abundance and cover decreased with increasing altitude.

In contrast, the abundance of other species, such as *Calamagrostis tolucensis* and *Gnaphalium sarmentosum*, decreased at low altitude yet increased at higher altitudes. Consequently, as altitude increases, the coverage of small, cold-resistant species, including several bunch grasses, also increases (Calderón de Rzedowski and Rzedowski, 2005; Arzac *et al.*, 2011). In a similar way, the diversity of cushion plants, such as *Arenaria bryoides*, increased with increasing altitudes, as well as those that form rosettes, such as *Gnaphalium sarmentosum* and *Hieracium dysonymum* S. F. Blake (chamaephytic plants). These plant forms are adapted to environmental conditions such as extreme low temperatures and intense winds, which are frequent at high altitudes (Körner, 2017; Körner *et al.*, 2021).

In the present study, the dominant group of chamaephytes may regulate microsite conditions (Cavieres and Badano, 2009; Heslop-Harrison, 2017; dos Santos *et al.*, 2022) and may possibly facilitate the establishment of other groups, thereby compensating for the negative effects of extreme environmental conditions (Brooker *et al.*, 2008). As climate changes, chamaephytes that have a facilitative relationship with forest species such as *Pinus hartwegii* could enable these species to establish at increasingly higher altitudes (Maestre *et al.*, 2009; Ramírez-Contreras and

Rodríguez-Trejo, 2009). Even so, altitudinal migrations are extremely complex and depend on numerous factors that could also potentially limit the capacity of *Pinus hartwegii* to colonize new sites at higher altitudes. This relationship is evidence that particular species and life forms that can favor the colonization process of *Pinus hartwegii* at new sites and at greater altitudes in *Nevado de Toluca*, especially given the changing climate.

Conclusions

The diversity and distribution of life forms along alpine treeline ecotones influence the capacity of species present in these environments to react to changing environmental conditions. Ecotones with a diffuse structure are the most abundant around the world, such as that of treeline of *Pinus hartwegii* in *Nevado de Toluca*. Altitudinal advance and recession of plants has been frequently reported in such diffuse ecotones, favoring the increase in the diversity of life forms found in this transition zone. The present study found a consistent pattern in the life forms accompanying *Pinus hartweqii* along its altitude gradient. As altitude increases, the abundance of forest species along the altitudinal gradient of Pinus hartwegii reduces, and P. *hartwegii* is increasingly accompanied by high-mountain grasses indicating a potential relationship between life forms yet not necessarily between species. However, Lupinus montanus, Calamagrostis tolucensis, Festuca tolucensis, and Arenaria bryoides, presented important relationships with the presence of Pinus hartwegii. Considering the changing climate in central Mexico, Pinus hartwegii could be benefitted by certain life forms like chamaephytes that encourage its establishment at increasingly higher altitudes. The results of this work show a close relationship between chamaephytes and terophytes in the *Pinus hartwegii* treeline ecotone, so it would be valuable to assess further this interaction between life forms.

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Conflicts of Interest

The authors declare no conflict of interest.

Author contributions

Marlín Pérez-Suárez, Farid Uriel Alfaro-Ramírez, Jorge Enrique Ramírez-Albores and Francisca Ofelia Plascencia-Escalante: conceived and designed the study; Farid Uriel Alfaro-Ramírez and Jorge Enrique Ramírez-Albores: collected the data. All authors analyzed the data, wrote, reviewed the manuscript and provided suggestions for it.

Referencias

Aitken, S. N., S. Yeaman, J. A. Holliday, T. Wang and S. Curtis-McLane. 2008. Adaptation, migration or extirpation: climate change outcomes for tree populations. Evolutionary Applications 1(1):95-111. Doi: 10.1111/j.1752-4571.2007.00013.x. Alberto, F. J., S. N. Aitken, R. Alía, S. C. González-Martínez, ... and O. Savolainen. 2013. Potential for evolutionary responses to climate change-evidence from tree populations. Global Change Biology 19(6):1645-1661. Doi: 10.1111/gcb.12181. Alfaro-Ramírez, F. U., J. E. Ramírez-Albores, J. J. Vargas-Hernández, S. Franco-Maass and M. Pérez-Suárez. 2020. Potential reduction of Hartweg's Pine (Pinus Lindl.) geographic distribution. PloS ONE 15(2):1-18. hartwegii Doi: 10.1371/journal.pone.0229178.

Alfaro-Ramírez, F. U., J. T. Arredondo-Moreno, M. Pérez-Suárez and Á. R. Endara-Agramont. 2017. *Pinus hartwegii* Lindl. treeline ecotone: structure and altitudinal limits at Nevado de Toluca, Mexico. Revista Chapingo Serie de Ciencias Forestales y del Ambiente 23(2):261-273. Doi: 10.5154/r.rchscfa.2016.10.055.

Anthelme, F., L. A. Cavieres and O. Dangles. 2014. Facilitation among plants in alpine environments in the face of climate change. Frontiers in Plant Science 5:387. Doi: 10.3389/fpls.2014.00387.

Arriaga, L. y L. Gómez. 2004. Posibles efectos del cambio climático en algunos componentes de la biodiversidad de México. In: Martínez, J. y A. Fernández (Coord.). Cambio climático: una visión desde México. Instituto Nacional de Ecología (INE) y Secretaría de Medio Ambiente y Recursos Naturales (Semarnat). Coyoacán, México D. F., México. pp. 255-265.

Arzac, A., E. Chacón-Moreno, L. D. Llambí y R. Dulhoste. 2011. Distribución de formas de vida de plantas en el límite superior del ecotono bosque paramo en los Andes tropicales. Ecotrópicos 24(1):26-46. http://www.saber.ula.ve/handle/123456789/35696. (3 de septiembre de 2018).
Badano, E. I., R. O. Bustamante, E. Villarroel, P. A. Marquet and L. A. Cavieres. 2015. Facilitation by nurse plants regulates community invasibility in harsh

environments. Journal of Vegetation Science 26(4):756-767. Doi: 10.1111/jvs.12274.

Bader, M. Y., M. Rietkerk and A. K. Bregt. 2007. Vegetation structure and temperature regimes of tropical alpine treelines. Arctic, Antarctic, and Alpine Research 39(3):353-364. Doi: https://doi.org/10.1657/1523-0430(06-055)[BADER]2.0.CO;2.

Brooker, R. W., F. T. Maestre, R. M. Callaway, C. L. Lortie, ... and R. Michalet. 2008. Facilitation in plant communities: the past, the present, and the future. Journal of Ecology 96(1):18-34. Doi: 10.1111/j.1365-2745.2007.01295.x.

Buckland, S. T., A. E. Magurran, R. E. Green and R. M. Fewster. 2005. Monitoring change in biodiversity through composite indices. Philosophical Transactions of the Royal Society B 360:243-254. Doi: 10.1098/rstb.2004.1589.

Calderón de Rzedowski, G. y J. Rzedowski. 2005. Flora fanerogámica del Valle de México. Instituto de Ecología, A. C. y Comisión Nacional para el Estudio de la Biodiversidad (Conabio). Pátzcuaro, Mich., México. 1406 p.

Callaway, R. M. 2007. Positive interactions and interdependence in plant communities. Springer. Dordrecht, ZH, Netherlands. 404 p.

Callaway, R. M., R. W. Brooker, P. Choler, Z. Kikvidze, ... and B. J. Cook. 2002. Positive interactions among alpine plants increase with stress. Nature 417(6891):844-848. Doi: 10.1038/nature00812.

Camarero, J. J. y M. J. Fortin. 2006. Detección cuantitativa de fronteras ecológicas y ecotonos. Ecosistemas 15(3):76-87. https://www.revistaecosistemas.net/index.php/ecosistemas/article/view/164. (31 de agosto de 2018).

Cavieres, L. A. and E. I. Badano. 2009. Do facilitative interactions increase species richness at the entire community level? Journal of Ecology 97(6):1181-1191. Doi: 10.1111/j.1365-2745.2009.01579.x.

Challenger, A. y J. Soberón. 2008. Los ecosistemas terrestres. In: Sarukhán, J. (Coord.). Capital natural de México, Vol. I: Conocimiento actual de la biodiversidad. Comisión Nacional para el Estudio de la Biodiversidad (Conabio). Tlalpan, México D. F., México. pp. 87-108.

Chao, A., K. H. Ma and T. C. Hsieh. 2022. A brief introduction to iNEXT Online: software for interpolation and extrapolation of species diversity. Institute of Statistics, National Tsing Hua University. http://chao.stat.nthu.ed.tw/wordpress/software-download/inext-online/. (11 de octubre de 2023).

Chao, A., N. J. Gotelli, T. C. Hsieh, E. L. Sander, ... and A. M. Ellison. 2014. Rarefaction and extrapolation with Hill numbers: a framework for sampling and estimation in species diversity studies. Ecological monographs 84(1):45-67. Doi: 10.1890/13-0133.1.

Cox, K. D., M. J. Black, N. Filip, M. R. Miller, ... and S. E. Dudas. 2017. Community assessment techniques and the implications for rarefaction and extrapolation with Hill numbers. Ecology and Evolution 7(24):11213-11226. Doi: 10.1002/ece3.3580.

dos Santos, P., M. Â. Brilhante, T. F. E. Messerschmid, H. C. Serrano, ... and J. de Vos. 2022. Plant growth forms dictate adaptations to the local climate. Frontiers in Plant Science 13:1023595. Doi: 10.3389/fpls.2022.1023595.

Eastman, J. R. 2012. Idrisi Selva. Guía para SIG y procesamiento de imágenes. Manual Versión 17. Clark University. Worcester, MA, Estados Unidos de América. 321 p. https://clarklabs.org/wp-content/uploads/2016/10/IDRISI-Selva-Spanish-Manual.pdf. (14 de septiembre de 2018).

García, E. 1990. Carta de climas. Atlas Nacional de México. Instituto de Geografía de la Universidad Nacional Autónoma de México. Coyoacán, México D. F., México. 116 p. Harsch, M. A. and M. Y. Bader. 2011. Treeline form-a potential key to understanding treeline dynamics. Global Ecology and Biogeography 20(4):582-596. Doi: 10.1111/j.1466-8238.2010.00622.x.

Heslop-Harrison, J. S. 2017. Morphology, adaptation and speciation. Annals of Botany 120(5):621-624. Doi: 10.1093/aob/mcx130.

Holtmeier, F. K. 2009. Mountain timberlines: Ecology, patchiness, and dynamics (Vol. 36). Springer Dordrecht. Geneva, GE, Switzerland. 438 p.

Kark, S. 2017. Effects of ecotones on biodiversity. In: Reference Module in Life Sciences. Encyclopedia of Biodiversity. Elsevier. Amsterdam, NH, Netherlands. pp. 142-148.

Körner, C. 2017. Alpine ecosystems. In: Reference Module in Life Sciences. Encyclopedia of Biodiversity. Elsevier. Amsterdam, NH, Netherlands. pp. 148-157.

Körner, C. and J. Paulsen. 2004. A world-wide study of high altitude treeline temperatures. Journal of Biogeography 31(5):713-732. Doi: 10.1111/j.1365-2699.2003.01043.x.

Körner, C., D. Urbach and J. Paulsen. 2021. Mountain definitions and their consequences. Alpine Botany 131:213–217. Doi: 10.1007/s00035-021-00265-8.

Kozlowski, T. T. and S. G. Pallardy. 2002. Acclimation and adaptive responses of woody plants to environmental stresses. The Botanical Review 68:270-334. Doi: 10.1663/0006-8101(2002)068[0270:AAAROW]2.0.CO;2.

Maestre, F. T., R. M. Callaway, F. Valladares and J. L. Lortie. 2009. Refining the stress-gradient hypothesis for competition and facilitation in plant communities. Journal of Ecology 97(2):199-205. Doi: 10.1111/j.1365-2745.2008.01476.x.

Matteucci, S. D. y A. Colma. 1982. Metodología para el estudio de la vegetación. Secretaría General de la Organización de los Estados Americanos y Programa Regional de Desarrollo Científico y Tecnológico. Washington, DC, Estados Unidos de América. 159 p.

https://aulavirtual.agro.unlp.edu.ar/pluginfile.php/76505/mod_resource/content/3/ MatteucciColma1982.pdf. (26 de agosto de 2018).

Parmesan, C. 2006. Ecological and evolutionary responses to recent climate change.

Annual Review of Ecology Evolution and Systematics 37:637-669. Doi: 10.1146/annurev.ecolsys.37.091305.110100.

Pérez-Suárez, M., J. E. Ramírez-Albores, J. J. Vargas-Hernández and F. U. Alfaro-Ramírez. 2022. A review of the knowledge of Hartweg's Pine (*Pinus hartwegii* Lindl.): current situation and the need for improved future projections. Trees 36(1):25-37. Doi: 10.1007/s00468-021-02221-9.

Rahman, M. M. and Z. Govindarajulu. 1997. A modification of the test of Shapiro and Wilk for normality. Journal of Applied Statistics 24(2):219-236. Doi: 10.1080/02664769723828.

Ramírez-Contreras, A. y D. A. Rodríguez-Trejo. 2009. Plantas nodriza en la reforestación con *Pinus hartwegii* Lindl. Revista Chapingo Serie Ciencias Forestales y del Ambiente 15(1):43-48. https://www.scielo.org.mx/pdf/rcscfa/v15n1/v15n1a5.pdf. (13 de septiembre de 2022).

Raunkiaer, C. 1934. The life forms of plants and statistical plant geography. Oxford at the Clarendon Press. Oxford, OX, United Kingdom. 632 p.

Richardson, A. D. and A. J. Friedland. 2009. A review of the theories to explain arctic and alpine treelines around the world. Journal of Sustainable Forestry 28(1-2):218-242. Doi: 10.1080/10549810802626456.

SAS. 2008. JMP Statistical Discovery (Version 8.0). Cary, NC, United States of America. SAS Institute Inc. http://www.jmp.com/es/. (14 de septiembre de 2022).

Treml, V. and T. Chuman. 2015. Ecotonal dynamics of the altitudinal forest limit are affected by terrain and vegetation structure variables: An example from the Sudetes Mountains in central Europe. Artic, Antarctic, and Alpine Research 47(1): 133-146. Doi: 10.1657/AAAR0013-108.

Vargas M., F. 1984. Parques Nacionales de México y Reservas Equivalentes. Pasado, presente y futuro. Instituto de Investigaciones Económicas de la Universidad Nacional Autónoma de México. Coyoacán, México D. F., México. 266 p.

Wielgolaski, F. E., K. Laine, J. Inkeröinen and O. Skre. 2017. Enhancing the

resilience capacity of sensitive mountain forest ecosystems under environmental change (SENSFOR). Climate Research 73(1-2):3-6. Doi: 10.3354/cr01482. Zar, J. H. 1998. Biostatistical analysis. Pearson College. New Jersey, NJ, United States of America. 663 p.

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