

Restoration of a degraded oak forest in Mexico City by introducing tree native species



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Abstract

Background: Uncontrollable urban expansion and population growth converted Mexico City into one of the largest cities of the world. Ravines remain important oak forest relicts that should be conserved and restored to recover the biodiversity and the ecological services they provide.

Question: Is an open area equally appropriate for ecological restoration as is the forest edge? Does plant performance of introduced species vary between the open area and the forest edge?

Study species: *Prunus serotina*, *Quercus rugosa*, and *Crataegus mexicana*.

Study site: Oak forest in Barranca de Tarango, Mexico City; July 2010 to July 2012.

Methods: Plants of these species were planted in the forest edge and in an open area adjacent to it, where soil properties and light conditions were measured.

Results: Soil properties and leaf area index were better and significantly larger, respectively, in the forest edge than in the open area. Stem height, basal area and number of sprouts of these plants did not differ between sites or years. Plants of *C. mexicana* produced the largest number of sprouts, followed by *Q. rugosa*. Survival of plants was similar in the open area and the forest edge; however, *P. serotina* had the lowest survival, and *C. mexicana* the highest.

Conclusions: *Q. rugosa* and *C. mexicana* with higher survival are more likely to perform better in both sites than *P. serotina*, on account of their ability to produce more sprouts. The forest edge is better than the open area for introducing native species in Barranca de Tarango. Low performance of these species was mainly due to the construction of a highway that crosses the study area.

Key words: forest edge, oak forest, ravine, restoration, sprouts.

Restauración de un bosque de encino degradado en la Ciudad de México mediante introducción de especies arbóreas nativas

Resumen

Antecedentes: La expansión urbana y el crecimiento poblacional convirtieron a la Ciudad de México en una de las más grandes del mundo. Las barrancas, como relictos importantes de bosques de encinos, deben ser conservadas y restauradas para recuperar la biodiversidad y los servicios ambientales que proporcionan.

Preguntas: ¿Una zona abierta es igualmente apropiada para la restauración ecológica como lo es el borde del bosque? ¿El desempeño de especies introducidas varía entre la zona abierta y el borde del bosque?

Especies de estudio: *Prunus serotina*, *Quercus rugosa* y *Crataegus mexicana*.

Sitio de estudio: Bosque de encino en la Barranca de Tarango, Ciudad de México; julio de 2010 a julio 2012.

Métodos: Estas especies se plantaron en el borde del bosque y en un área abierta adyacente a éste, donde se midieron propiedades del suelo y condiciones lumínicas.

Resultados: Las propiedades del suelo fueron mejores y el índice de área foliar significativamente mayor en el borde que en el área abierta. La altura, área basal del tallo y el número de rebrotes no difirieron entre sitios ni entre años. Plantas de *C. mexicana* produjeron la mayor cantidad de rebrotes, seguidas de *Q. rugosa*. La supervivencia fue similar en ambas zonas, pero *P. serotina* tuvo la menor supervivencia y *C. mexicana* la mayor.

Conclusiones: *Q. rugosa* y *C. mexicana* se desempeñan mejor que *P. serotina* en ambos sitios, debido a su habilidad para producir rebrotes. El borde de bosque presenta mejores condiciones para introducir especies nativas en la Barranca de Tarango. El pobre desempeño de estas especies se debió a la construcción de una carretera que cruza el área de estudio.

Palabras clave: barranca, bosque de encino, rebrotes, restauración, zona de borde.

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During the second half of the last century urban and population growth of Mexico City converted it into one of the largest cities of the world. Between 1950 and 1980 mean annual growth rate was 4.8 %, mainly due to rural immigration (Ezcurra 1990). In only one decade (1970-1980) more than 3 million people from rural areas established in Mexico City (Calderón & Hernández 1987). The expansion of the city continues nowadays by unregulated growth of new marginal settlements, thus decreasing agricultural and natural areas around it. This, in turn, reduces groundwater recharge and carbon dioxide capture while increasing air pollution and modifying climatic conditions (Maderey-Rascón & Carrillo-Rivera 2005, Awal *et al.* 2010, Ayanlade 2016, Lu *et al.* 2016).

The persistent social pressure upon natural resources involves deforestation of large areas of land causing reductions and fragmentation of forest areas and increases in forest edges (Lovejoy *et al.* 1986, Nascimento *et al.* 2006, Laurance *et al.* 2011). Ravines in Mexico City, particularly those located in the southwestern part of the city, have escaped from the excessive growth of the urban area, due to its topographic characteristics and the difficulties for human settlements. These and the climatic conditions have contributed to maintain vegetation relicts of oak forests, which provide numerous ecological services. However, the lack of coordination and agreement between local authorities and the unplanned urban development have caused serious problems to the ravines. Among these are stream contamination, slope destabilization, imbalance of hydrological functions, loss of biodiversity, and a constant danger to the inhabitants that illegally have occupied the land (Canedo-Castro 2007, SMAGDF 2010).

One of the main goals in restoration programs is the conservation of vegetation by creating buffer zones that minimize temperature and moisture changes (Gascon & Lovejoy 1998), increasing the size of the fragments and reducing the edge effects. Environmental conditions vary among the forest, the edges and the outside. While incidence of light, wind and temperature are larger in the forest edge than inside the forest, moisture content in the air and soil are smaller in the edge than in the forest (Williams-Linera *et al.* 2002). Microclimatic conditions outside the forest are even more severe than in the forest edge, resulting in higher temperature and lower relative humidity of the air, lower soil moisture retention and lower litter decomposition (Saunders *et al.* 1991, Matlack 1993, Alvarez-Aquino *et al.* 2004, Quintana-Ascencio *et al.* 2004). On the other side, tolerance of plants to different micro environmental conditions (sunlight, wind, high and low temperatures and low levels of moisture) might have significant effects on their establishment.

In the face of continued forest loss and degradation, we conducted an experiment aiming to restore a disturbed area of an oak forest in Mexico City, under the principle that ecological restoration is “an intentional activity that initiates or accelerates the recovery of an ecosystem with respect to its health, integrity and sustainability (SER 2002)”. Ecological restoration aims to return the ecosystem that has been degraded, damaged, transformed or destroyed to its historic conditions (SER 2002). Hence, it is important to identify the reference ecosystem that will serve as the model to achieve such activity. Barranca de Tarango is an area originally covered by a temperate oak forest (reference ecosystem) and a xerophytic scrubland (*sensu* Rzedowski 1981), which has passed through several external pressures (García-Romero 2001, Canedo-Castro 2007), causing severe perturbations and needed to be restored. According to previous studies, the oak forest relicts were confined to the northern and southern parts of Barranca de Tarango, whereas induced grasslands, secondary vegetation and isolated trees were found between these forested areas (unpublished data). In order to contribute to connect these areas, improve the environmental conditions of Barranca de Tarango and preserve the original vegetation, we committed ourselves in assisting in the recovery of this forest by employing one of the most widely used ecological restoration practices consisting on planting tree native species (Harrington 1999, SER 2002) in a forest edge and an open area. It is known that forest edges are suitable places for the restoration of disturbed ecosystems (Bonfil & Soberón 1999, Asbjornsen *et al.* 2004, López-Barrera & Newton 2005, López-Barrera *et al.* 2006, Bueno & Llambí 2015), but, is the open area equally appropriate for restoration as is the forest edge? To answer this question, we established experimental plots in a forest edge and an open area adjacent to it in order to introduce three native species. The aims of this study were: i) to characterize micro environmental conditions of the forest edge and open area by means of soil and light variables during the dry and rainy seasons, and ii) to estimate whether

Author Contributions

Elizabeth Chávez-García conceived, designed and performed the experiments, analyzed the data and wrote the paper.

Ana Mendoza conceived, designed and directed the research and reviewed drafts of the paper.

the performance (survival and growth) of these tree species differs between sites, on account of their differences in microclimatic conditions. Our hypothesis is that plants of *Quercus rugosa* Née (Fagaceae) will perform better in the forest edge than in the open area, since this is a shade tolerant species (Vázquez-Yanes *et al.* 1999). Contrary, plants of *Crataegus mexicana* Moc. Sessé (Rosaceae) and *Prunus serotina* subsp. *capuli* (Cav.) McVaugh (Rosaceae) will respond more favorably to the open area conditions because of their intolerance to shade conditions (Auclair 1975, Vázquez-Yanes *et al.* 1999). Results of this research will provide information on some actions to be taken for restoring urban forests of one of the world's largest cities that even now continues to be threatened by urban development.

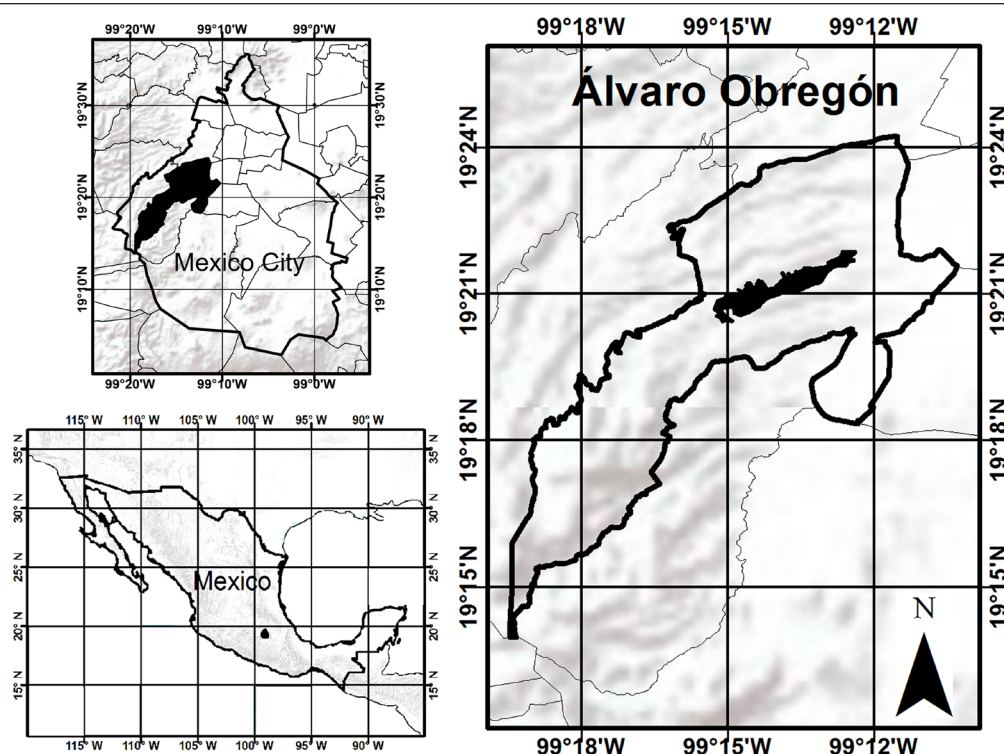
Material and methods

Study site. The study was carried out in Barranca de Tarango (Figure 1), located in Delegación Álvaro Obregón, Mexico City ($19^{\circ} 14'N - 19^{\circ} 25'N$ and $99^{\circ} 10'W - 99^{\circ} 20'W$), at an altitude between 2,330 and 2,580 m (GODF 2009, Guerra-Martínez 2012). It occupies an area of approximately 280 ha with a length of 6 km. The oak forest in the area comprises two associations: *Quercus castanea* - *Quercus obtusata* with tree density of 448 individuals ha^{-1} , and *Quercus castanea* - *Quercus rugosa* with tree density of 700 individuals ha^{-1} (Guerra-Martínez 2012).

Study species. The study species were *Prunus serotina* subsp. *capuli*, *Quercus rugosa*, and *Crataegus mexicana*; from now on the first species will be referred to as *Prunus serotina*. These tree species were chosen according to the following criteria: a) they are native of remnant oak forests of the city, b) they were found in previous studies carried out in the Barranca de Tarango (Guerra-Martínez 2012), and c) they are dispersed by small mammals and birds (Vázquez-Yanes *et al.* 1999, GDF 2001, Rodríguez-Sánchez & Cohen-Fernández 2003). Furthermore, these species are known for their tolerance to stressful conditions such as drought or pollution, and have been recommended for restoring degraded areas (Vázquez-Yanes *et al.* 1999, Martínez-González 2008).

Fieldwork. With the aid of satellite images and field surveys, in June 2010 we selected two experimental sites in Barranca de Tarango, one in the forest edge and the other in an open area.

Figure 1. Study site located in Barranca de Tarango, Delegación Álvaro Obregón, Mexico City, Mexico.



The forest edge was considered as the strip of land of approximately 50 m, extending from the interior of the forest to the area totally devoid of vegetation, while the open area was that outside the forest at nearly 150-200 m far from the edge with few or no trees at all. In each site two plots were established. Each plot in the open area was of approximately 1,800 m², located at 2,538 - 2,629 m asl, while plots in the forest edge were approximately 1,200 m² each, located at 2,547 - 2,553 m asl.

Site characterization. Environmental conditions such as air temperature and humidity, soil properties and solar radiation were used to characterize both sites. Air temperature and humidity were registered using sensors (HOBO Pro v2) placed in each plot. These variables were registered every two hours from August 2010 to July 2012. However, no data were collected in the open area during the last six months (February to July 2012) because sensors were stolen. Six sensors were placed randomly in the study area, three in the forest edge and three in the open area. The collected data were analyzed using monthly averages.

Soil properties in both sites were described by means of two soil profiles according to Siebe *et al.* (2006). Profiles were 40 cm in depth in the open area and 64 cm in the forest edge. Soils were classified according to the World Reference Base for Soil Resources (FAO-ISRIC-IUSS 2007). Parameters used to describe the soil in the field were texture, organic matter, bulk density, pH, aggregate stability, aeration capacity, available water, and field capacity.

To describe the light environment of each site, we used hemispherical photographs, which have been used to estimate solar radiation penetration and canopy architecture (Chazdon & Field 1987, Becker *et al.* 1989, Rich 1990, Chianucci & Cutini 2012). Eleven photographs were taken in each area, always looking upward with the camera aligned to magnetic north and placed at a height between 57 and 67 cm from the soil surface. We used a Coolpix 990 digital camera (Nikon) set on stabilizing equipment designed by WinScanopy. Photographs were taken in each plot at a distance of 14 m between them during the dry (February 2011) and the rainy seasons (August 2011). Photographs were processed using Gap Light Analyzer (GLA) Version 2.0 (Frazer *et al.* 1999). The following parameters were considered for the analyses: i) canopy openness, percentage of open sky seen from beneath the forest canopy; ii) leaf area index, total leaf area per unit of land; iii) direct and diffuse radiation, amount of direct and diffuse sunlight, and iv) the total transmitted radiation, the sum of direct and diffuse radiations (Frazer *et al.* 1999, Régent Instruments 2002).

Plant performance (growth and survival). Plants used in this study were purchased in a nursery located in Mexico City. Plants of *Quercus rugosa* and *Crataegus mexicana* were 21 months old, whereas plants of *Prunus serotina* were 17 months old at the time of planting. We planted a total of 480 plants, 120 were planted in each of the two plots in the open area (240) and the same number in the two plots of the forest edge. The number of plants of each species varied according to the densities obtained by Guerra-Martínez (2012) in a previous study in Barranca de Tarango. The number of plants was 60 per plot of *Q. rugosa* and 30 plants per plot of each species of *C. mexicana* and *P. serotina*. Individuals of the three species were planted randomly in July 2010 at a distance of approximately 3 m between plants.

Before being planted, each plant was numbered and the height and diameter at the base of the stem were recorded. As some plants produced stems by clonal growth, the height of the main stem and the diameter of each individual stem also were recorded. Stems produced by clonal growth will be referred to as sprouts. Height was measured from the base of the stem to the apical bud. Stem diameter was measured at the base of it with a digital calliper. A permanent mark was placed at the base of the stem as a reference for measuring height and stem diameter during the following censuses (every 4 months). Growth of plants was assessed as the increase in height and basal area of the stem, as well as the number of sprouts. To estimate the basal area (BA) we used the following formula:

$$BA = \pi (d/2)^2 \quad (1)$$

where π is the constant 3.1416, and d is the stem diameter.

From the date of planting, plants were registered monthly during two years. In order to estimate plant survival, in each census we recorded plant mortality and the possible causes of death.

The following criteria were used to resolve whether a plant was dead: 1) the plant was not found in two consecutive records (disappearance), 2) the plant was uprooted, and 3) the plant was found completely dry, without foliage or sprouts (desiccation).

Data analysis. To prove whether monthly air temperature and relative humidity differed between sites (open area and forest edge), we used one-way ANOVA tests. Differences in canopy openness, leaf area index, direct, diffuse and total radiation between sites and seasons (dry and rainy) were analyzed by fitting generalized linear models (GLM) (McCullagh & Nelder 1989). For the first variable (canopy openness), we used a binomial link distribution, whereas for the other four variables we used a Poisson link distribution. The null hypothesis considered that these variables did not differ between sites or seasons.

In order to test whether stem height and basal area of plants differed between sites each year, we used ANCOVA's tests. Stem height and basal area of plants of each species at the moment of planting (2010) were used as covariables to analyze height and basal area (response variables) in 2011. The same criterion was used one year later; height and basal area in 2012 were the response variables in the analyses, whereas stem height and basal area in 2011 were used as covariables. To compare plant size according to the number of sprouts of each species, in each site during 2011 and 2012, we used Kruskal-Wallis tests. Finally, differences in survival of plants of each species in each site were analyzed using a binomial logistic regression, with site as independent variable and death or survival of individuals as the binary response variable (Crawley 2005). The null hypothesis was that survivorship of plants is not affected by site. Analyses were carried out at a 95 % level of significance, using STATISTICA (StatSoft, Inc. 2004), whereas light variables and survival analyses were performed with R (R Development Core Team 2014).

Results

Site characterization. The lowest average temperatures were registered in January 2012 in the open area (8.6 °C) and in December 2010 in the forest edge (9.8 °C). The highest average temperatures were recorded in May 2011; 20 °C was registered in the open area and 18.6 °C in the forest edge. However, temperature did not differ between sites ($F = 0.059$, $p = 0.8$). The lowest average relative humidity was registered in March 2011, being higher in the forest edge (42.7 %) than in the open area (39 %); the highest average relative humidity was recorded in July 2011; 88.6 % in the open area and 91.1 % in the forest edge. However, no significant differences were found between sites ($F = 0.23$, $p = 0.64$).

Soils were developed on pyroclastic materials in which the processes of formation have led to Regosols (open area) and Phaeozems (forest edge). Regosols are characterized by weakly developed soils in unconsolidated materials without diagnostic horizons and can have a low water

Table 1. Soil characteristics in the open area and in the forest edge in Barranca de Tarango, Mexico City.

	Open area	Forest edge
Soil type	Haplic Regosol (tephric)	Haplic Phaeozem (tephric)
Depth	40 cm	64 cm
Texture	Fine silt loam and clay loam	Silty clay loam
Organic matter	1.8 %	4 %
Bulk density	Medium (1.4 -1.5 g/cm ³)	Low-medium (1.0 -1.38 g/cm ³)
pH	6.5 - 7	6
Rock fragments in the soil	Low (< 2 %)	Low (< 2 %)
Root density	Medium at top of soil (6-10 fine roots per dm ²)	Medium at top of soil (6 -10 fine roots per dm ²)
Aggregate stability	Moderate	High
Aeration capacity	Low to medium (4 - 7.5 %)	Medium (8 %)
Available water	Low (55.8 L/m ²)	Medium (125 L/m ²)
Field capacity	Low (148 L/m ²)	Medium (275 L/m ²)

Table 2. Light variables in the open area and the forest edge in Barranca de Tarango, Mexico City during the dry and rainy seasons. Values are means \pm 1SD.

Light variables	Open area		Forest edge	
	Dry season	Rainy season	Dry season	Rainy season
Canopy openness (%)	72.3 \pm 10.4	66.3 \pm 6.6	51.5 \pm 16.9	48.4 \pm 12.5
Leaf area index (m ² m ⁻²)	0.2 \pm 0.2	0.2 \pm 0.1	0.6 \pm 0.4	0.6 \pm 0.3
Direct radiation (mol m ⁻² d ⁻¹)	16.9 \pm 1.3	20.3 \pm 1.2	11.1 \pm 5.3	15.2 \pm 3.4
Diffuse radiation (mol m ⁻² d ⁻¹)	15.5 \pm 1.7	18.2 \pm 1.0	11.5 \pm 3.6	13.0 \pm 2.8
Total transmitted radiation (mol m ⁻² d ⁻¹)	32.4 \pm 2.9	38.6 \pm 2.1	22.6 \pm 8.8	28.2 \pm 6.1

retention capacity. Phaeozems also developed in unconsolidated materials, including dark soils rich in organic matter (FAO-ISRIC-IUSS 2007, Table 1). Soil textures showed a homogeneous mixture of the three primary particles (clay, silt and sand). Stoniness was low (less than 2 %) and root density was medium in the first centimeters of the soil in both sites. Bulk density was slightly higher in the open area than in the forest edge. Soil depth, organic matter content, aggregate stability, aeration capacity, amount of available water and field capacity were higher in the forest edge than in the open area (Table 1).

Canopy openness, direct, diffuse and total radiation were higher in the open area than in the forest edge, whereas leaf area index was the other way around, lower in the open area than in the forest edge during the dry and rainy seasons (Table 2). Values of all light variables were smaller in the dry than in the rainy season, except canopy openness, whose value was larger in the dry than in the rainy season. Statistical analyses showed that all light variables differed significantly between sites, while direct, diffuse and total radiation also differed between seasons (Table 3).

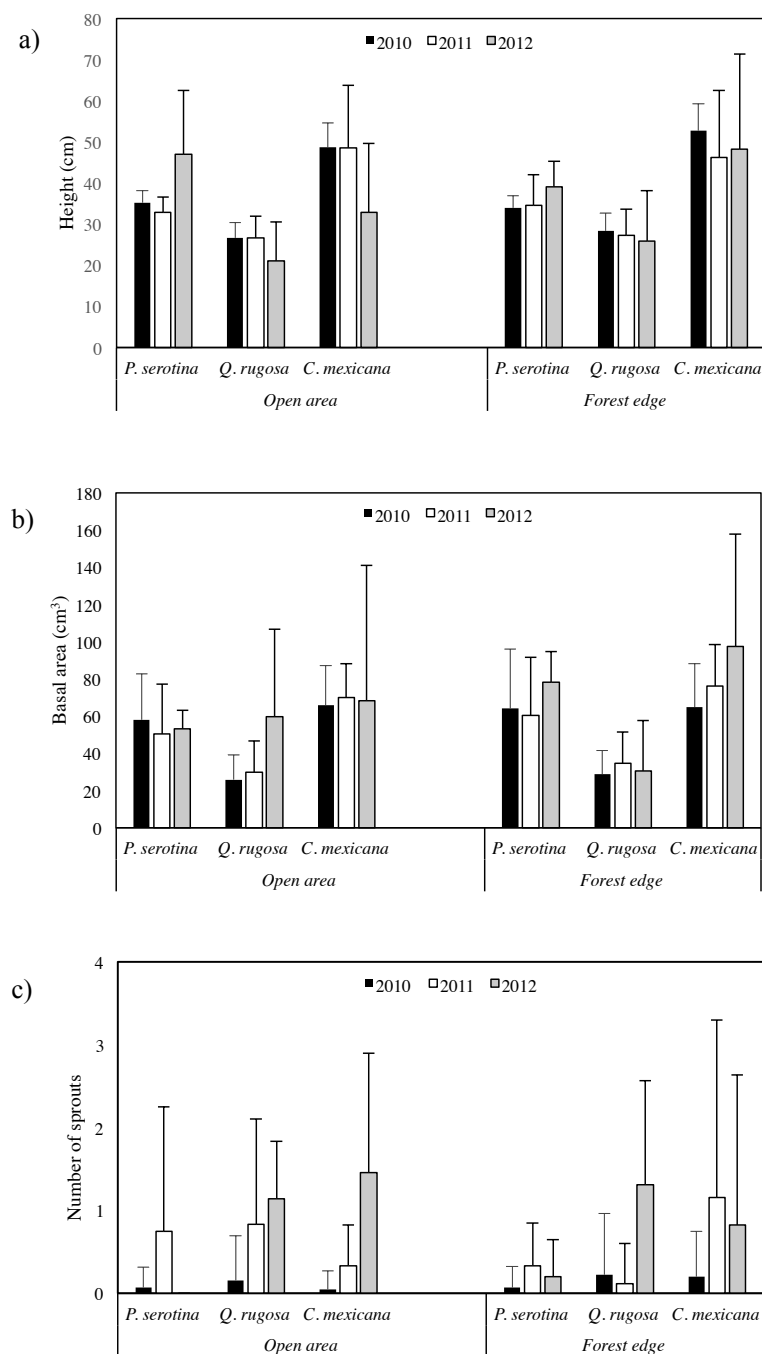
Table 3. Generalized linear model (GLM) results of the light variables measured in the open area and the forest edge in Barranca de Tarango, Mexico City. Estimate values \pm SE.

Parameter		Estimate	t value	p
Canopy openness	(Intercept)	1.8536 \pm 0.3815	4.859	< 0.001
	site	-0.7100 \pm 0.1710	-4.152	< 0.001
	season	-0.2468 \pm 0.1706	-1.446	0.1556
Leaf area index	(Intercept)	1.0918 \pm 0.6096	1.791	0.0811
	site	1.4196 \pm 0.2821	5.033	< 0.001
	season	0.1187 \pm 0.1898	0.625	0.5353
Direct radiation	(Intercept)	7.5844 \pm 0.0605	125.191	< 0.001
	site	-0.2934 \pm 0.0763	-3.846	< 0.001
	season	-0.1851 \pm 0.0758	-2.443	0.0190
Diffuse radiation	(Intercept)	7.5002 \pm 0.1300	57.672	< 0.001
	site	-0.2711 \pm 0.0606	-4.474	< 0.001
	season	0.1193 \pm 0.0601	1.983	0.0541
Total transmitted radiation	(Intercept)	8.2003 \pm 0.1401	58.520	< 0.001
	site	-0.2922 \pm 0.0651	-4.483	< 0.001
	season	0.1634 \pm 0.0647	2.525	0.0155

Growth. Since the beginning of the experiment (2010), the largest plants in height and basal area were those of *Crataegus mexicana*, followed by *Prunus serotina*, and those of *Quercus rugosa* were the smallest (Figure 2a, b; Appendix 1). The first year (2011), in the open area the three species increased in the number of sprouts (Figure 2c; Appendix 1). Additionally, *Q. rugosa* and *C. mexicana* also increased in basal area. One year later (2012), *Q. rugosa* again increased in basal area and in the number of sprouts, whereas plants of *P. serotina* increased in height and basal area, and those of *C. mexicana* only increased in basal area.

In the forest edge *Prunus serotina* was the only species that increased in height in 2011; this species also increased in the number of sprouts as did *Crataegus mexicana*, and in *Quercus ru-*

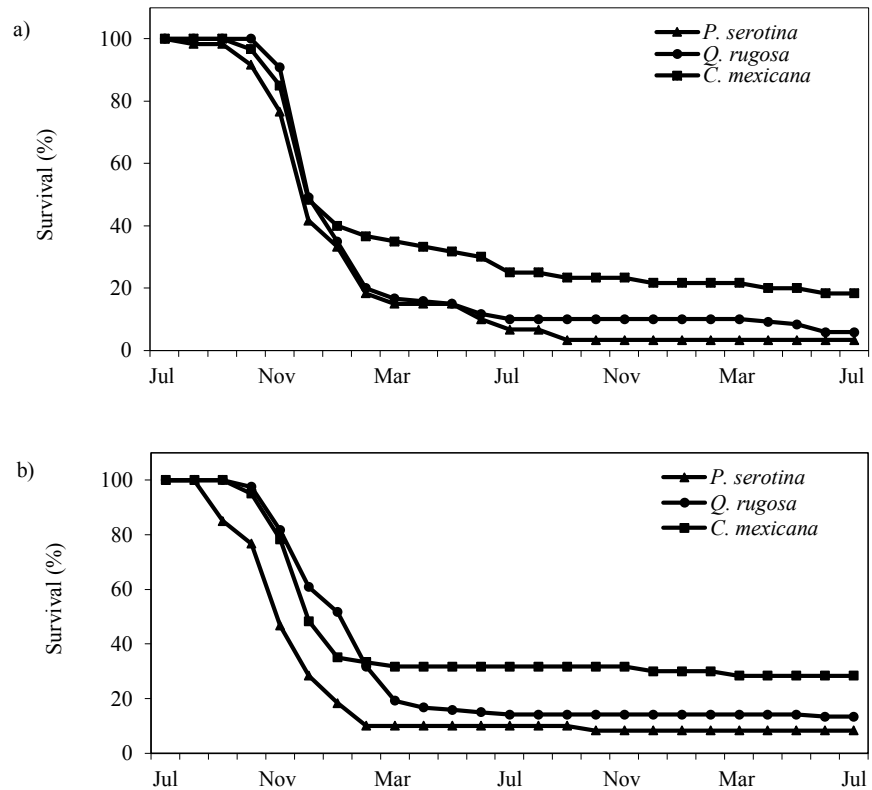
Figure 2. Growth in a) height, b) basal area, and c) number of sprouts of plants of *P. serotina*, *Q. rugosa* and *C. mexicana* planted in the open area and the forest edge in Barranca de Tarango, Mexico City, one (2011) and two (2012) years after being planted (2010). Error bars represent +1 standard deviation.



Q. rugosa the increment was in basal area (Figure 2 a, b, c; Appendix 1). In 2012, *P. serotina* and *C. mexicana* increased in height and basal area and *Q. rugosa* in number of sprouts. As plants of *P. serotina* experienced high mortality during the course of the experiment, no statistical analyses were carried out with this species due to the small sample size. However, results of the statistical analyses for the other two species did not show significant differences in height between the open area and the forest edge (*Q. rugosa* $F = 0.37$, $p = 0.551$; *C. mexicana* $F = 1.95$, $p = 0.175$), in basal area (*Q. rugosa* $F = 2.45$, $p = 0.133$; *C. mexicana* $F = 1.83$, $p = 0.187$) and number of sprouts (*Q. rugosa* $F = 0.39$, $p = 0.75$; *C. mexicana* $F = 100.45$, $p = 0.07$).

As no significant differences were found in height, basal area and number of sprouts between the open area and the forest edge, data from both sites were grouped together for each species in order to analyze whether these variables differed between years (Appendix 2). The only species

Figure 3. Survival of plants of *P. serotina*, *Q. rugosa* and *C. mexicana* planted in: a) the open area, and b) the forest edge in Barranca de Tarango, Mexico City from July 2010 date of planting) to July 2012 (end of the study).



that increased in height was *Prunus serotina* and this only occurred in 2012; it also increased in basal area and number of sprouts in 2012 and in 2011, respectively. *Quercus rugosa* and *Crataegus mexicana* decreased in height in both years, but always increased in basal area and number of sprouts relative to the year before (2010 < 2011 < 2012). However, no significant differences were found between years either in height (*Q. rugosa* $F = 2.72$, $p = 0.114$; *C. mexicana* $F = 2.10$, $p = 0.158$), basal area (*Q. rugosa* $F = 2.53$, $p = 0.126$; *C. mexicana* $F = 0.42$, $p = 0.518$) or number of sprouts (*Q. rugosa* $\chi^2 = 3.43$, $p = 0.180$; *C. mexicana* $\chi^2 = 3.71$, $p = 0.156$).

Survival. Plant survival of the three species was similar in both sites during the course of the study (Figure 3). During the first three months, plants in both sites showed high survival followed by an abrupt decrease. Mortality of plants of *Prunus serotina* occurred earlier in the forest edge. From January 2011, onwards, mortality of plants of the three species remained relatively constant both in the open area and the forest edge. No significant differences were found in plant survival of these species between sites (estimate = 0.727 ± 0.516 , t value = 1.410, $p = 0.189$). At the end of the study (Figure 3), plants of *Crataegus mexicana* had the highest survival (18.3 % in the open area and 28.3 % in the forest edge; 23.3 % for both sites), and *P. serotina* had the lowest (3.3 % in the open area and 8.3 % in the forest edge; 5.8 % for both sites). Plants of *Quercus rugosa* had an intermediate survival in both sites (5.8 % in the open area and 13.3 % in the forest edge; 9.5 % for both sites). When the experiment finished, 8.3 % of plants of the three species survived in the open area, while 15.8 % survived in the forest edge.

Discussion

Since *Quercus rugosa* is a shade tolerant species and *Crataegus mexicana* and *Prunus serotina* are both intolerant to the same condition, we expected that *Q. rugosa* would have a better performance in the forest edge, while the other two would have it better in the open area. However, we did not observe a clear pattern of growth of each species in either site; rather, it seems that each had its own mechanism for growing. In the open area plants of the three species grew in the number of sprouts in 2011, *Q. rugosa* and *C. mexicana* also increased in basal area and

none of the three species increased in height. In 2012, plants of *Q. rugosa* again increased in the number of sprouts and in basal area, *P. serotina* in height and basal area, but *C. mexicana* only in the number of sprouts. Contrary, in the forest edge plants of *Q. rugosa* increased in basal area in 2011 and in the number of sprouts in 2012. Nonetheless, in this site, plants of *P. serotina* increased in height in both years, in the number of sprouts in 2011 and in basal area in 2012, while those of *C. mexicana* increased in basal area in both years and additionally increased in the number of sprouts in 2011 and in height in 2012.

These results suggest that patterns of resource allocation of these species are not fixed; rather they can be modified to modulate growth, possibly due to the amount of resources stored before and after they were planted and/or the resources available in each site (*i.e.*, light incidence or soil quality). It has been suggested that the amount of reserves in a developing seedling is crucial for enhancing the probability of establishment under hazardous environments (Westoby *et al.* 1996, Muller *et al.* 2000, Walters & Reich 2000). Factors such as phenotypic plasticity induced by the environment (Whitman & Agrawal 2009), or the developmental ontogeny of each species (Vesk 2006) might also explain the growing strategy followed by each species.

Evidently, height was not the best strategy for growing of plants of *Quercus rugosa* that never increased, but rather decreased or almost remained the same. *Prunus serotina* was the only species that always increased in height, except in the open area in 2011, but plants of *Crataegus mexicana* only increased in the forest edge in 2012. Decreases in height of the main stem of plants of the three species were mainly due to stem breakage. Therefore, height is not a reliable variable to measure growth of these species. Increases in basal area and number of sprouts were better strategies of growing for these species. It is likely that the increases observed in basal area were mainly due to the number of new sprouts, whereas decreases could be related to volume loss associated with water deficit (Cabrera-García *et al.* 1998). Seedlings of other oak species like *Q. rugosa*, *Q. glaucoides* M. Martens & Galeotti and *Q. candicans* Née also experienced biomass losses when planted in the forest or in a degraded grassland (Cabrera-García *et al.* 1998, Bonfil & Soberón 1999, Rubio-Licona 2009, Gómez-Cirilo 2010), due to either the death of the main stem that was broken before, or to the dry conditions in the sites where they were planted.

Our results show that clonal growth was the best strategy of growing for these species. It has been suggested that this is an important mechanism of plant regeneration and propagation that allows plant persistence in local and disturbance-prone environments (Espelta *et al.* 1999, Bond & Midgley 2001, Chong *et al.* 2007). We found that plants of the three species in both sites had the ability to produce sprouts from basal meristems. However, it varied among species, and between years and sites, although the differences were not significant. Increases in the number of sprouts of *Quercus rugosa* and *Crataegus mexicana* occurred each year in the open area, whereas in the forest edge only happened in 2011 in *C. mexicana* and in 2012 in *Q. rugosa*. Similar results were reported by López-Barrera *et al.* (2006) who found that seedlings of several species of *Quercus* planted in open areas produced more sprouts than those along the forest edge. In the case of *Prunus serotina*, Auclair (1975) pointed out that clonal growth in plants of this species depends on forest structure.

It seems that the ability to produce sprouts is an important strategy for the survival of plants of these species, because if the main stem dies, the plant can produce new shoots, thus buffering severe disturbances. Our results suggest that clonal growth of these species is not a mechanism to propagate as in other species, but rather is a mechanism of growing at early stages (Mendoza & Franco 1998). In addition, this mechanism reduced mortality risks of the whole plant, enhancing their survival in the open area and forest edge. Therefore, clonal growth is likely to confer advantages (Puijalon & Bornette 2006) to plants of these species that were exposed to damage (see below).

Producing sprouts is not only a mechanism to grow at an individual level, but it is also valuable at the population level. Minnich *et al.* (1994) and Alfonso-Corrado *et al.* (2007) suggested that populations of *Quercus eduardii* Trel. and *Q. potosina* Trel. have survived after severe logging due to their ability to produce sprouts. Additionally, a demographic study carried out by Alfonso-Corrado *et al.* (2007) showed that clonal growth was more important to the population growth rate than seed production, suggesting that the former is crucial for the regeneration of

these species in Sierra Fría, Aguascalientes, Mexico. Thus, for restoration purposes it is convenient to use species that produce sprouts to rehabilitate degraded areas.

It is likely that plants faced some transplanting stress while undergoing a period of adjustment to the new environment (Rietveld 1989). Of the three species, *Prunus serotina* was more affected than the other two species, although it was pointed out as a species tolerant to drought, or shallow and stony soils (Vázquez-Yanes *et al.* 1999), and common in degraded oak forests (García-Romero 2001). Plants of this species showed signs of stress (more leaf losses and changes in foliage color) to the conditions of the experimental sites since the very beginning. This effect also has been observed in species like *Quercus pyrenaica* Willd. and *Pinus halepensis* Mill. (Castro *et al.* 2006, Ariza-Mateos *et al.* 2008). The difficulty of plants of *P. serotina* to cope with the environment is clearly shown in the survival curves of this species in both sites. Even if the three species showed similar survival curves, plants of *P. serotina* experienced higher mortality, especially in the forest edge, where they started to die soon after being planted. We have no explanation whatsoever of its low performance in the study area, particularly in the forest edge, but it is likely that its intolerance to shadow or the effect of some herbivore or pathogen (Auclair 1975, Vázquez-Yanes *et al.* 1999, Hartman & McCarthy 2004) might have contributed to this result.

Plants survived better in the forest edge than in the open area, although there were no statistical differences between the experimental sites. At the end of the study, plants of *Crataegus mexicana* had the highest survival in both sites and *Quercus rugosa* an intermediate. The study sites were rather homogenous in terms of temperature and relative humidity; however, soil and light characteristics were more favorable in the forest edge than the open area (Tables 1, 2). These differences might explain the better performance of plants in the forest edge.

Survival of plants of the three species contrasts with the results of Hernández-García (2011) and Corona-Álvarez (2011), who found that survival of plants of *Quercus rugosa* (12 and 36 months old), was above 90 %, and that of plants of *Quercus mexicana* Bonpl. and *Q. rugosa* (36 months old) was above 80 %, respectively. Plants of these species also were planted in Barranca de Tarango, Mexico City (the same study site). However, twelve months old seedlings of *Quercus rugosa* planted in a degraded forest in Mexico City experienced high mortality (90 %) one year after planted (Cabrera-García *et al.* 1998) similar to what we found in this study. Our results also contrast with those reported by Löf *et al.* (2004), Schreeg *et al.* (2005) and Skousen *et al.* (2009) who found that survival of plants of *Prunus serotina* and *P. avium* (L.) L. was high (> 80 %), one and three years after being planted, and between 50 and 80 % during the second year in a reclaimed surface mine (Skousen *et al.* 2009). Similarly, in northwest Florida, 24 months old plants of *Crataegus marshallii* Eggl. survived very well (> 90 %) either in shade or full sun, with supplemental irrigation during the summer months (Fletcher *et al.* 2012) and those of *C. monogyna* Jacq. had more than 80 % survival three years after planted in a farmland in Denmark (Löf *et al.* 2004).

The micro environmental conditions each plant encountered in its new habitat and the amount of stored resources influenced survival and growth of these species, as is shown by the survival curves and the large standard deviations observed for each growth variable, particularly during 2012. This suggests that each plant will allocate more resources to growing if microsite conditions where they were planted are appropriate (light, soil, humidity), whereas those planted in sites with poor conditions will invest more resources in establishing and surviving. It is also important to consider the quality of plants to be introduced. Ramírez-Contreras & Rodríguez-Trejo (2009) pointed out that survival of established seedlings of *Quercus rugosa* was more than 37 % when plants were taller and with larger stem diameters on account of their ability to compete with herbaceous plants. Monitoring the experimental plants is critical to determining whether these will follow certain patterns of growth over time. Nonetheless, a two-year period recording plant performance might not be enough to conclude about such patterns.

We presume that one of the main reasons for the low performance of these species is that while we started this experiment, the mayor of Mexico City authorized the construction of an expressway that crosses Barranca de Tarango, classified as “valued environmental area” (GODF 2009), meaning that it should be protected for the conservation of the biodiversity of the area (GODF 2010). Although authorities and builders assured that our experiment was going to be monitored and untouched, this did not occur. Lorries, machinery and workers moving through-

out the area provoked side effects upon our experiment. We do not doubt that under other circumstances plants would have performed better as was shown by Hernández-García (2011) and Corona-Álvarez (2011) in Barranca de Tarango and by Löf *et al.* (2004), Schreeg *et al.* (2005), Skousen *et al.* (2009) and Fletcher *et al.* (2012) in other places.

Quercus rugosa, a shade tolerant species and *Crataegus mexicana*, an intolerant one, responded similarly in both sites. These species had the highest survival, probably because of their ability to produce more sprouts, strategy that allows these plants to reduce mortality risks. Therefore, both species are recommended to be planted in either site in Barranca de Tarango. Nevertheless, the forest edge seems to be better than the open area for introducing native species, mainly because of the better environmental conditions, but also for its greater distance to the construction of the highway. This study gives cues on certain mechanisms followed by these species after being planted. Also, it makes an important contribution to improve the selection of species to be introduced, as well as the area where they have to be planted in order to increase the probability of survival and the likelihood to persist in a new habitat.

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Appendix 1. Height, basal area and number of sprouts of plants of *Prunus serotina*, *Quercus rugosa* and *Crataegus mexicana* in 2011 and 2012, planted in 2010 in an open area and in the forest age in Barranca de Tarango, Mexico City. Values are means \pm 1SD.

			2010	2011	2012
Height (cm)	Open area	<i>P. serotina</i>	35.21 \pm 2.9	32.90 \pm 3.8	47.0 \pm 15.6
		<i>Q. rugosa</i>	26.63 \pm 3.8	26.63 \pm 5.4	21.0 \pm 9.5
		<i>C. mexicana</i>	48.6 \pm 6.0	48.5 \pm 15.2	32.8 \pm 16.8
	Forest edge	<i>P. serotina</i>	33.9 \pm 3.0	34.6 \pm 7.4	39.0 \pm 6.3
		<i>Q. rugosa</i>	28.3 \pm 4.4	27.2 \pm 6.5	25.8 \pm 12.4
		<i>C. mexicana</i>	52.7 \pm 6.5	46.2 \pm 16.4	48.3 \pm 23.1
Basal area (cm ²)	Open area	<i>P. serotina</i>	58.0 \pm 24.8	50.4 \pm 26.7	53.1 \pm 10.2
		<i>Q. rugosa</i>	25.9 \pm 13.3	29.9 \pm 16.8	59.7 \pm 47.0
		<i>C. mexicana</i>	65.9 \pm 21.4	69.9 \pm 18.4	68.3 \pm 72.8
	Forest edge	<i>P. serotina</i>	64.1 \pm 31.8	60.5 \pm 31.0	78.3 \pm 16.3
		<i>Q. rugosa</i>	28.7 \pm 12.8	34.6 \pm 16.9	30.5 \pm 27.1
		<i>C. mexicana</i>	64.8 \pm 23.2	76.0 \pm 22.6	97.3 \pm 60.6
Number of sprouts	Open area	<i>P. serotina</i>	0.07 \pm 0.25	0.75 \pm 1.50	
		<i>Q. rugosa</i>	0.16 \pm 0.53	0.83 \pm 1.27	1.14 \pm 0.69
		<i>C. mexicana</i>	0.05 \pm 0.22	0.33 \pm 0.49	1.45 \pm 1.44
	Forest edge	<i>P. serotina</i>	0.07 \pm 0.25	0.33 \pm 0.52	0.20 \pm 0.45
		<i>Q. rugosa</i>	0.22 \pm 0.75	0.12 \pm 0.49	1.31 \pm 1.25
		<i>C. mexicana</i>	0.20 \pm 0.55	1.16 \pm 2.14	0.82 \pm 1.81

Appendix 2. Height, basal area and number of sprouts of plants of *Prunus serotina*, *Quercus rugosa* and *Crataegus mexicana* in 2011 and 2012. Plants were introduced in 2010 in Barranca de Tarango, Mexico City. Values are means \pm 1SD.

	2010	2011	2012
Height (cm)			
<i>P. serotina</i>	34.6 \pm 3.0	33.9 \pm 6.0	41.3 \pm 9.1
<i>Q. rugosa</i>	27.5 \pm 4.2	27.0 \pm 6.0	24.3 \pm 11.6
<i>C. mexicana</i>	50.7 \pm 6.5	47.2 \pm 15.6	42.2 \pm 21.9
Basal area (cm ²)			
<i>P. serotina</i>	61.0 \pm 28.6	56.4 \pm 28.3	71.1 \pm 18.6
<i>Q. rugosa</i>	27.3 \pm 13.1	32.7 \pm 16.8	39.4 \pm 35.9
<i>C. mexicana</i>	65.5 \pm 22.2	73.3 \pm 20.8	85.9 \pm 65.9
Number of sprouts			
<i>P. serotina</i>	0.1 \pm 0.3	0.5 \pm 1.0	0.1 \pm 0.4
<i>Q. rugosa</i>	0.2 \pm 0.6	0.4 \pm 0.9	1.3 \pm 1.1
<i>C. mexicana</i>	0.1 \pm 0.4	0.8 \pm 1.7	1.1 \pm 1.7