



Performance of 15 tropical tree species recruited or transplanted on restoration settings

ALEJANDRA GUZMÁN-LUNA¹, CRISTINA MARTÍNEZ-GARZA^{2*}

Botanical Sciences
94 (4): 757-773, 2016

DOI: 10.17129/botsci.659

Abstract

Background. Minimal restoration intervention includes actions to stop disturbance so natural succession may take place whereas maximal intervention involves the establishment of plantings.

Questions. To evaluate the success of minimal versus maximal restoration intervention, the performance of recruits and transplants was assessed. To this end, performance of 15 native tree species was predicted using life-history, their origin (recruits or transplants) and 12 plant functional traits.

Study site and years of study. This study was carried out in pastures at Los Tuxtlas, Veracruz, Mexico in 5 years old restoration settings.

Methods. Pioneer and non-pioneer species were planted in 16 30 × 30 m plots whereas natural recruitment was evaluated in plantings and at eight additional fenced plots.

Results. Overall 15 species recruited or planted, pioneers had higher performance than non-pioneer. Transplant shock in terms of survival and height growth rates was overcome after 5 years probably as a result of increases in diameter growth rates.

Conclusions. Tree species are divided in three groups to give recommendation for restoration: (1) Species in the *Good recruiters* group do not need to be transplanted; if seed sources are not close, we recommend direct seeding (i.e., *Albizia purpusii*, *Cedrela odorata*, *Cecropia obtusifolia*). (2) Species in the *Good transplants* group show very low or nil recruitment; they should be transplanted (i.e., *Ochroma pyramidalis*, *Ficus yoponensis*, *Cojoba arborea*). (3) Species in the *Poor transplants* group should be transplanted but once a canopy has developed (i.e., *Amphitecna tuxtlanensis*, *Brosimum alicastrum*, *Bernoullia flammea*).

Key words: functional traits, *Helicarpus appendiculatus*, life history, natural recruitment, transplant shock

Desempeño de 15 especies de árboles tropicales reclutados o trasplantados en áreas de restauración

Resumen

Antecedentes. La mínima intervención de restauración incluye detener la perturbación para que la sucesión se lleve a cabo mientras que la máxima involucra el establecimiento de plantaciones.

Preguntas. Para evaluar el éxito de la intervención mínima en comparación con la máxima se midió el desempeño de reclutas y trasplantes. Con este fin se buscó explicar el desempeño de 15 especies de árboles nativos mediante su historia de vida, su origen (reclutas o trasplantes) y 12 caracteres funcionales.

Sitio de estudio y fechas. Este estudio se llevó a cabo en Los Tuxtlas, México dentro de un proyecto de restauración de 5 años.

Métodos. Árboles pioneros y no-pioneros fueron plantados en 16 parcelas de 30 × 30 m; el reclutamiento natural se evaluó en las plantaciones y en ocho parcelas adicionales.

Resultados. Incluyendo las 15 especies evaluadas, las pioneras tuvieron mejor desempeño que las no-pioneras. El stress del trasplante terminó después de 5 años, probablemente debido al incremento en diámetro.

Conclusiones. Las especies han sido divididas en tres grupos para dar recomendaciones sobre su uso en restauración: (1) Las especies del grupo *Buenos reclutas* no necesitan ser plantadas; si no hay fuentes de semillas cercanas se recomienda introducirlas por siembra directa (i.e., *Albizia purpusii*, *Cedrela odorata*, *Cecropia obtusifolia*). (2) Las especies del grupo *Buenos trasplantes* mostraron bajo o nulo reclutamiento, estas especies deben ser trasplantadas (i.e., *Ochroma pyramidalis*, *Ficus yoponensis*, *Cojoba arborea*). (3) Las especies en el grupo *Malos trasplantes* pueden ser plantadas una vez que exista un dosel (i.e., *Amphitecna tuxtlanensis*, *Brosimum alicastrum*, *Bernoullia flammea*).

Palabras clave: *Helicarpus appendiculatus*, historia de vida, reclutamiento natural, stress debido al trasplante, caracteres funcionales

¹ Posgrado en Ciencias Biológicas, Universidad Nacional Autónoma de México, Mexico City, Mexico,

² Centro de Investigación en Biodiversidad y Conservación, Universidad Autónoma del Estado de Morelos, Cuernavaca, Mexico

*Corresponding author: cristina.martinez@uaem.mx

Restoration actions may include different degrees of intervention to recover the structure and function of degraded tropical rain forest. Minimal intervention includes actions to stop disturbance so natural succession may take place whereas maximal intervention involves the establishment of restoration plantings (SER 2004, Morrison & Lindell 2011, Martínez-Garza *et al.* 2016). Many studies have quantified the natural recovery of vegetation of the rain forest after disturbance (*e.g.* Uhl 1987, Martínez-Ramos & García-Orth 2007) whereas other studies have evaluated the performance of tree species in plantings (*e.g.* Davidson *et al.* 1998, Hooper *et al.* 2002, Carpenter *et al.* 2004, dos Santos *et al.* 2006, Douterlunne *et al.* 2010). However, we are not aware of any study that compares the performance of naturally regenerating seedlings (*i.e.* recruits) and transplants in the same restoration setting to evaluate the success of minimal versus maximal restoration intervention.

Under minimal restoration intervention, natural succession takes place and pioneer species may establish. Pioneer tree species are those that colonize naturally in early successional environments, due to the high dispersal capacity of their small seeds and their rapid growth rates associated to high availability of resources (Swaine & Whitmore 1988, Whitmore 1989). Given that restoration plantings are expensive, pioneers are frequently selected because of high survival in such conditions (Davidson *et al.* 1998, dos Santos *et al.* 2006). On the other hand, late-successional non-pioneer species do not naturally recruit in early successional environments: they have large seeds dispersed by animals that do not cross open areas (Westoby 1998, Martínez-Garza & Howe 2003). However, some non-pioneers may perform as well as pioneers when transplanted to restoration areas in the tropics (Hooper *et al.* 2002, Carpenter *et al.* 2004) and they may arrive at older restoration plantings (De la Peña *et al.* 2013). Given that pioneer species naturally establish in early successional habitats, it is expected that they outperform non-pioneer species early in restoration settings. However, as succession takes place, may pioneers continue to outperform non-pioneers irrespective of their origin, recruited or transplanted?

When nursery-raised seedlings are transplanted, they may experience stress known as transplant shock. Transplant shock is defined as low survival and growth of seedlings transplanted to the field compared to naturally recruited seedlings (Close *et al.* 2005), usually due to the low contact between the roots of the seedlings and the soil (Burdett 1990). Transplant shock had been recorded in temperate forest (see for example, Bernier 1993); however, in the rain forest of Chiapas, Mexico, transplanted individuals realized higher survival than recruits established after direct seeding (Douterlunne *et al.* 2010). Transplant shock has never been evaluated for tropical tree species exposed to severely adverse environmental conditions of degraded areas; there, smaller recruits may be more affected by adverse soil conditions than larger transplants even after experiencing transplant shock.

Planning for restoration plantings in tropical rain forests involve the challenge of selecting appropriate species from around 53,000 species (Slik *et al.* 2015). Species selected for restoration plantings are usually those with high survival and growth rates in early successional environments (Vázquez-Yanes *et al.* 1999); however, evaluation of performance have been done for few species, mainly for forestry use (Evans & Turnbull 2004). Survival and growth of species in early successional environments may be predicted by their functional traits; a functional trait is a measurable property of the organisms that strongly influences its performance (McGrill *et al.* 2006). For example, in the cloud forest, transplanted species with large leaves and high dry matter content (Saldaña-Acosta *et al.* 2009) or in the rain forest, transplanted species with deep canopies (Martínez-Garza *et al.* 2013b) showed higher performance in restoration plantings. On the other hand, tree species that recruit naturally early in succession have functional traits associated to high dispersal capacity, as small seed size (Lohbeck *et al.* 2013). Also, seed size is related to establishment: larger seeds have higher establishment capacity (Coomes & Grubb 2003, Poorter *et al.* 2008). Plant strategies of both, transplants and naturally recruited trees include multiple functional traits (multivariate plant strategies; Weiher *et al.* 1999, Violle *et al.* 2007) which may coincide, so both successfully establish in early successional environments irrespective of their origin. Alternatively, individual functional traits that predict performance may be different among transplanted and naturally recruited trees, therefore individual traits may better predict performance of tree species than multivariate plant strategies.

The main objective of this study is to propose management and restoration recommendations

for restoration plantings of tropical tree species by using life-history, the origin of the seedlings and plant functional traits to predict survival and growth rates of 15 tropical tree species in 5 years old restoration settings. We tested the following hypotheses: 1) pioneer species (both recruits and transplants) perform better than non-pioneers; 2) transplants have lower performance than recruits because of “transplant shock”; 3) multivariate functional traits predict performance of tree species better than individual traits irrespective of seedling origin; multivariate plant strategy for high performance in early successional environments include larger leaves, deep canopies and small seeds. Finally, we give recommendations to select tree species for restoration plantings under different scenarios based in the life history category, functional traits, number of recruits registered during 5 years, and performance of tree species.

Methods

Research site. The Los Tuxtlas Biological Station (LTBS) lies within a reserve of 640 ha of lowland tropical rain forest in the state of Veracruz, southeast Mexico. The forest has a closed canopy ~35 m high where *Nectandra ambigens* (S.F. Blake) C.K. Allen (Lauraceae) is the most common species in the canopy and *Pseudolmedia oxyphyllaria* Donn. Sm. (Moraceae) and *Astrocaryum mexicanum* Liebm (Arecaceae) are abundant in the mid-canopy and understory, respectively (Bongers *et al.* 1988). Mean annual rainfall at the station from 1997 to 2007 was $4,275 \pm 404$ mm; the dry season extends from March to May and the rainy season from June to February; the mean annual maximum temperature was 28 °C (R. Coates, National University of Mexico, Veracruz, personal communication). Our site is a cow pasture that was grazed intensively for 30–40 years, and is embedded in primary and secondary forest on a hill 180–260 m above sea level, facing north-east to the Gulf of Mexico. Where cattle have access, pasture grasses are a closely cropped, 3–10 cm high mix of exotic [*Cynodon plectostachyus* (K. Schum.) Pilg., *Capriola dactylon* (L.) Kuntze, *Brachiaria decumbens* Stapf and *Brachiaria brizantha* (A. Rich.) Stapf] and native grasses [*Axonopus compressus* (Sw.) P. Beauv., *Panicum* spp., *Paspalum conjugatum* P.J. Bergius]. The soil is sandy loam classified as Vitric Andosol originating from basalt and andesite mixed with volcanic ash; the texture is mainly clay (48.5 %) with an acidic pH (4.9) (González-Soriano *et al.* 1997).

Experimental settings. A 3 × 8 grid of 24 fenced plots (30 × 30 m, each plot separated by 35 m) was established along an altitudinal gradient in a 12-ha pasture of the agricultural colony of Ruiz Cortines in August 2006, adjacent to the LTBS. Barbed wire fences were held up by living poles of *Gliricidia sepium* (Jaq.) Kunz (Fabaceae) every two meters. Plots on the grid are within 500–1,200 m of the edge of the LTBS and 90 m from the nearby secondary forest. Standing trees

Table 1. Family, Life history and the Origin of the seedling of 15 tropical tree species in restoration settings at Los Tuxtlas, Veracruz, Mexico.

Species	Family	Life History	Origin of Seedling
<i>Albizia purpusii</i> Britton & Rose	Mimosaceae	Non-pioneer	Both
<i>Amphitecna tuxtlensis</i> A.H. Gentry	Bignoniaceae	Non-pioneer	Transplanted
<i>Bernoullia flammea</i> Oliv.	Malvaceae	Non-pioneer	Transplanted
<i>Brosimum alicastrum</i> Sw.	Moraceae	Non-pioneer	Transplanted
<i>Bursera simaruba</i> (L.) Sarg.	Burseraceae	Non-pioneer	Recruited
<i>Cecropia obtusifolia</i> Bertol.	Cecropiaceae	Pioneer	Both
<i>Cedrela odorata</i> L.	Meliaceae	Pioneer	Both
<i>Cojoba arborea</i> (L.) Britton & Rose	Fabaceae	Non-pioneer	Transplanted
<i>Eupatorium galeottii</i> B. L. Rob.	Asteraceae	Pioneer	Recruited
<i>Ficus yoponensis</i> Desv.	Moraceae	Pioneer	Transplanted
<i>Helicocarpus appendiculatus</i> Turcz.	Malvaceae	Pioneer	Both
<i>Ochroma pyramidalis</i> (Cav. ex Lam.) Urb.	Malvaceae	Pioneer	Transplanted
<i>Tabebuia guayacan</i> (Seem.) Hemsl.	Bignoniaceae	Non-pioneer	Transplanted
<i>Tetrorchidium rotundatum</i> Standl.	Euphorbiaceae	Non-pioneer	Recruited
<i>Trema micrantha</i> (L.) Blume.	Cannabaceae	Pioneer	Recruited

within the 12 ha were cut in 2006 (Howe *et al.* 2010). From September to December 2006, 144 seedlings of each of 24 native tree species were transplanted to 16 of the 24 plots (see details in Martínez-Garza *et al.* 2013a). Seedlings were 4–7 months old at the time of planting, and their average height was 17.8 cm (range 5–40 cm across species). The entire area for the plantings was 0.92 ha (16 plots), and the area where natural recruitment could occur was 1.38 ha, including eight additional plots where no plantings were established.

Data collection. For this study, we selected 15 species following three criteria: 1) given that individuals were considered replicates, only species with more than 17 individuals across all plots were selected; 2) transplanted species that were also recorded as recruited and 3) a similar number of pioneer and non-pioneer species. We selected 11 species transplanted and eight registered as naturally recruited; four species had individuals from both origins (Table 1). *A priori* distinction between pioneer and non-pioneer species was based on the literature (Martínez-Ramos 1985, Popma *et al.* 1992, Ibarra-Manríquez & Oyama 1992). This study includes two censuses: June 2007 and 2011; at each census, we measured height and diameter at the base of all individuals transplanted and recruited of the 15 tree species selected. Growth rates were calculated as the difference between the first measurement and the last, divided by the number of months elapsed. Performance of all tree species transplanted (Martínez-Garza *et al.* 2013b) and richness and abundance of recruits (de la Peña-Domene *et al.* 2013) has been evaluated before. The present study adds an explicit comparison of survival and growth rates of recruits with a pertinent subset of the transplanted species (see above) and their functional traits.

Data Analysis. Performance of pioneer and non-pioneer species.- Two one-way analyses of variance (ANOVA) were carried out to test for differences in diameter and height growth between pioneer and non-pioneer species. The life history category was the independent variable. The proportion of survivors for each species at all the plots (seven pioneer species and eight non-pioneer species) was analyzed with a General Linear Model (GLM). For the analysis of growth rates, individuals were used as replicates ($N = 319$ pioneers and $N = 277$ non-pioneers): given that they were growing in mixed stands at each plot under variable soil conditions, we consider the performance of each plant to be independent.

Performance of transplants and recruits.- To evaluate transplant shock we tested survival, diameter and height growth rates of recruits and transplants for the four species with individuals in both groups (*Cecropia obtusifolia*, *Cedrela odorata*, *Helicarpus appendiculatus* and *Albizia purpusii*). For these species, we had more transplants than recruits; therefore, we randomly selected the number of transplants necessary to balance the data points we had for recruits of each species (up to 15 individuals for each species; see Appendix 2). T-tests were used to evaluate survival and growth rates in diameter and height for recruits and transplants. Height and diameter growth rates of *Albizia purpusii* and *Cecropia obtusifolia* were transformed with the natural logarithm to homogenize variances. Overall species, life-history was not included in the analysis because variances for the interaction with origin of the seedling could not be homogenized. To evaluate survival overall recruits ($N = 8$ species) and transplants ($N = 11$ species), species were used as replicates; individuals at all the plots were considered as one population. To evaluate growth rates overall recruited and transplanted species, a subset of individuals (up to 15 individuals for each species) were chosen randomly to balance sample size (see above; $N = 91$ recruits and $N = 138$ transplants).

Functional traits as predictors of species performance.- To evaluate multivariate plant strategies, a principal components analysis (PCA; Duntzman 1989) was run using 12 functional traits related to leaf display (Leaf Area, Leaf Mass per unit Area [LMA], Leaf Dry Matter Content [LDMC]), tree architecture (Crown Area, Crown Length, Height to the First Branch [Height^{1st} B], Maximal tree height), and reproduction (Flower size [Flower S], Fruit size [Fruit S], Fruit Weight [Fruit W], Seed Weight [Seed W] and Seed Number per Fruit [Seed Number]) (Appendix 1). Leaf and architectural traits were measured in the same plants measured for performance whereas reproductive traits were taken from Ibarra-Manríquez & Oyama (1992). Given that tree architecture changes when trees increase in size, plants were compared at a standardized size of 30 mm of stem diameter at the base. To this end, species-specific regression curves were

obtained relating the architectural trait of interest to stem diameter. When the regression of a trait was not significant we used the average of stem diameter calculated with the five closest values to 30 mm stem diameter. We used all the single functional traits and the two first axes of the PCA to predict tree performance using regressions.

Species for different restoration strategies.- to give restoration and management recommendations for different scenarios, tree species were grouped based in their life-history category, survival, growth rates and number of individuals naturally recruited during 5 years in the restoration settings.

All analyses were done with STATISTICA (StatSoft 2004). Variables were back transformed to report original values in the result section. Means and standard deviation of variables are shown throughout results. From now on, species are mentioned by genus only.

Results

Performance of pioneer and non-pioneer species. Survival varied six times among species. The non-pioneer *Albizia* showed the highest survival (100 %) whereas the non-pioneer *Brosimum* showed the lowest (15.6 %; Appendix 1). Overall recruits and transplants, pioneer species showed higher survival ($75.61 \pm 8.43 \%$, $N = 7$) than non-pioneer species ($50.45 \pm 7.88 \%$, $N = 8$); the GLM reveled significantly higher survival of pioneer species ($F_{(1,13)} = 4.75$, $P < 0.05$).

Diameter growth varied an order of magnitude among species. The pioneer *Ochroma* had the highest diameter growth rates (11.6 ± 6.4 mm/month) whereas the non-pioneer *Brosimum* realized the lowest (0.5 ± 0.2 mm/month; Appendix 1). Pioneer species showed two times higher diameter growth rates (3.8 ± 3.2 mm/month) than non-pioneer species (1.59 ± 1.03 mm/month). The analysis of variance reveled significantly higher growth rates in diameter of pioneer species compared with non-pioneers ($F_{(1,595)} = 6.44$, $P < 0.03$).

Height growth rates varied three times among tree species. The pioneer *Helicocarpus* showed the highest growth rates in height (11.0 ± 3.0 cm/month) whereas the non-pioneer *Amphitecna*

Table 2. Pearson correlation coefficients (r) between 12 functional traits and survival (%), growth rate in height (cm/month) and diameter (mm/month) for 15 tree species recruited and transplanted to restoration settings at Los Tuxtlas, Veracruz, Mexico. Correlations with the axes 1 and 2 from the PCA (multivariate trait axes; see Fig. 2) are also shown. LMA refers to leaf mass per unit area, and LDMC to leaf dry matter content. Coefficients of determination (r^2) are shown in the figures for selected correlations.

Functional Traits	Survival	Growth rates	
		Height	Diameter
Foliar			
Leaf Area (cm ²)	0.40	0.70**	0.55
LMA (g/m ²)	-0.44	0.00	0.00
LDMC (g/g)	0.37	0.35	0.00
Architectural			
Crown Area (cm ²)	0.42	0.76**	0.68**
Crown Length (cm)	0.32	0.66**	0.54*
Height 1 st Branch (cm)	0.14	0.30	0.00
Maximal tree height (m)	0.00	0.20	0.14
Reproductive			
Flower size (mm)	0.35	0.14	0.63**
Fruit size (mm)	0.00	-0.10	0.00
Fruit weight (g)	0.41	-0.17	0.00
Seed weight (g)	-0.65**	-0.46*	0.28
Seed Number per Fruit	0.14	-0.39	0.24
Axis from PCA			
Axis 1	0.56*	0.86***	0.47*
Axis 2	0.33	0.00	0.00

* $P < 0.05$, ** $P < 0.01$, *** $P < 0.001$

Table 3. T-test for Survival (%), growth rates in Diameter (mm/month) and Height (cm/month) of individuals naturally recruited and transplanted to restoration settings at Los Tuxtlas, Veracruz, Mexico. A t-test of performance overall the four species is also shown. Means and standard errors are shown.

Performance				Species		
		<i>Albizia</i>	<i>Helicocarpus</i>	<i>Cedrela</i>	<i>Cecropia</i>	Overall
Survival	Recruits	100.00	70.83	85.71	94.44	87.75 ± 6.35
	Transplants	66.66	85.41	56.25	87.50	73.96 ± 7.53
						$t_{(1,6)} = 1.40\text{NS}$
Diameter	Recruits	1.19 ± 1.64	3.17 ± 1.52	1.33 ± 0.28	1.06 ± 1.12	1.10 ± 1.20
Growth	Transplants	0.86 ± 1.60	7.58 ± 0.91	1.12 ± 0.42	4.26 ± 1.15	3.12 ± 1.20
Rates		$t_{(1,10)} = 0.47\text{ NS}$	$t_{(1,24)} = 2.49^*$	$t_{(1,8)} = 0.43\text{NS}$	$t_{(1,24)} = 7.70^{***}$	$t_{(1,72)} = 3.04^{**}$
Height	Recruits	2.19 ± 2.16	12.53 ± 0.77	3.87 ± 0.70	15.44 ± 1.76	8.13 ± 1.20
Growth	Transplants	1.59 ± 1.44	8.32 ± 0.75	3.42 ± 1.54	15.52 ± 1.60	6.41 ± 1.18
Rates		$t_{(1,10)} = 0.38\text{ NS}$	$t_{(1,24)} = 3.89^{***}$	$t_{(1,8)} = 0.26\text{ NS}$	$t_{(1,24)} = 0.04\text{ NS}$	$t_{(1,72)} = 0.97\text{ NS}$

* $P < 0.05$, ** $P < 0.01$, *** $P < 0.001$; NS = Not significant

showed the lowest (3.40 ± 1.09 cm/month). Pioneers showed significantly higher growth rates in height (9.39 ± 3.66 cm/month) than non-pioneer species (5.85 ± 1.59 cm/month; $F_{(1,595)} = 5.98$, $P < 0.02$).

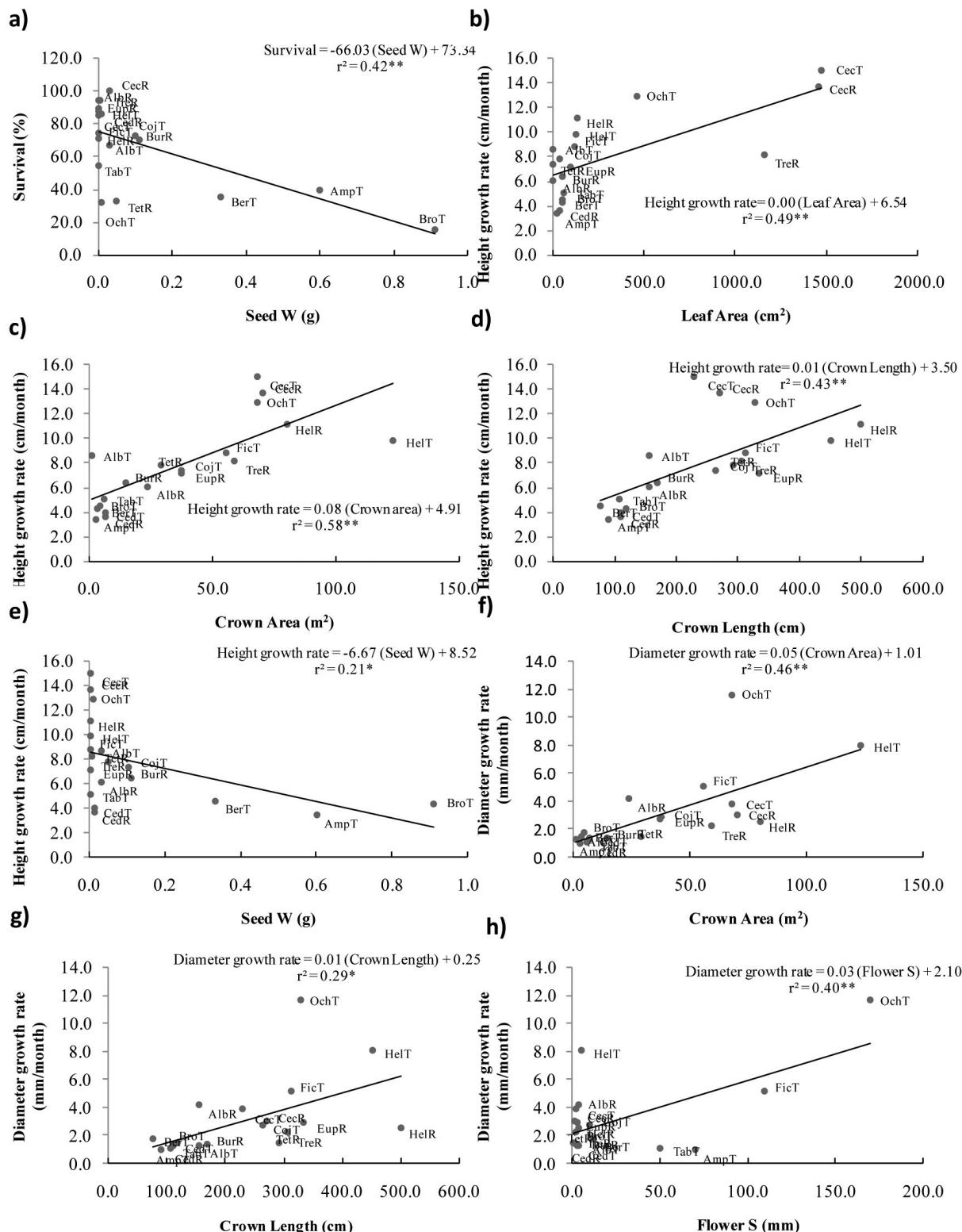
Performance of recruits and transplants. Species with recruits and transplants (4 spp).- For three pioneer species (*Helicocarpus*, *Cecropia* and *Cedrela*) and one non-pioneer species (*Albizia*), we recorded performance of recruits and transplants. Transplants of *Cedrela* and *Albizia* suffered $> 30\%$ decrease in its survival compared to naturally recruited individuals whereas *Cecropia* experienced a 7 % reduction of survival when transplanted (Table 3). On the other hand, *Helicocarpus* showed an increment of 17 % in survival when transplanted (Table 3). Overall these four species, recruits showed higher survival ($87.75 \pm 6.35\%$) than transplants ($73.96 \pm 7.53\%$), however, analysis revealed that differences in survival were not statistically significant ($t_{(1,6)} = 1.40$, $P > 0.74$).

The non-pioneer *Albizia* and the pioneer *Cedrela* showed a similar 27 % decrease in diameter growth rates when transplanted, however, diameter growth rate were not statistically different by origin for both species (Table 3). On the other hand, transplanted individuals of *Helicocarpus* and *Cecropia* showed significantly higher diameter growth rates (7.58 and 4.26 mm/month respectively) than recruits (3.17 and 1.06 mm/month; Table 3). Overall the four species, recruits showed significantly lower diameter growth rates (1.10 ± 1.20 mm/month; $N = 36$) than transplants (3.13 ± 1.20 mm/month; $N = 36$) ($t_{(1,72)} = 3.04$, $P < 0.01$; Table 3).

Transplants of *Albizia*, *Cedrela* and *Cecropia* showed 30, 10 and 9 % higher growth rates in height respectively, than naturally recruited individuals; however, height growth rates were statistically similar for these species (Table 3). *Helicocarpus* transplants decreased significantly its height growth rates by 11 % compared to recruits (Table 3). Overall the four species, recruits showed higher height growth rates (8.13 ± 1.20 cm/month) than transplants (6.41 ± 1.18 cm/month) but differences were not statistically significant ($t_{(1,72)} = 0.97$, $P > 0.73$).

All species (11 spp).- Survival of 11 transplanted species (5 pioneer and 6 non-pioneer species) varied 5 times: from 15 % (*Bernoullia*) to 87.5 % (*Cecropia*). Survival of naturally recruited species (5 pioneer and 3 non-pioneer species) was $> 70\%$ excepting for the non-pioneer *Tetrorchidium* (33 %; Appendix 2). Overall species, recruits showed significantly higher survival ($80.51 \pm 19.25\%$) than transplants ($56.32 \pm 22.31\%$; $F_{(1,18)} = 5.43$, $P < 0.03$).

Diameter growth rates of transplanted individuals of 11 species varied 12 times, from 0.93 ± 0.72 mm/month (*Amphitecna*) to 11.6 ± 5.42 mm/month (*Ochroma*) whereas diameter growth rates of recruits (eight species) varied 4 times from 1.19 ± 0.61 mm/month (*Cedrela*) to 4.16 ± 3.29 mm/month (*Albizia*; Appendix 2). Overall species, transplanted individuals showed higher



* $P < 0.05$, ** $P < 0.01$, *** $P < 0.001$

Figure 1. Correlations between performance and individual functional traits: Survival with (a) Seed weight; Height growth rates with (b) Leaf Area, (c) Crown Area, (d) Crown Length and, (e) Seed Weight. Diameter growth rates with (f) Crown Area, (g) Crown Length and (h) Flower Size. Values of r^2 , regression lines and equations are shown. Acronyms refer to the first three letter of the genus name of 15 species recruited (R) or transplanted (T) on restoration settings at Los Tuxtlas, Veracruz, Mexico

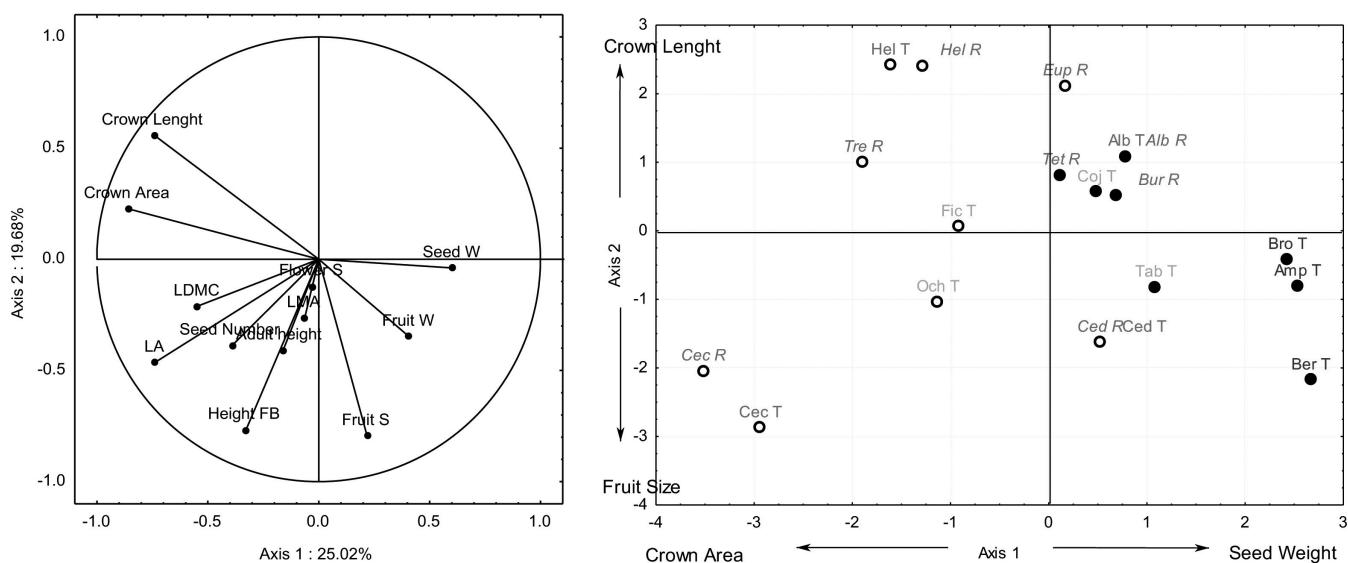


Figure 2. (a) Trait loading and (b) species scores of PCA axes 1 and 2 of an ordination based on 12 functional traits of 15 tree pioneer (open symbols) and non-pioneer (close symbols) species recruited (R) or transplanted (T) on restoration settings at Los Tuxtlas, Veracruz, Mexico. Groups of species for different restoration strategies are shown: good recruiters (red letters), good transplants (green letters) and poor transplants (blue letters). Acronyms refer to the first three letter of the genus name.

diameter growth rate (1.88 ± 1.10 mm/month; $N = 138$) than recruits (1.51 ± 1.12 mm/month; $N = 91$); however, analysis revealed that differences in diameter were not statistically significant; $F_{(1, 227)} = 2.12$, $P > 0.15$.

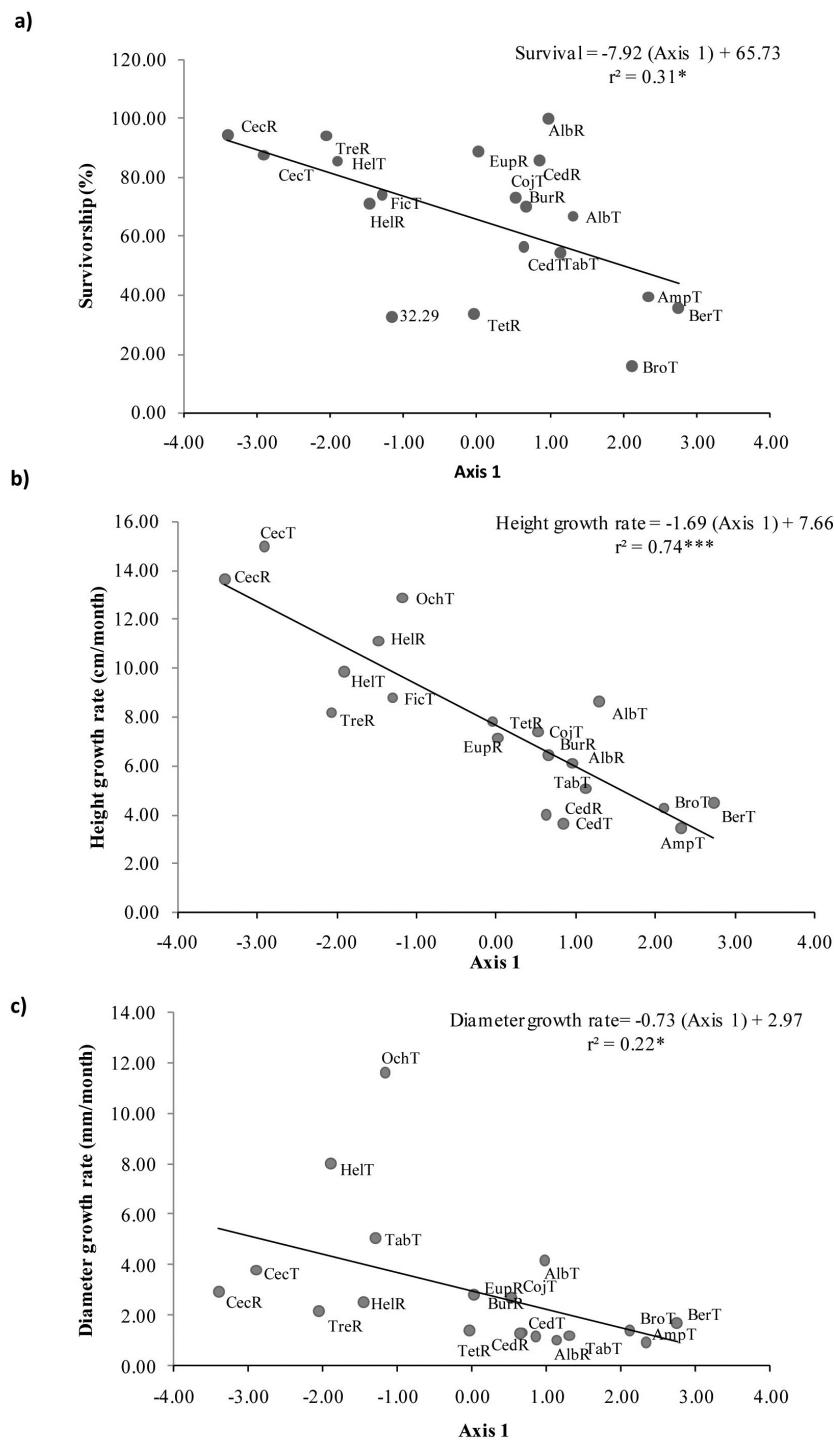
Height growth rates of recruits and transplants had similar variation among species, ca. 4 times; *Cecropia* showed the highest height growth rates of recruits (13.64 ± 5.84 cm/month) and transplants (14.98 ± 60.32 cm/month) whereas *Cedrela* (3.59 ± 1.55 cm/month) and *Amphitecna* (3.4 ± 1.09 cm/month) showed the lowest height growth rates for recruits and transplants respectively (Appendix 2). Overall species, recruits realized significantly higher growth rates in height (7.73 ± 1.08 cm/month) than transplants (5.70 ± 1.07 cm/month); $F_{(1, 227)} = 7.58$, $P < 0.001$.

Functional traits as predictors of species performance. Considering 8 species of recruits and 11 species of transplants, survival was negatively correlated with Seed Weight ($r = -0.65$, $P < 0.01$; Figure 1a; Table 2). Height growth rate was positively correlated with Leaf Area ($r = 0.70$, $P < 0.01$; Figure 1b), Crown Area ($r = 0.76$, $P < 0.001$; Figure 1c), Crown Length ($r = 0.66$, $P < 0.01$; Figure 1d) and negatively with Seed Weight ($r = -0.46$, $P < 0.05$; Figure 1e). Diameter growth rate was positively correlated with Crown Area ($r = 0.68$, $P < 0.01$; Figure 1f), Crown Length ($r = 0.54$, $P < 0.05$; Figure 1g), and Flower Size ($r = 0.63$, $P < 0.01$; Figure 1h).

The first two axes of the PCA explained 44.70 % of the variation among species (Figure 2a). The PCA axis 1 was related to decreasing Crown Area and increasing Seed Weight while axis 2 was related to increasing Crown Length and decreasing Fruit Size (Figure 2b). Axis 1 of the PCA was positively correlated with survival ($r = 0.56$, $P < 0.05$; Figure 3a), Height Growth rates ($r = 0.86$, $P < 0.001$; Figure 3b) and Diameter Growth rates ($r = 0.47$, $P < 0.05$; Figure 3c; Table 2). Axis 2 of the PCA was not correlated with tree performance (Table 2).

Groups of species for different restoration strategies. Tree species were divided in three groups based on their life history category, performance of recruits and transplants, functional traits and the number of new recruits recorded during the first five years of the experiment (Appendix 2). The first group was named “Good recruiters” and it included five pioneer and three non-pioneer species. For these eight species, we recorded, 31 ± 17 recruits in average, excluding *Bursera* which recruited most individuals (337 individuals; Appendix 2). These species showed high survival (77 ± 19.19 %), height growth rates (8.44 ± 3.49 cm/month; Appendix 2) and long crowns

Figure 3. Correlations between performance and multivariate functional traits. PCA axis 1 with (a) Survival, (b) Height Growth rates and, (c) Diameter growth rates. Values of r^2 , regression lines and equations are shown. Acronyms refer to the first three letter of the genus name of 15 species recruited (R) or transplanted (T) on restoration setting at Los Tuxtlas, Veracruz, Mexico.



* $P < 0.05$, ** $P < 0.01$, *** $P < 0.001$.

(Figure 2b). The second group was named “*Good transplants*” and included two pioneer and two non-pioneer species. The species in this group had intermediate values of survival ($58.32 \pm 19.59\%$), height growth rates (8.50 ± 3.29 cm/month) and functional traits (Figure 2b). They showed high diameter growth rates (5.10 ± 4.01 mm/month) and only one recruit was registered for this group (Appendix 2). The third group was named “*Poor transplants*” and included three non-pioneer species. Species in this group had the lowest survival ($30.13 \pm 12.72\%$) and height growth rates (4.03 ± 0.55 cm/month) and the heaviest seeds (Figure 2b). No recruits were registered for this group (Appendix 2).

Discussion

To guide the selection of species for restoration plantings, the life-history, the origin of the seedlings (recruits or transplants), and plant functional traits were used to predict performance of 15 tropical tree species. Overall 15 species, pioneer species had higher performance than non-pioneers whereas recruits had higher performance than transplants. The crown area was the individual functional traits most powerful to predict height growth rates whereas the multivariate plant trait axis 1 related to seed weight and crown area showed even higher power of prediction of height growth rates. Species that recruit well in early successional environments do not need to be planted whereas transplants with low survival and growth rates should be introduced at later stages of succession.

Performance of pioneer and non-pioneer species. According to our predictions, pioneer showed higher survival and growth rates than non-pioneers after 5 years. Our results agree with what has been reported in other studies: for example, in a rain forest in Brazil, seven pioneer species showed higher survival (95 %) compared with one non-pioneer species (65 %) after 2 months of planting (dos Santos *et al.* 2006). In the humid forest of Ecuador's Amazonia, seven pioneers showed twice the survival (90 %) of eight non-pioneers (45 %) after 2.5 years of planting (Davidson *et al.* 1998). In plantings located in Ecuador, Brazil and Mexico, at the short term (< 3 years of planting) pioneer species had higher growth rates than non-pioneer species (Davidson *et al.* 1998, dos Santos *et al.* 2006, Román-Dañobeytia *et al.* 2012). Pioneer species seem to realize higher survival and growth rates at the short-term (< 5 years of planting) under the aggressive environmental conditions of pastures.

Performance of recruits and transplants. According to studies in the temperate forest, plantings may experience transplant shock which is observed as a decrease in performance when planted in the field (Close *et al.* 2005). This negative effect may last from a couple of years to decades (South & Zwolinski 1997). For this study, overall the four species with naturally recruited individuals and transplants, recruits showed similar survival and height growth rates compared to transplants. Therefore, it seems that, transplant shock in terms of these two measures of performance had been overcome after 5 years. This is the first record of transplant shock length evaluated for native tree species in the tropics.

Transplants may suffer from a low contact between the roots and the soil which causes water limitation (Burdett 1990). In this scenario, drought stress may trigger an increment in the growth of roots into deeper soil layers to improve uptake of water (Larcher 1985). In our study, transplants might have suffered some water limitation. This is supported by higher diameter growth rates of transplant compared to recruits; basal diameter has been related to root biomass for juvenile trees growing in experimental plantings (Martínez-Garza *et al.* 2013a). Evaluation at species level revealed a contrasting response of *Helicocarpus*: recruits of this species showed higher height growth rates than transplants suggesting that this species have not overcome transplant shock. For the other species, the increment in stem diameter of transplants in response to drought stress may have allowed them to reach similar survival and height growth rates compared to recruits, ending transplant shock after 5 years.

Under minimal intervention, pioneer species arriving by dispersal events initiate succession (Holl 1999, Martínez-Garza *et al.* 2009, Chazdon 2014) whereas plantings, as a maximal restoration intervention are frequently established with pioneer transplants (Lamb *et al.* 2005, Duterlouge *et al.* 2008). Others suggest to plant as many pioneer and late-successional species as possible (framework species; Tucker & Murphy 1997) or mostly late-successional species to skip early successional stages (Martínez-Garza & Howe 2003). If recruited tree species may outperform transplants, why spend resources establishing plantings? Far away from old growth forest, composition of regenerating forest may remain with few species for decades (Finegan 1996; pioneer desert, Martínez-Garza & Howe 2003), on the other hand, plantings implicate choosing tree species to accelerate and somehow guide succession. Further, plantings ameliorate environmental conditions and favor recruitment of mostly pioneers at the beginning (de la Peña *et al.* 2013) and also non-pioneer species at later stages (de la Peña *et al.* 2014). Plantings

in island arrangements have also proven useful to favor natural recruitment in large areas at lower cost (Zahawi *et al.* 2013). To maximize natural recruitment at the lowest cost, different densities and arrangement of plantings should be tested; however, selection of successful species from a large pool (~53,000 species) for different scenarios is still a challenge.

Functional traits as predictors of the species performance. The multivariate plant strategy related to seed weight and crown area predicted performance better than individual functional traits, according to our expectation. By individual traits, species with smaller seed weight had higher survival and height growth rates whereas species with high crown area had higher diameter growth rates. In the multivariate trait space, species with high scores on the first strategy axis were the ones with large seeds contained in heavy fruits and smaller crown areas as *Bernoulillia*, *Amphitecna* and *Brosimum*. Reproductive traits as high seed mass and few seeds per fruit are characteristics of non-pioneers species that establish in small gaps within the forest; they show lower inherent growth rates and poor dispersal (Martínez-Ramos 1985, Denslow 1987, Paine *et al.* 2015) and therefore, they do not recruit in early successional environments (de la Peña *et al.* 2013). Further, they show a slow life history which refers to low population growth rates strongly influenced by survival (Adler *et al.* 2014). The second axis separated species that developed deep crowns versus species with small fruits as *Cecropia* and *Ochroma*. These two species are well-known pioneers (Martínez-Ramos 1985), they have small seeds which move easily in fragmented landscapes (Estrada *et al.* 1984) and fast life histories with high population growth rates influenced by fecundity (Adler *et al.* 2014). Individual traits and multivariate strategies showed that pioneer or non-pioneer species with small seeds and large crowns may show higher performance in early successional environments as natural recruits or if transplanted, however, not all species need to be transplanted.

Groups of species for different restoration strategies. Tropical deforestation has multiple origins and causes corresponding to various biophysical and socio-economic conditions and therefore, it generates different scenarios for restoration (reviewed in Ceccon *et al.* 2015). Minimal intervention may be enough in some cases (Holl & Aide 2011) whereas other situation may call for maximal intervention with the use of plantings. To select successful tree species for restoration plantings, we usually choose species known to have higher performance in early successional environments. However, those few pioneer species with well-known performance in early successional environments may not need to be planted: in our study, the species in the *Good recruiters* category are natural colonizers (Dalling *et al.* 1997). They produce many seeds with high dispersal ability; for example, an average of 0.64 seeds/m² of these species were recorded arriving to experimental plots (Martínez-Garza *et al.* 2009). Even when all these species are also successful if planted (Vázquez-Yanes *et al.* 1999), we suggest favoring its natural recruitment in landscapes including forest fragments. Plantings (maximal intervention) are more expensive than establishing cattle exclusions to favor recruitment (minimal intervention). For example, each individual of *Cedrela odorata* propagated in a Mexican nursery costs \$0.84 USD (reviewed in Guzman-Luna 2012); to establish the plantings, additional cost of transplant should be added. Further, the *Good recruiters* showed a seed size that varied 2 orders of magnitude (Appendix 1); given that only small seeds arrive more than 10 m from the forest border to open areas (Aide & Cavelier 1994, Martínez-Garza & Gonzalez-Montagut 1999), direct seeding should be considered to favor recruitment of larger seeded species (Camargo *et al.* 2002, Cole *et al.* 2011). Finally, these species develop large crowns that shade out the grasses and may serve as attractors of frugivorous fauna that bring large forest seeds from the mature forest.

The species in the *Good transplants* category are pioneers or non-pioneer species with intermediate values of all functional traits (Figure 2b); they were not registered in the seed rain for the first year of the experiment except for *Ochroma*, for which a seed rain of 0.01 seeds/ m² was recorded (Martínez-Garza *et al.* 2009). These species may show higher performance as transplant than by direct seeding (Douterlunne *et al.* 2010). Therefore, even when the species in this group are not usually recorded as establishing in early successional environments, they are successful if transplanted. Finally, the species in the *Poor transplants* category are all non-pioneers with large seeds (≥ 0.3 g; Appendix 1) and they have never been recorded in the seed

rain of early successional environments (Martínez-Garza *et al.* 2009). We suggest transplanting these species or testing its introduction by direct seeding once a canopy has developed.

To recover the structure and processes of tropical forest, we recommend taking into account landscape matrix: restoration areas in complex matrices, those than include isolated trees, living fences or forest fragments should be excluded from disturbance to favor natural recruitment. In degraded isolated areas invaded by exotic grasses, plantings of a mix of small-seeded pioneer and non-pioneer species should be established for a rapid recovery of a canopy to suppress grasses. In any scenario, we recommend further augmentation of tree diversity with direct seeding or planting of non-pioneer species with large seeds after a canopy develops.

Acknowledgments

We thank two anonymous reviewers for helpful comments on early versions of this manuscript. We are grateful to Rene Gadea, Eladio Velazquez and Luz Maria Ayestarán for expert field assistance and to Karen Holl for discussion with her group of students that benefited early versions of this manuscript. The authors also gratefully acknowledge financial and logistical support from the University of Illinois at Chicago, the Universidad Autónoma del Estado de Morelos (UAEM), the Universidad Nacional Autónoma de México (UNAM), the Director and staff of the LTBS (UNAM). This work was supported by the National Science and Foundation (#DEB 0516259 to Henry F. Howe and C-MG), the National Council of Science and Technology of Mexico (CONACyT) provided a scholarship to A-GL and the Packard Foundation provided support for field work. AGL was supported in part by the University of California in the Lab of Karen Holl while writing.

Literature cited

Adler PB, Salguero-Gómez R, Compagnoni A, Hsu JS, Ray-Mukherjee J, Mbeau-Ache C, Franco M. 2014. Functional traits explain variation in plant life history strategies. *Proceeding of National Academy of Sciences* **111**:740-745. DOI: 10.1073/pnas.1315179111

Aide TM & Cavelier J. 1994. Barriers to lowland tropical forest restoration in the Sierra Nevada de Santa Martha, Colombia. *Restoration Ecology* **2**:219-229. DOI: 10.1111/j.1526-100X.1994.tb00054.x

Bernier PY. 1993. Comparing natural and planted black spruce seedlings. 1. Water relations and growth. *Canadian Journal of Forest Research* **23**:2427-2434. DOI: 10.1139/x93-300

Bongers F, Popma J, Meave del Castillo J, Carabias J. 1988. Structure and floristic composition of the low land rain forest of Los Tuxtlas, Mexico. *Vegetatio* **74**:55-80. DOI: 10.1007/BF00045614

Burdett AN. 1990. Physiological processes in plantation establishment and the development of specifications for forest planting stock. *Canadian Journal Forest Research* **20**:415-427. DOI: 10.1139/x90-059

Camargo CJL, Kossman-Ferraz ID, Imakawa MA. 2002. Rehabilitation of degraded areas of central Amazonia using direct sowing of forest tree seeds. *Restoration ecology* **10**:636-644. DOI: 10.1046/j.1526-100X.2002.01044.x

Carpenter FL, Nichols JD, Sandi E. 2004. Early growth of native and exotic trees planted on degraded tropical pasture. *Forest Ecology and Management* **196**:367- 378. DOI: 10.1016/j.foreco.2004.03.030

Ceccon E, Barrera-Cataño JI, Aronson J, Martínez-Garza C. 2015. The socioecological complexity of ecological restoration in Mexico. *Restoration Ecology* **23**:331-336. DOI: 10.1111/rec.12228

Chazdon RL. 2014. *Second growth: The promise of tropical forest regeneration in an age of deforestation*. Chicago: University of Chicago Press.

Close DC, Beadle CL, Brown PH. 2005. The Physiological basis of containerised tree seedling 'transplant shock': a review. *Australian Forestry* **68**:112-120. DOI: 10018/913679

Cole RJ, Holl KD, Keene CL, Zahawi RA. 2011. Direct seeding of late-successional trees to restore tropical Montane forest. *Forest Ecology and Management* **261**:1590-1597. DOI: 10.1016/j.foreco.2010.06.038

Coomes DA & Grubb PJ. 2003. Colonization, tolerance, competition and seed-size variation within functional groups. *Trends in Ecology & Evolution* **18**:283-291. DOI: 10.1016/S0169-5347(03)00072-7

CONAFOR. 2010. Comisión Nacional Forestal. <conafor.gob.mx> (Accessed February 13, 2010).

Dalling JW, Swaine MD, Garwood NC. 1997. Soil seed bank community dynamics in seasonally moist lowland tropical forest, Panama. *Journal of Tropical Ecology* **13**:659-680. DOI: 10.1017/S0266467400010853

Davidson R, Gagnon D, Mauffette Y, Hernandez H. 1998. Early survival, growth and foliar nutrients in

native Ecuadorian trees planted on degraded volcanic soil. *Forest Ecology and Management* **105**:1-19. DOI: 10.1016/S0378-1127(97)00295-8

de la Peña-Domene M, Martínez-Garza C, Howe HF. 2013. Early recruitment dynamics in tropical restoration. *Ecological Applications* **23**:1124-1134. DOI: 10.2307/23441611

de la Peña-Domene M, Martínez-Garza C, Palmas-Pérez S, Rivas-Alonso E, Howe HF. 2014. Roles of birds and bats in early tropical-forest restoration. *Plos One*, **9**, e104656. DOI: 10.1371/journal.pone.0104656

Denslow JS. 1987. Tropical rainforest gaps and tree species diversity. *Annual Review of Ecology and Systematics* **18**:431-451. DOI: 10.1146/annurev.es.18.110187.002243

dos Santos Jr. UM, Gonçalves JFdC, Feldpausch TR. 2006. Growth, leaf nutrient concentration and photosynthetic nutrient use efficiency in tropical tree species planted in degraded areas in central Amazonia. *Forest Ecology and Management* **226**:299-309. DOI: 10.1016/j.foreco.2006.01.042

Douterlunge DN, Levy-Tacher S, Golicher DJ, Román-Dañoebetya F. 2010. Applying indigenous knowledge to the restoration of degraded tropical rain forest clearings dominated by bracken fern. *Restoration Ecology* **18**:322-329. DOI: 10.1111/j.1526-100X.2008.00459.x

Estrada A, Coates-Estrada R, Vázquez-Yanes C. 1984. Observations on fruiting and dispersers of *Cecropia obtusifolia* at Los Tuxtlas, Mexico. *Biotropica* **16**:315-318. DOI: 2387942

Evans J & Turnbull JW. 2004. *Plantation Forestry in the Tropics*. Oxford: Oxford University Press.

Finegan B. 1996. Pattern and process in Neotropical secondary rain forests: The first 100 years of succession. *Trends in Ecology & Evolution* **11**:119-124. [http://dx.doi.org/10.1016/0169-5347\(96\)81090-1](http://dx.doi.org/10.1016/0169-5347(96)81090-1)

González-Soriano E, Dirzo R, Vogt RC, eds 1997. *Historia natural de Los Tuxtlas*. México City: Instituto de Biología, Universidad Nacional Autónoma de México.

Guzmán-Luna A. 2012. Éxito de árboles provenientes de reclutamiento natural y propagados en vivero en plantaciones de restauración ecológica. MSc. Thesis, Universidad Nacional Autónoma de México.

Holl KD. 1999. Factors limiting tropical rain forest regeneration in abandoned pasture: seed rain, seed germination, microclimate and soil. *Biotropica* **31**:229-242. DOI: 10.1111/j.1744-7429.1999.tb00135.x

Holl KD & Aide TM. 2011. When and where to actively restore ecosystems? *Forest Ecology Management* **261**:1558-1563. DOI: 10.1016/j.foreco.2010.07.004

Hooper E, Condit R, Legendre P. 2002. Responses of 20 native tree species to reforestation strategies for abandoned farmland in Panama. *Ecological Applications* **12**:1626-1641. DOI: 10.1890/1051-0761(2002)012[1626:RONTST]2.0.CO;2

Howe HF, Urincho-Pantaleon Y, de la Peña-Domene M, Martínez-Garza C. 2010. Early seed fall and seedling emergence: precursors to tropical restoration. *Oecologia* **164**:731-740. DOI: 10.1007/s00442-010-1669-2

Ibarra-Manríquez G & Oyama K. 1992. Ecological correlates of reproductive traits of Mexican rain forest trees. *American Journal of Botany* **79**:383-394. DOI: 2445150

Lamb D, Erskine PD, Parrotta JA. 2005. Restoration of degraded tropical forest landscapes. *Science* **310**:1628-1632. DOI: 10.1126/science.1111773

Larcher W. 1995. *Physiological Plant Ecology: Ecophysiology and Stress Physiology of Functional Groups*. Springer-Verlag, Berlin.

Lohbeck M, Poorter L, Lebrilla-Trejos E, Martínez-Ramos M, Meave JA, Paz H, Pérez-García EA, Romero-Pérez IE, Tauro A, Bongers F. 2013. Successional changes in functional composition contrast for dry and wet tropical forest. *Ecology* **94**:1211-1216. DOI: 10.1890/12-1850.1

Martínez-Garza C & González-Montagut R. 1999. Seed rain from forest fragment into tropical pastures in Los Tuxtlas, Mexico. *Plant Ecology* **145**: 255-265. DOI: 10.1023/A:1009879505765

Martínez-Garza C & Howe HF. 2003. Restoring tropical diversity: beating the time tax on species loss. *Journal of Applied Ecology* **40**:423-429. DOI: 10.1046/j.1365-2664.2003.00819.x

Martínez-Garza C, Flores-Palacios A, de la Peña-Domene M, Howe HF. 2009. Seed rain in a tropical agricultural landscape. *Journal of Tropical Ecology* **25**:541-550. DOI: DOI: 10.1017/S0266467409990113

Martínez-Garza C, Tobon W, Campo J, Howe HF. 2013a. Drought mortality of tree seedlings in an eroded tropical pasture. *Land Degradation & Development* **24**:287-295. DOI: 10.1002/lrd.1127

Martínez-Garza C, Bongers F, Poorter L. 2013b. Are functional traits good predictors of species performance in restoration planting in tropical abandoned pastures? *Forest Ecology and Management* **303**:35-45. DOI: 10.1016/j.foreco.2013.03.046

Martínez-Garza C, Osorio-Beristain M, Alcalá-Martínez R, Valenzuela-Galván D, Mariano N. 2016. Ocho años de restauración experimental en las selvas estacionales de México. In: Ceccon E & Martínez-Garza C, eds. *Experiencias mexicanas en la restauración de ecosistemas*. Mexico City: UNAM, CONA-BIO, UAEM, 385-406.

Martínez-Ramos M. 1985. Claros, ciclos vitales de los árboles tropicales y regeneración natural de las selvas altas perennifolias. In: Gómez-Pompa A & Del-Amo S, eds. *Investigaciones sobre la Regeneración de selvas altas en Veracruz, México*. Mexico City: Alhambra Mexicana, 191-240.

Martínez-Ramos M & García-Orth X. 2007. Sucesión ecológica y restauración de las selvas húmedas. *Boletín de la Sociedad Botánica de México* **80 (Suplemento)**:69-84.

McGill BJ, Enquist BJ, Weiher E, Westoby M. 2006. Rebuilding community ecology from functional traits. *Trends in Ecology & Evolution* **21**:178-185. DOI: 10.1016/j.tree.2006.02.002

Morrison EB & Lindell CA. 2011. Active or passive forest restoration? Assessing Restoration alternatives with avian foraging behavior. *Restoration Ecology* **19**:170-177. DOI: 10.1111/j.1526-100X.2010.00725.x

Paine CET, Amissah L, Auge H, Baraloto C, Baruffol M, Bourland N, Bruelheide H, Daïnou K, de Gouvenain RC, Doucet J-L, Doust S, Fine PVA, Fortunel C, Haase J, Holl KD, Jactel H, Li X, Kitajima K, Koricheva J, Martínez-Garza C, Messier C, Paquette A, Philipson C, Piotto D, Poorter L, Posada JM, Potvin C, Rainio K, Russo SE, Ruiz-Jaen M, Scherer-Lorenzen M, Webb CO, Wright SJ, Zahawi RA, Hector A. 2015. Globally, functional traits are weak predictors of juvenile tree growth, and we do not know why. *Journal of Ecology* **103**:978-989. DOI: 10.1111/1365-2745.12401

Popma J, Bongers F, Werger JA. 1992. Gap-dependence and leaf characteristics of trees in a tropical lowland rain forest in Mexico. *Oikos* **63**:207-214. DOI: 3545380

Poorter L, Wright SJ, Paz H, Ackerly DD, Condit R, Ibarra-Manríquez G, Harms KE, Licona JC, Martínez-Ramos M, Mazer SJ, Muller-Landau HC, Peña-Claros M, Webb CO, Wright IJ. 2008. Are functional traits good predictors of demographic rates? Evidence from five Neotropical forests. *Ecology* **89**: 1908-1920. DOI: 10.1890/07-0207.1

Román-Dañobeytia FJ, Levy-Tacher S, Aronson J, Ribeiro-Rodrigues R, Castellanos-Albores J. 2012. Testing the performance of fourteen native tropical tree species in two abandoned pastures of the Lacandon rainforest region of Chiapas, Mexico. *Restoration Ecology* **20**: 378-386. DOI: 10.1111/j.1526-100X.2011.00779.x

Saldaña-Acosta A, Meave JA, Sanchez-Velásquez LR. 2009. Seedling biomass allocation and vital rates of cloud forest tree species: Responses to light in shade house conditions. *Forest Ecology and Management* **258**:1650-1659. DOI: 10.1016/j.foreco.2009.07.027

SER /Society for Ecological Restoration International Science & Policy Working Group). 2004. *The SER International Primer on Ecological Restoration*. www.ser.org & Tucson: Society for Ecological Restoration International. <www.ser.org/resources/resources-detail-view/ser-international-primer-on-ecological-restoration> (accessed September, 17, 2015)

Slik JWF, Arroyo-Rodriguez V, Aiba S, Alvarez-Loayza P, Alves LF, Ashton P, Balvanera P, Bastiangan ML, Bellingham PJ, van den Berg E, Bernacci L, Bispo PD, Blanc L, Bohning-Gaese K, Boeckx P, Bongers F, Boyle B, Bradford M, Brearley FQ, Hockemba MBN, Bunyavejchewin S, Matos DCL, Castillo-Santiago M, Catharino ELM, Chai SL, Chen YK, Colwell RK, Robin CL, Clark C, Clark DB, Clark DA, Culmsee H, Damas K, Dattaraja HS, Dauby G, Davidar P, DeWalt SJ, Doucet JL, Duque A, Durigan G, Eichhorn KAO, Eisenlohr PV, Eler E, Ewango C, Farwig N, Feeley KJ, Ferreira L, Field R, de Oliveira AT, Fletcher C, Forshed O, Franco G, Fredriksson G, Gillespie T, Gillet JF, Amarnath G, Griffith DM, Grogan J, Gunatilleke N, Harris D, Harrison R, Hector A, Homeier J, Imai N, Itoh A, Jansen PA, Joly CA, de Jong BHJ, Kartawinata K, Kearsley E, Kelly DL, Kenfack D, Kessler M, Kitayama K, Kooyman R, Larney E, Laumonier Y, Laurance S, Laurance WF, Lawes MJ, do Amaral IL, Letche SG, Lindsell J, Lu XH, Mansor A, Marjokorpi A, Martin EH, Meilby H, Melo FPL, Metcalfe DJ, Medjibe VP, Metzger JP, Millet J, Mohandass D, Montero JC, Valeriano MD, Mugerwa B, Nagamasu H, Nilus R, Ochoa-Gaona S, Onrizal, Page N, Parolin P, Parren M, Parthasarathy N, Paudel E, Permana A, Piedade MTF, Pitman NCA, Poorter L, Poulsen AD, Poulsen J, Powers J, Prasad RC, Puyravaud JP, Razafimahaimodiso JC, Reitsma J, dos Santos JR, Spironello WR, Romero-Saltos H, Rovero F, Rozak AH, Ruokolainen K, Rutishauser E, Saiter F, Saner P, Santos BA, Santos F, Sarker SK, Sardichanh M, Schmitt CB, Schongart J, Schulze M, Suganuma MS, Sheil D, Pinheiro ED, Sist P, Stevart T, Sukumar R, Sun IF, Sunderland T, Suresh HS, Suzuki E, Tabarelli M, Tang JW, Targhetta N, Theilade I, Thomas DW, Tchouto P, Hurtado J, Valencia R, van Valkenburg J, Do TV, Vasquez R, Verbeeck H, Adekunle V, Vieira SA, Webb CO, Whitfeld T, Wich SA, Williams J, Wittmann F, Woll H, Yang XB, Yao CYA, Yap SL, Yoneda T, Zahawi RA, Zakaria R, Zang RG, de Assis RL, Luize BG, Venticinque EM. 2015. An estimate of the number of tropical tree species. *Proceedings of the National Academy of Sciences of the United States of America* **112**:7472-7477. DOI: 10.1073/pnas.1423147112

South DB & Zwolinski JB. 1997. Transplant Stress Index: a proposed method of quantifying planting check. *New Forests* **13**: 315-328. DOI: 10.1023/A:1006546627342

Swaine MD & Whitmore TC. 1988. On the definition of ecological species groups in tropical rain forests. *Vegetatio* **75**:81-86. DOI: 10.1007/BF00044629

Tucker NIJ & Murphy TM. 1997. The effects of ecological rehabilitation on vegetation recruitment: some observations from the Wet Tropics of North Queensland. *Forest Ecology and Management* **99**:133-152. DOI: 10.1016/S0378-1127(97)00200-4

Uhl C. 1987. Factors controlling succession following slash-and-burn agriculture in Amazonia. *Journal of Ecology* **75**:377-407. DOI: 2260425

Received:
October 21st, 2015

Accepted:
February 11th, 2016

Vázquez-Yanes C, Batís AIM, Alcocer MIS, Gual MD, Sánchez DC. 1999. Árboles y arbustos nativos potencialmente valiosos para la restauración ecológica y la reforestación. Project J-084-CONABIO. Mexico City.

Violle C, Navas ML, Vile D, Kazakou E, Fortunel C, Hummel I, Garnier E. 2007. Let the concept trait be functional! *Oikos* **116**:882-892. DOI: 10.1111/j.0030-1299.2007.15559.x

Weiher E, van der Werf A, Thompson K, Roderick M, Garnier E, Eriksson O. 1999. Challenging Theophrastus: A common core list of plant traits for functional ecology. *Journal of Vegetation Science* **10**:609-620. DOI: 10.2307/3237076

Westoby M. 1998. A leaf-height-seed (LHS) plant ecology strategy scheme. *Plant and soil* **199**:213-227. DOI: 10.1023/A:1004327224729

Whitmore TC. 1989. Canopy gaps and the two major groups of forest trees. *Ecology* **70**:536-38. DOI: 1940195

Zahawi RA, Holl KD, Cole RJ, Reid JL. 2013. Testing applied nucleation as a strategy to facilitate tropical forest recovery. *Journal of Applied Ecology* **50**:88-96. DOI: 10.1111/1365-2664.12014

Appendix 1. Sixteen functional traits of recruits (R) and/or transplants (T) of 15 native tropical tree species in restoration settings at Los Tuxtlas, Veracruz, Mexico. Functional leaf traits: Leaf Area (cm²), Leaf Mass per unit area (LMA; g/m²) Leaf dry matter content (LDMC; g/g); Architecture traits: Crown Area (m²), Crown Length(cm), Height to the First Branch (Height FB; cm), Maximal tree height (Max tree height; m); Reproductive traits: Flower size (Flower S; mm), Fruit size (Fruit S; mm), Fruit Weight (Fruit W; mg), Seed Weight (Seed W; mg) and Seed Number per Fruit (Seed Number).

Species	Leaf Area	LMA	LDMC	Crown Area	Crown Length	Height 1st b	Adult height	Flower size	Fruit size	Fruit Weight	Seed Weight	Seed Number
<i>Albizia</i> R	1.00	0.57	0.003	23.82	110.36	102.73	25	3.50	110.00	770.00	30.00	10
<i>Albizia</i> T	1.00	0.57	0.003	11.46	156.57	112.70	25	3.50	110.00	770.00	30.00	10
<i>Amphitecna</i>	26.12	109.45	0.29	27.98	91.18	51.88	10	70.00	150.00	21,000	600.00	50
<i>Bernoullia</i>	55.33	69.14	0.23	4.62	76.93	79.53	25	12.00	300.00	100,000	330.00	150
<i>Brosimum</i>	52.15	99.06	0.32	3.58	118.08	38.85	40	0.60	22.00	1,950.00	910.00	3
<i>Bursera</i>	50.14	46.06	0.23	14.62	169.03	136.12	40	2.50	12.50	250.00	110.00	1
<i>Cecropia</i> R	1,453.10	80.48	11.69	70.33	270.00	420.84	35	1.80	250.00	3,800.00	0.90	4,758
<i>Cecropia</i> T	1,468.00	84.11	0.27	68.14	229.72	386.51	35	1.80	250.00	3,800.00	0.90	4,758
<i>Cedrela</i> R	40.50	55.75	0.29	67.28	110.46	74.40	30	3.00	300.00	5,450.00	10.00	30
<i>Cedrela</i> T	40.50	55.75	0.29	210.10	499.93	82.37	30	3.00	300.00	5,450.00	10.00	30
<i>Cojoba</i>	0.36	41.13	0.56	37.48	264.34	47.92	25	10.00	260.00	1,200.00	100.00	12
<i>Eupatorium</i>	99.61	64.48	0.64	37.57	333.72	30.07	5	2.80	3.50	0.90	0.90	1
<i>Ficus</i>	117.32	85.51	0.25	55.75	311.94	36.81	35	110.00	18.00	7,600.00	1.90	4,000
<i>Helicocarpus</i> R	134.22	42.13	0.56	80.17	499.93	58.88	30	4.00	12.00	5.00	1.50	1
<i>Helicocarpus</i> T	126.46	63.39	0.24	123.26	123.26	451.14	30	4.00	12.00	5.00	1.50	1
<i>Ochroma</i>	458.28	63.39	0.24	68.18	328.80	261.15	30	170	200	30	7.00	800
<i>Tabebuia</i>	58.45	64.39	0.34	5.94	107.11	62.40	30	50.00	300.00	500.00	1.00	500
<i>Tetrorchidium</i>	38.70	97.45	0.37	29.28	292.14	175.00	20	1.50	7.00	160.00	48.00	2
<i>Trema</i>	11.53	63.77	7.35	59.17	306.00	57.28	15	2.20	2.70	12.00	3.00	1

Appendix 2. Survival (%), Height Growth Rate (Height GR; cm/month), Diameter Growth Rate (Diameter GR; mm/month), and number of recruits between 2007 and 2011 for 15 tropical tree species grouped in three categories based in life history, functional traits, performance and number of recruits in restoration settings at Los Tuxtlas, Veracruz, Mexico. Averages (\pm standard error) growth rates were calculated with a subsample of individuals to balance number of individuals among species and between seedlings origins; species are referred by genus.

Species	Survival	Height GR	Diameter GR	Recruits
Good recruiters				
<i>Bursera</i>	69.84	6.43 \pm 1.13	1.31 \pm 1.13	337
<i>Albizia</i> *	83.00	1.86 \pm 1.50	1.01 \pm 1.39	57
<i>Eupatorium</i>	88.88	7.70 \pm 1.05	2.54 \pm 1.18	53
<i>Cecropia</i> *	90.97	14.08 \pm 1.10	2.13 \pm 1.18	36
<i>Helicarpus</i> *	78.12	9.83 \pm 1.07	2.67 \pm 1.37	23
<i>Tetrorchidium</i>	33.33	6.09 \pm 1.15	2.42 \pm 1.09	20
<i>Trema</i>	94.11	9.63 \pm 1.11	9.63 \pm 1.11	13
<i>Cedrela</i> *	70.98	2.90 \pm 1.27	1.04 \pm 1.21	13
Good transplants				
<i>Ochroma</i>	32.29	9.25 \pm 1.19	7.47 \pm 1.23	1
<i>Ficus</i>	73.95	7.56 \pm 1.14	3.70 \pm 1.23	0
<i>Cojoba</i>	72.91	8.33 \pm 1.09	2.62 \pm 1.09	0
<i>Tabebuia</i>	54.16	4.44 \pm 1.24	0.85 \pm 1.29	0
Poor Transplants				
<i>Amphitecna</i>	39.36	2.63 \pm 1.09	0.48 \pm 1.23	0
<i>Bernoullia</i>	35.41	3.43 \pm 1.31	1.18 \pm 1.24	0
<i>Brosimum</i>	15.62	6.46 \pm 1.14	0.45 \pm 1.18	0

* Average performance for recruits and transplant are given for species with individuals from both origins