ANALYSIS OF THE α-CYPERMETHRIN BIODEGRADATION BY Klebsiella pneumoniae YH43

Análisis de la biodegradación de α-cipermetrina por Klebsiella pneumoniae YH43

Verónica HERRERA-MAYORGA¹, Diana Laura OCHOA-ESPARZA², Timoteo DELGADO-MALDONADO³, Alma D. PAZ-GONZÁLEZ³, Francisco PAREDES-SÁNCHEZ¹ and Gildardo RIVERA³*

¹ Universidad Autónoma de Tamaulipas, Blvd. Enrique Cárdenas González 1201 Poniente, Col. Jardín, 89840 El Mante, Tamaulipas, México.

² Instituto Politécnico Nacional, Circuito del Gato 202, Ciudad Administrativa, 98160 Zacatecas, Zacatecas, México.

³ Instituto Politécnico Nacional, Centro de Biotecnología Genómica, Laboratorio de Biotecnología Farmacéutica, Blvd. del Maestro s/n, Col. Narciso Mendoza, 88710 Reynosa, Tamaulipas, México.

*Author for correspondence: gildardors@hotmail.com

(Received: June 2021; accepted: September 2022)

Key words: pyrethroid, bacterium, biodegradation, α -cyano-3-phenoxybenzylalcohol.

ABSTRACT

Cypermethrin (CYP) is an insecticide pyrethroid used in the agriculture for its broadspectrum activity. However, recent studies have shown the accumulation of CYP in soil, and water and a high toxicity against different species, including humans. On the other hand, the use of bacteria in bioremediation represents an eco-friendly alternative for the removal of different xenobiotics. In this study, the aim was to isolate and identify bacteria from agricultural soils to determine their ability to tolerate and biodegrade CYP. Additionally, some metabolites of the CYP biodegradation were detected by ultra-performance liquid chromatography with mass spectrometry (UPLC-MS). *Klebsiella pneumoniae* YH43 isolate used CYP as a carbon source, degrading 87.55 % at a concentration of 100 mg/L. α -cyano-3-phenoxybenzyl alcohol and *cis,trans*-3-(2,2dichlorovinyl)-2,2-dimethylcyclopropane carboxylate were identified as the main metabolites in the pathway of the CYP biodegradation. These results suggest that *K. pneumoniae* YH43 has the potential to be used in the bioremediation of contaminated soils with CYP.

Palabras clave: piretroide, bacteria, biotransformación, alcohol α-ciano-3-fenoxibencil.

RESUMEN

La cipermetrina (CYP) es un insecticida piretroide utilizado en la agricultura por su actividad de amplio espectro. Sin embargo, estudios recientes han evidenciado la acumulación de CYP en suelo y agua y una alta toxicidad en diferentes especies, incluido el ser humano. Por otra parte, el uso de bacterias en la biorremediación representa una alternativa sostenible para la remoción de distintos xenobióticos. En este estudio, el objetivo fue aislar e identificar bacterias de suelos agrícolas con habilidad para tolerar y biodegradar CYP. Adicionalmente, varios metabolitos de la degradación de CYP fueron identificados por cromatografía de líquidos de ultra alta resolución con espectrometría de masas (UPLC-MS, por su sigla en inglés). *Klebsiella pneumoniae* YH43 usó a CYP como fuente de carbono, causando una degradación del 87.55 % a una concentración de 100 mg/L. El alcohol α -ciano-3-fenoxibencil y el carboxilato de *cis,trans*-3-(2,2-diclorovinil)-2,2-dimetilciclopropano fueron los principales metabolitos identificados en la ruta de degradación de CYP. Los resultados sugieren que *Klebsiella pneumoniae* YH43 tiene potencial para ser utilizada en la biorremediación de suelos contaminados con CYP.

INTRODUCTION

In the last two decades, the world population increased by 25.7 %, from 6.2 to 7.7 billion (GBD 2020), which caused an increase in a good quality and safe food production (Fung et al. 2018). In this sense, the intensive use of pesticides in agriculture has made it possible to satisfy the population's food needs (Carvalho 2017). Pyrethroid insecticides stand-out because of their broad-spectrum activity towards the Lepidoptera order, their moderate persistence under field conditions, high potency, and cost-effectiveness (Weston et al. 2009, 2011, Pankaj et al. 2016).

Pyrethroids are synthetic insecticides derived from pyrethrin (derived from Chrysanthemum plants). This kind of insecticides has been used for more than 40 years and represent more than 30 % of the global insecticide market (Pietrantonio et al. 2007, Yao et al. 2015, Kalita et al. 2017). Cypermethrin (CYP) is a pyrethroid insecticide used worldwide in many crops due to its high efficacy (Tang et al. 2018). However, several studies have shown its acute toxicity to fish, invertebrates, and humans; therefore, their intensive use negatively impacts the environment and human health (Jin et al. 2011, Hocine et al. 2016, Ullah et al. 2018). The toxicity produced by CYP has been attributed mainly to two degradation metabolites, cis-3-(2',2'-dichlorovinyl)-2,2-dimethylcyclopropane carboxylic acid (cis-DCCA), and 3-phenoxybenzoic acid (3-PBA) (Yao et al. 2015). Different strategies have been implemented in agriculture to reduce the impact of insecticides, such as integrated pest management (IPM), reduction in the use of insecticides, the use of genetically modified crops (GE), the creation and use of biopesticides, and the implementation of precision agriculture (Soberón et al. 2016). So far, bioremediation has been proposed as the most complete method of biodegradation of insecticides to restore the environment (Asim et al. 2021).

Bioremediation is considered an eco-friendly and cost-effective procedure for removing xenobiotics

(polycyclic aromatic hydrocarbons, pesticides, chlorinated compounds, phenols, and others) of contaminated environments (Mao and Guan 2016, Mishra et al. 2021). Particularly, native microorganisms of contaminated environments represent an important source of bioremediation agents. Several studies aim to find genera of soil-derived bacteria with insecticide-degrading activity (Cycoń and Piotrowska-Seget 2016). Constant exposure to insecticides has allowed certain bacteria to develop the ability to use them as a substrate, generating less toxic or more assimilable metabolites that mitigate the environmental impact (Cycoń et al. 2009, Akbar and Sultan 2016). Bacteria such as strain Klebsiella sp. ZD112 (Wu et al. 2006), Serratia sp. JCN13, and Pseudomonas aeruginosa CH7 (Zhang et al. 2010, 2011), and Bacillus subtilis (Gangola et al. 2018) can degrade CYP through different mechanisms of action, mainly by hydrolytic enzymes (Fig. 1).

In this study, the main objective was to isolate and identify new microorganisms capable of using CYP as a carbon or nitrogen source for growth, which can be used in CYP remotion. Bacteria were isolated from agricultural soil samples and their ability to biodegrade CYP in liquid media was determined. Main metabolites generated by the CYP degradation using strain *Klebsiella pneumoniae* YH43 were identified as biotransformation products.

MATERIALS AND METHODS

Sample collection

The sample collection site was in Fresnillo City, Zacatecas, Mexico (23° 12' N-103° 30' W). The soils are used for bean and corn crops, and CYP ([cyano-(3phenoxyphenyl)methyl]-3-(2,2-dichloroethenyl)-2,2dimethylcyclopropane-1-carboxylate) is periodically applied. Approximately one month after the last application of CYP, a total of 10 different soil samples were collected from a depth of 0-20 cm.



Fig. 1. Biodegradation of cypermethrin by *B. subtilis* and *Klebsiella* sp. Production of (A) 3-(2,2-dichloroethenyl)-2,2-dimethylcyclopropanecarboxylate, (B) 3-(2,2-dichlorovinyl)-2,2-dimethylcyclopropanecarboxylic acid, and (C) and 2-hydroxy-2-(3-phenoxyphenyl) acetonitrile.

Subsequently, the samples were mixed and homogenized (bean and corn crop soil, respectively) to obtain two composite samples (Ammari et al. 2015).

The samples were then transferred in airtight plastic bags to the Laboratorio de Biología of the Unidad Profesional Interdisciplinaria de Ingeniería, Instituto Politécnico Nacional Campus Zacatecas (Laboratory of Biology, Interdisciplinary Professional Engineering Unit, Zacatecas Campus of the National Polytechnic Institute) (UPIIZ-IPN).

Isolation of bacteria

One gram of the composite sample soil from each culture was used and dissolved in 20 mL of minimal saline medium (MSM). The composition of the MSM was g/L: 0.5 g (NH₄)₂SO₄, 0.3 g K₂HPO₄, 0.01 g KCl, 0.01 g NaCl, 0.8 g dextrose, and 100 mg/L CYP \geq 98% from Sigma-Aldrich. The sample was incubated for 24 h at 35 °C and 120 rpm. Samples were assayed by triplicate under sterile conditions.

After the incubation period, plates with nutrient agar (MCD) supplemented with CYP (100 mg/L) were inoculated by rod extension with the previously obtained culture. The plates were incubated in static conditions at 35 °C for 24 h. The colonies were characterized morphologically with an optical microscope taking into consideration elevation, edge, shape, surface, and color. The colonies with different characteristics were selected and isolated.

Molecular identification

Genomic DNA extraction was performed on individually identified colonies. The Wizard Genomic DNA purification kit (Promega A1120, USA) was used according to the protocol described by the manufacturer. The 16s rDNA region was amplified by end-point PCR, using the Bac1-FW 5'-AGAGA-GTTTGATCVTGGCTCAG-3' and 16S-1400 RV 5'-GCGGTGTGTGTGTACAAGGCCCG-3' oligo-nucleotides reported by Criollo et al. (2012). The reaction was adjusted to a final volume of 25 μ L: 50 ng/ μ L gDNA, 1 X DNA polymerase Taq Buffer, 2.5 mM MgCl₂, 0.2 mM dNTPs, 0.2 μ M BAC-1FW, and 16S-1400 RV, 0.125 U/ μ L GoTaq Polymerase. Amplification conditions were 94 °C for 3 min, followed by 30 cycles of denaturation at 94 °C for 30 s, alignment at 57 °C for 30 s, extension at 72 °C for 30 s, and a final extension at 72 °C for 3 min in a Labnet Multigene TM mini thermal cycler.

The PCR products were purified following the ExoSAP-IT protocol (Affymetrix, Santa Clara, CA) and sequenced according to the conditions indicated in the BigDye Terminator v.3.1 cycle sequencing kit of the ABI 3130 system (Applied Biosystems, Foster City, CA). The electropherogram was visualized, edited, and assembled with Chromas Lite 2.1 (Technelysium) and SeqMan software from the Lasergene suite DNASTAR 8 (Madison, WI). The assembled nucleotide sequences were compared with the NCBI nr/nt database using the BLASTn tool to identify selected ClustalW bacteria (homology > 99 %; Koolivand et al. 2019). A phylogenetic tree was constructed with the sequences obtained using the ClustalW method (Tamura et al. 2007).

Growth in the presence of CYP as a carbon source

All assays were performed by triplicate under sterile conditions. Erlenmeyer flasks containing 20

mL of MSM supplemented with CYP (100 mg/L) were inoculated with 3×10^6 cells/mL of each selected bacterium. Flasks were incubated at 35 °C and 120 rpm in a ZHWY 200B incubator (Zhicheng). Aliquots of 5 mL were drawn at different times (0, 6, 12, 24, 24, 36, 48, 56, 72, 96, 120, 140, 180, and 360 h) and placed in plastic microwells to determine the optical density at 600 nm (OD₆₀₀ nm) in a Thermo Scientific Genesys 10S UV-Vis spectrophotometer. Three flasks were introduced as controls: 1 (MSM), 2 (MSM with inoculum and without CYP), and 3 (MSM only with CYP).

CYP biodegradation assay

For the CYP biodegradation assays, the bacterium that previously showed the best growth in the presence of CYP was selected. The bacterium was reactivated under the above-mentioned conditions. From this seed flask, an inoculum of 3×10^{6} cells/mL was collected by triplicate and added to a flask containing 30 mL of MSM supplemented with CYP (100 mg/L). The culture was monitored, and aliquots of 1.0 mL were drawn at different times (0, 6, 12, 24, 36, 48, 56, 72, 96, 120, 140, 180, and 360 h) to determine the percentage of CYP biodegradation. Biodegradation was calculated from a standard curve with different CYP concentrations (5, 20, 30, 40, 50, 60, 80, and 100 mg/L). The results were read by triplicate in a spectrophotometer (UV-Vis Thermo Scientific Genesys 10S) at a $\lambda = 210$ nm and plotted by simple linear regression.

Degradation products extraction

Thirteen samples were obtained by triplicate from the different times, placed in Falcon tubes (Corning), and vortexed for 3 min. Subsequently, they were centrifuged at 14 000 rpm for 30 min and the supernatant was recovered. Metabolites were extracted by the liquid-liquid technique with ethyl acetate 2:1 (v/v). The organic phase was recovered, and the solvent was removed under reduced pressure at 40 °C. The obtained product was resuspended in 5 mL of methanol. This solution was filtered with a 0.22 μ m pore membrane (Millex GP, Millipore Express PES membrane). The filtrate was used for substrate analysis by ultra-performance liquid chromatography-mass spectrometry (UPLC-MS).

Analysis by UPLC-MS

Samples were analyzed at the Centro de Biotecnología Genómica of the Instituto Politécnico Nacional (Genomic Biotechnology Center of the National Polytechnic Institute). The UPLC-MS analysis was performed using an Acquity UPLC system coupled to a Waters QDA mass detector column (Milford, MA): Acquity UPLC Cortecs C₁₈ 1.6 μ m, 3.0 × 100 mm. The injection volume was 1 μ L at a flow rate of 0.5 mL/min. The column temperature was maintained at 40 °C during the analysis, using a gradient elution with a mobile phase of acetonitrile and 0.1 % formic acid in water at a ratio of 30:70. The total run time for samples with CYP was 8 min. Data acquisition, data management, and instrument control were performed using Empower software.

RESULTS

Isolation and identification

In this work, bacteria from soil samples of bean and corn crops from Fresnillo City, Zacatecas, Mexico were obtained. Six bacterial strains were morphologically identified from a primary culture (soil inoculated in MSM with CYP) and subsequently transferred to an enriched medium supplemented with 100 mg/L of CYP. Two strains were isolated from corn soil and four from bean soil. Molecular identification by 16S rDNA sequencing showed that the bacteria belonged to the genera *Acinetobacter*, *Serratia, Pseudomonas*, and *Klebsiella* (**Table I**), in particular two strains of the genus *Klebsiella* and two strains of the genus *Pseudomonas*.

 TABLE I. BACTERIA ISOLATED FROM CORN AND BEAN CROP SOILS

 WITH ADDED CYPERMETHRIN MEDIA AND IDENTIFIED BY

 16S RDNA SEQUENCING.

Code	Crop	Bacteria	Access number
FP2	Bean	Acinetobacter calcoaceticus	AB680365.1
FP1	Bean	Klebsiella pneumoniae (YH43)	AP014950.1
MP2	Corn	Klebsiella pneumoniae (KP-1)	CP012883.1
FP4	Bean	Pseudomonas plecoglossicida (NyZ12)	KY646089.1
MP4	Corn	Pseudomonas putida (n7)	JQ782496.1
FP5	Bean	Serratia marcescens	KX928068.1



Fig. 2. Growth at 180 h of isolated bacterial strains using cypermethrin (100 mg/L) as a carbon source.

Growth in the presence of CYP as a carbon source

To select the most tolerant strain with the best growth, the six identified strains were inoculated in MSM with CYP at 100 mg/L as the maximum concentration. **Figure 2** shows the growth behavior of each bacterium after 180 h of culture. The best adapted genus was *Klebsiella*. *K. pneumoniae* YH43 showed the best adaptation using CYP as the only carbon source, with a greater exponential growth in the experiment. *K pneumoniae* KP-1 and *S. marcescens* spp. also showed an exponential growth, but lower than the latter. On the contrary, *P. plecoglossicida* NyZ12, *A. calcoaceticus*, and *P. putida* n7 showed a lower growth in all the evaluated times.

CYP degradation analysis

The percentage of CYP degradation by *K. pneu-moniae* YH43 in a liquid medium (pH 7.0) at 35 °C was calculated from the equation of the straight line generated from the standard curve (**Fig. 3**). The OD of

Fig. 3. Cypermethrin standard curve obtained by ultraviolet visible (UV-Vis) spectrophotometry at $\lambda = 210$ nm.

each flask with MSM and CYP (100 mg/L) determined at different times, was substituted into the equation to determine changes in CYP concentration. CYP degradation by *K. pneumoniae* YH43 was 87.55 % at 360 h of culture.

UPLC-MS analysis

Finally, an UPLC-MS analysis was done to know the potential metabolites of the CYP degradation by K. pneumoniae YH43. Initially, the CYP standard at time 0 h was identified at a time of 2.94 min with a molecular weight of 401.21 (Fig. 4A) and different metabolites previously reported were detected (Table II). After 360 h, the metabolites detected were methyl-3-(2,2-dichlorovinyl)-2-2-dimethylcyclopropane-1-carboxylate; methyl-3-(2,2-dichlorovinyl)-2-(hydroxymethyl)-2-methylcyclopropane-1-carboxylate; 3-(2,2-dichlorovinyl)-2-hydroxymethyl)-2-methylcyclopropane-1-carboxylic acid; methyl-3-(2,2-dichlorovinyl)-2-(hydroxymethyl)-2-methylcyclopropane-1-carboxylate; 3-phenoxybenzyl alcohol, and 3-phenoxybenzoic acid (3-PBA; Fig. 4B and 5).

DISCUSSION

In this work, six bacteria from pesticide contaminated soil were identified as agents capable to tolerate CYP. The presence of CYP in the medium caused a selective effect on the growth of the bacteria. Our findings are consistent with previous studies of strains resistant to pesticides isolated from soil and water samples contaminated from areas of high agricultural activity (Huang et al. 2018).

Commonly, *K. pneumoniae* is one of the most dangerous species to health, with high virulence and antibiotic resistance. In this study, the 16S rDNA gene



Fig. 4. Ultra-performance liquid chromatography-mass spectrometry (UPLC-MS) analysis. (A) Chromatogram of cypermethrin at time 0 h; (B) metabolites produced by *K. pneumoniae* YH43 after 360 h. a: methyl-3-(2,2-dichlorovinyl)-2-2-dimethyl-cyclopropane-1-carboxylate; b: methyl-3-(2,2-dichlorovinyl)-2-(hydroxymethyl)-2-methylciclopropane-1-carboxilate; c: 3-(2,2-dichlorovinyl)-2-hydroxymethyl)-2-methylcyclopropane-1-carboxylic acid; d: methyl-3-(2,2-dichlorovinyl)-2-(hydroxymethyl)-2-methylciclopropane-1-carboxilate; e: 3-phenoxybenzyl alcohol; and f: 3-phenoxibenzoic acid (3-PBA).

TABLE II.	METABOLITES IDENTIFIED BY ULTRA-PERFORMANCE LIQU	JID CHROMATOGRAPHY-MASS SPECTROMETRY
	UPLC-MS) OF α-CYPERMETHRIN (100 mg/L) DEGRADATIC	N BY K. pneumoniae YH43.

	Mw reported m/z	Time 96 h		Time 360 h		
Metabolites		Mw founded m/z	Rt (min)	Mw founded m/z	Rt (min)	⁻ % Identity
Cypermetrine standard	416.30	401.21	2.938	401.22	2.930	100
Alfa-cypermetrine	416.3	415.13	1.36	415.16	1.367	98
methylcyclopropane-1-carboxylic acid (<i>cis</i> -DCCA)	209.07	ND	ND	ND	ND	ND
3-phenoxibenzoic acid (3-PBA)	214.22	214.26	1.578	214.31	1.550	99
I cypermetrine	430.41	432	1.334	432	0.710	96
3-phenoxybenzyl-3-(2,2-dichlorovinyl)-2,2- dimethylcyclopropane-1-carboxylate	404.39	ND	ND	402	0.760	96
2-(aminooxy)-2-oxo-1-(3-phenoxyphenyl) ethyl- 3-(2,2-dichlorovinyl)-2,2-dimethylciclopropane-1- carboxilate	449.43	449.73	5.553	ND	ND	99
2-(methoxy)-2-oxo-1-(3-phenoxyphenyl) ethyl-3- (2,2-dichlorovinyl)-2,2-dimethylcyclopropane-1- carboxylate	464.44	464.35	0.710	464	0.660	99
Cyano (3-(4-methoxyphenoxy) phenyl) methyl-3- (2,2-dichlorovinyl)-2,2-dimethylcyclopropane-1 carboxylate	462.45	462	4.398	462	3.390	99
(3- phenoxyphenyl) methanol	210.31	ND	ND	209	2.93	98
Methyl-3-pheenoxybenzoate	238.32	238.04	1.31	238.03	1.295	99
3-(4-hydroxyphenoxy) benzoic acid	240.29	241.30	0.651	ND	ND	98
Methyl-3-(4-hydroxyphenoxy) benzoate	255.33	256.34	3.4	256	3.355	98
Methyl-3-(4-methoxiphenoxy) benzoate	270.36	269.26	2.30	269.15	2.44	98

Mw: molecular weight, Rt: retention time, ND: not detected.

	Mw reported m/z	Time 96 h		Time 360 h		
Metabolites		Mw founded m/z	Rt (min)	Mw founded m/z	Rt (min)	[—] % Identity
3-(2,2-dichlorovinyl)2,2-dimethylcyclopropane-1- carboxylic acid	215.11	214.26	1.626	214.31	1.557	98
Methyl-3-(2,2-dichlorovinyl)-2-2- dimethylcyclopropane-1-carboxylate	230.15	230.26	1.595	230.25	1.35	99
3-(2,2-dichlorovinyl)-1-methylcyclopropane-1,2- dicarboxylic acid	245.10	247.62	1.373	247.60	1.367	96
Dimethyl-3-(2,2-dichlorovinyl)-1- methylcyclopropane-1,2-dicarboxylate	275.17	ND	ND	275.07	0.833	99
3-(2,2-dichlorovinyl)-2-hydroxymethyl)-2- methylcyclopropane-1-carboxylic acid	232.12	ND	ND	233.12	1.343	98
Methyl-3-(2,2-dichlorovinyl)-2-(hydroxymethyl)-2- methylciclopropane-1-carboxilate	247.16	247.15	2.456	247.20	2.461	99
Cyano (3-(4-hydroxyphenoxy) phenyl) methyl-3- (2,2-dichlorovinyl)-2,2-dimethylcyclopropane-1 carboxilate	447.42	ND	ND	447	0.789	99
3-phenoxybenzaldehyde	198.26	ND	ND	ND	ND	ND
Dichloroethenyl, dimethylcyclopropane carboxylic acid	104.53	ND	ND	ND	ND	ND
3-phenoxybenzyl alcohol	199.0	ND	ND	200	1.462	98
(RS)- alpha-cyano-3-phenoxybenzyl alcohol	225.5	225.21	1.140	225	1.137	99
(1RS)- cis, trans-3- (2,2-dimethylcyclopropanecarb oxylate	208.06	ND	ND	208	6.991	99
3-phenyl benzoate	213.21	ND	ND	214.22	1.557	98
3,4-Dihydroxybenzoate	153.11	150.02	0.675	150.01	0.678	96

TABLE II. METABOLITES IDENTIFIED BY ULTRA-PERFORMANCE LIQUID CHROMATOGRAPHY-MASS SPECTROMETRY (UPLC-MS) OF α-CYPERMETHRIN (100 mg/L) DEGRADATION BY *K. pneumoniae* YH43.

Mw: molecular weight, Rt: retention time, ND: not detected.



Fig. 5. Metabolites detected in the pathway of cypermethrin degradation by the *K. pneumoniae* YH43 bacteria. (a) Methyl-3-(2,2-dichlorovinyl)-2-2-dimethylcyclopropane-1-carboxylate; (b) methyl-3-(2,2-dichlorovinyl)-2-(hydroxymethyl)-2-methylciclopropane-1-carboxilate; (c) 3-(2,2-dichlorovinyl)-2-hydroxymethyl)-2-methylcyclopropane-1-carboxylic acid; (d) methyl-3-(2,2-dichlorovinyl)-2-(hydroxymethyl)-2-methylciclopropane-1-carboxilate; (e) 3-phenoxybenzyl alcohol; and (f) 3-pheno-

sequence analysis identified FP1 and MP2 isolated as *K. pneumoniae* strains. Strain *K. pneumoniae* YH43 has been isolated and identified from potato crop soil samples and used as a model to study nitrogen fixation (Iwase et al. 2016). Its presence in this type of source could be due to its adaptation to soil conditions where CYP is commonly applied (Nyawade et al. 2020).

On the other hand, two strains of the genus *Pseudomonas* (*P. plecoglossicida* NyZ12 and *P. putida* n7) were identified. Several studies show that both strains are capable of catabolizing aromatic compounds from industrial chemicals and agrochemicals (Wongsa et al. 2004, Chetverikov et al. 2017, Nogales et al. 2017, Yan et al. 2017). In this study, both strains only tolerated CYP.

A. calcoaceticus was also identified in this study. A previous study showed that A. calcoaceticus can use the insecticide chlorpyrifos as sole carbon source for their growth (Zhao et al. 2014). Additionally, A. calcoaceticus can grow in the presence of polycyclic aromatic hydrocarbons such as fluorene and phenanthrene, as well as heavy metals such as copper and cadmium (Liu et al. 2016, Méndez et al. 2017). However, in this study A. calcoaceticus only tolerated CYP. The isolate FP5, identified as S. marcescens, can degrade compounds such as the herbicide nicosulfuron (Zhang et al. 2012) and hexachlorobutadiene. Therefore, S. marcescens is considered as useful for the bioremediation of contaminated environments (Li et al. 2008). In our study it only tolerated CYP.

K. pneumoniae YH43, K. pneumoniae KP-1, and S. marcescens showed exponential growth using CYP as the only carbon source. The rest of the strains showed atypical growth during the experiment, which suggests that these bacteria are tolerant to concentrations of CYP lower than 100 mg/L and possibly require more culture time to adapt to the CYP concentration tested. Strain K. pneumoniae YH43 showed the best growth; therefore, it was selected for the biodegradation assays. In this work, K. pneumoniae YH43 caused an 87.55 % degradation of CYP, in contrast to the findings of Tang et al. (2019), who reported that under similar conditions strain K. pneumoniae BPBA052 degraded 45.42 % of β -cypermethrin at a concentration of 100 mg/L over a period of 48 h.

According to the metabolites of CYP degradation by strain *K. pneumoniae* YH43 identified by UPLC-MS, the results suggest that the degradation occurred through ester bond hydrolysis, which agrees with the findings of Tallur et al. (2008) for A. calcoaceticus, Sphingomonas sp., and Micrococcus sp. This pathway is considered one of the main routes for CYP degradation (Yao et al. 2015, Huang et al. 2018). Methyl-3-(2,2-dichlorovinyl)-2-2-dimethylcyclopropane-1-carboxylate is a key metabolite in the degradation pathway that causes breaking of the ester bond and loss of insecticidal activity. Additionally, the compounds 3-phenoxybenzyl alcohol and 3-phenoxybenzoic acid (3-PBA) are described as end products of the CYP degradation pathway that can be further mineralized under natural environmental conditions to products such as gallic acid, catechol, and protocatechuic acid (Huang et al. 2018).

CONCLUSION

In this study, *K. pneumoniae* YH43 was isolated and identified as a microorganism capable of growing in MSM using CYP as a carbon source. This strain has the ability to cause an 87.55 % degradation of CYP. Additionally, two compounds were identified as main products of the CYP degradation: α -cyano-3-phenoxybenzyl alcohol and *cis,trans*-3-(2,2dichlorovinyl)-2,2-dimethylcyclopropane carboxylate, suggesting an ester bond hydrolysis. These findings suggest that *K. pneumoniae* YH43 can be used in the bioremediation of CYP contaminated soils.

ACKNOWLEDGMENTS

The authors want to thank Jacqueline Guadalupe Villela Buenrostro and Mayra Andrea de la Torre Ávila of the Programa de Ingeniería Ambiental of the Unidad Profesional Interdisciplinaria de Ingeniería, Campus Zacatecas of the Instituto Politécnico Nacional, for the support in carrying out the microbiological analyzes.

REFERENCES

- Akbar S. and Sultan S. (2016). Soil bacteria showing a potential of chlorpyrifos degradation and plant growth enhancement. Brazilian Journal of Microbiology 47 (3), 563-570. https://doi.org/10.1016/j.bjm.2016.04.009
- Ammari T.G., Tahhan R., Al Sulebi N., Tahboub A., Ta'any R.A. and Abubaker S. (2015). Impact of intensive greenhouse production system on soil quality. Pedosphere 25 (2), 282-293. https://doi.org/10.1016/ S1002-0160(15)60013-3

- Asim N., Hassan M., Shafique F., Ali M., Nayab H., Shafi N., Khawaja S. and Manzoor S. (2021). Characterizations of novel pesticide-degrading bacterial strains from industrial wastes found in the industrial cities of Pakistan and their biodegradation potential. PeerJ 9, e12211. https://doi.org/10.7717/peerj.12211
- Carvalho F.P. (2017). Pesticides, environment, and food safety. Food and Energy Security 6 (2), 48-60. https:// doi.org/10.1002/fes3.108
- Chetverikov S.P., Sharipov D.A., Korshunova T.Y. and Loginov O.N. (2017). Degradation of perfluorooctanyl sulfonate by strain *Pseudomonas plecoglossicida* 2.4-D. Applied Biochemestry and Microbiology 53, 533-538. https://doi.org/10.1134/ S0003683817050027
- Criollo P.J., Obando M., Sánchez L. and Bonilla R. (2013). Efecto de bacterias promotoras de crecimiento vegetal (PGPR) asociadas a *Pennisetum clandestinum* en el altiplano cundiboyacense. Ciencia and Tecnología Agropecuaria 13 (2), 189-195. https://doi.org/10.21930/ rcta.vol13 num2 art:254
- Cycoń M., Wójcik M. and Piotrowska-Seget Z. (2009). Biodegradation of the organophosphorus insecticide diazinon by *Serratia* sp. and *Pseudomonas* sp. and their use in bioremediation of contaminated soil. Chemosphere 76 (4), 494-501. https://doi.org/10.1016/j. chemosphere.2009.03.023
- Cycoń M. and Piotrowska-Seget Z. (2016). Pyrethroiddegrading microorganisms and their potential for the bioremediation of contaminated soils: A review. Front Microbiology 7, 1463. https://doi.org/10.3389/ fmicb.2016.01463
- Fung F., Wang H.S. and Menon S. (2018). Food safety in the 21st century. Biomedical Journal 41 (2), 88-95. https://doi.org/10.1016/j.bj.2018.03.003
- Gangola S., Sharma A., Bhatt P., Khati P. and Chaudhary P. (2018). Presence of esterase and laccase in *Bacillus* subtilis facilitates biodegradation and detoxification of cypermethrin. Scientific Reports 8, 12755. https://doi. org/10.1038/s41598-018-31082-5
- GBD (2020). Global age-sex-specific fertility, mortality, healthy life expectancy (HALE), and population estimates in 204 countries and territories, 1950-2019: A comprehensive demographic analysis for the Global Burden of Disease Study 2019. Lancet 396 (10258), 1160-1203. https://doi.org/10.1016/s0140-6736(20)30977-6
- Hocine L., Merzouk H., Merzouk S.A., Ghorzi H., Youbi M. and Narce M. (2016). The effects of alpha-cypermethrin exposure on biochemical and redox parameters in pregnant rats and their newborns. Pesticide Biochemistry and Physiology 134, 49-54. https://doi. org/10.1016/j.pestbp.2016.04.007

- Huang Y., Xiao L., Li F., Xiao M., Lin D., Long X., and Wu Z. (2018). Microbial degradation of pesticide residues and an emphasis on the degradation of cypermethrin and 3-phenoxy benzoic acid: A review. Molecules 23 (9), 2313. https://doi.org/10.3390/molecules23092313
- Iwase T., Ogura Y., Hayashi T. and Mizunoe Y. (2016). Complete genome sequence of *Klebsiella pneumoniae* YH43. Genome Announcement 4 (2), e00242-16. https://doi.org/10.1128/genomea.00242-16
- Jin Y., Zheng S. and Fu Z. (2011). Embryonic exposure to cypermethrin induces apoptosis and immunotoxicity in zebrafish (*Danio rerio*). Fish and Shellfish Immunology 30 (4-5), 1049-1054. https://doi.org/10.1016/j. fsi.2011.02.001
- Kalita M.K., Haloi K. and Devi D. (2017). Cypermethrin formulation (Ustad-10 EC) induces genotoxicity via apoptosis, affects nutritional physiology, and modulates immune response in silkworm *Philosamia ricini* (Lepidoptera: Saturniidae). Journal of Economy Entomology 110 (3), 1010-1024. https://doi.org/ doi:10.1093/jee/tox044
- Koolivand A., Abtahi H., Godini K., Saeedi R., Rajaei M.S., Parhamfar M. and Seifi H. (2019). Biodegradation of oil tank bottom sludge using a new two-phase composting process: Kinetics and effect of different bulking agents. Journal of Material Cycles and Waste Management 21, 1280-1290. https://doi.org/10.1007/ s10163-019-00881-x
- Li M.T., Hao L.L., Sheng L.X. and Xu J.B. (2008). Identification and degradation characterization of hexachlorobutadiene degrading strain *Serratia marcescens* HL1. Bioresource Technology 99 (15), 6878-6884. https:// doi.org/10.1016/j.biortech.2008.01.048
- Liu Z., Xie W., Li D., Peng Y., Li Z. and Liu S. (2016). Biodegradation of phenol by bacteria strain *Acinetobacter calcoaceticus* PA isolated from phenolic wastewater. International Journal of Environmental Research and Public Health 13 (3), 300. https://doi.org/10.3390/ ijerph13030300
- Mao J. and Guan W. (2016). Fungal degradation of polycyclic aromatic hydrocarbons (PAHs) by *Scopulariopsis brevicaulis* and its application in bioremediation of PAH-contaminated soil. Acta Agriculturae Scandinavica, Section B – Soil and Plant Science 66 (5), 399-405. https://doi.org/10.1080/09064710.201 5.1137629
- Méndez V., Fuentes S., Morgante V., Hernández M., González M., Moore E. and Seeger M. (2017). Novel hydrocarbonoclastic metal-tolerant *Acinetobacter* and *Pseudomonas* strains from Aconcagua river oil-polluted soil. Journal of Soil Science and Plant Nutrition 17 (4), 1074-1087. https://doi.org/10.4067/ S0718-95162017000400017

- Mishra S., Lin Z., Pang S., Zhang W., Bhatt P. and Chen S. (2021). Recent advanced technologies for the characterization of xenobiotic-degrading microorganisms and microbial communities. Frontiers in Bioengineering and Biotechnology 9, 632059. https://doi.org/10.3389/ fbioe.2021.632059
- Nogales J., García J.L. and Díaz E. (2017). Degradation of aromatic compounds in *Pseudomonas*: A systems biology view. In: Aerobic utilization of hydrocarbons, oils and lipids (F. Rojo, Ed.). Springer, Cham, Switzerland AG, pp. 1-49. https://doi.org/10.1007/978-3-319-39782-5 32-1
- Nyawade, S.O., Karanja, N.N., Gachene, C.K.K. Gitari H.I., Schulte-Geldermann E. and Parker M. (2020). Optimizing soil nitrogen balance in a potato cropping system through legume intercropping. Nutrient Cycling in Agroecosystems 117, 43-59. https://doi. org/10.1007/s10705-020-10054-0
- Pankaj B., Sharma A., Gangola S., Khati P., Kumar G. and Srivastava A. (2016). Novel pathway of cypermethrin biodegradation in a *Bacillus* sp. strain SG2 isolated from cypermethrin-contaminated agriculture field. 3 Biotech 6 (1), 45. https://doi.org/10.1007/s13205-016-0372-3
- Pietrantonio P.V., Junek T.A., Parker R., Mott D., Siders K., Troxclair N., Vargas-Camplis J., Westbrook J.K. and Vassiliou V.A. (2007). Detection and evolution of resistance to the pyrethroid cypermethrin in *Helicoverpa zea* (Lepidoptera: Noctuidae) populations in Texas. Environmental Entomology 36 (5), 1174-1188. https://doi.org/10.1603/0046-225x(2007)36[1174:dae ort]2.0.co;2
- Soberón M. (2016). Reference module in food science: Strategies to reduce insecticide use in agricultural production. Elsevier Reference Collection in Food Science. Elsevier, Amsterdam, Netherlands, 5 pp. https:// doi:10.1016/b978-0-08-100596-5.03063-8
- Tallur P.N., Megadi V.B. and Ninnekar H.Z. (2008). Biodegradation of cypermethrin by *Micrococcus* sp. strain CPN 1. Biodegradation 19 (1), 77-82. https:// doi.org/10.1007/s10532-007-9116-8
- Tamura K., Dudley J., Nei M. and Kumar S. (2007). MEGA4: Molecular evolutionary genetics analysis (MEGA) software version 4.0. Molecular Biology and Evolution 24 (8), 1596-1599. https://doi.org/10.1093/ molbev/msm092
- Tang J., Hu Q., Liu B., Lei D., Chen T., Sun Q., Zeng C. and Zhang Q. (2019). Efficient biodegradation of 3-phenoxybenzoic acid and pyrethroid pesticides by the novel strain *Klebsiella pneumoniae* BPBA052. Canadian Journal of Microbiology 65 (11), 795-804. https://doi:10.1139/cjm-2019-0183
- Tang W., Wang D., Wang J., Wu Z., Li L., Huang M., Xu S. and Yan D. (2018). Pyrethroid pesticide residues

in the global environment: An overview. Chemosphere 91, 990-1007. https://doi:10.1016/j.chemosphere.2017.10.115

- Ullah S., Zuberi A., Alagawany M., Farag M.R., Dadar M., Karthik K., Tiwari R., Dhama K. and Iqbal H.M.N. (2018). Cypermethrin induced toxicities in fish and adverse health outcomes: Its prevention and control measure adaptation. Journal of Environmental Management 206, 863-871. https://doi.org/10.1016/j. jenvman.2017.11.076
- Weston D.P., Holmes R.W. and Lydy M.J. (2009). Residential runoff as a source of pyrethroid pesticides to urban creeks. Environmental Pollution 157 (1), 287-294. https://doi.org/10.1016/j.envpol.2008.06.037
- Weston D.P., Asbell A.M., Hecht S.A., Scholz N.L. and Lydy M.J. (2011). Pyrethroid insecticides in urban salmon streams of the Pacific Northwest. Environmental Pollution 159 (10), 3051-3056. https://doi. org/10.1016/j.envpol.2011.04.008
- Wongsa P., Tanaka M., Ueno A., Hasanuzzaman M., Yumoto I. and Okuyama H. (2004). Isolation and characterization of novel strains of *Pseudomonas aeruginosa* and *Serratia marcescens* possessing high efficiency to degrade gasoline, kerosene, diesel oil, and lubricating oil. Current Microbiology 49 (6), 415-422. https://doi. org/10.1007/s00284-004-4347-y
- Wu P.C., Liu Y.H., Wang Z.Y., Zhang X.Y., Li H., Liang W.Q., Luo N., Hu J.M., Lu J.Q., Luan T.G. and Cao L.X. (2006). Molecular cloning, purification, and biochemical characterization of a novel pyrethroidhydrolyzing esterase from *Klebsiella* sp. strain ZD112. Journal of Agriculture and Food Chemistry 54 (3), 836-842. https://doi.org/10.1021/jf052691u
- Yan D.Z., Li X., Li C.Z., Mao L.Q., Chi X.Q., Zhou N.Y. and Liu D.Y. (2017). Genome-wide identification and characterization of genes encoding cyclohexylamine degradation in a novel cyclohexylamine-degrading bacterial strain of *Pseudomonas plecoglossicida* NyZ12. Journal of Biotechnology 251, 166-173. https://doi.org/10.1016/j.jbiotec.2017.04.036
- Yao G., Jing X., Peng W., Liu X., Zhou Z. and Liu D. (2015). Chiral insecticide α-cypermethrin and its metabolites: Stereoselective degradation behavior in soils and the toxicity to earthworm *Eisenia fetida*. Journal of Agriculture and Food Chemistry 63 (35), 7714-7720. https://doi.org/10.1021/acs.jafc.5b03148
- Zhang C., Jia L., Wang S., Qu J., Li K., Xu L., Shi Y. and Yan Y. (2010). Biodegradation of beta-cypermethrin by two *Serratia* spp. with different cell surface hydrophobicity. Bioresource Technology 101 (10), 3423-3429. https://doi.org/10.1016/j.biortech.2009.12.083
- Zhang C., Wang S. and Yan Y. (2011). Isomerization and biodegradation of beta-cypermethrin by *Pseudomonas*

aeruginosa CH7 with biosurfactant production. Bioresource Technology 102 (14), 7139-7146. https://doi. org/10.1016/j.biortech.2011.03.086

- Zhang H., Mu W., Hou Z., Wu X., Zhao W., Zhang X., Pan H. and Zhang S. (2012). Biodegradation of nicosulfuron by the bacterium *Serratia marcescens* N80. Journal of Environmental Science and Health, Part B 47 (3), 153-60. https://doi.org/10.1080/03601234.2012.632249
- Zhao L., Wang F. and Zhao J. (2014). Identification and functional characteristics of chlorpyrifos-degrading and plant growth promoting bacterium *Acinetobacter calcoaceticus*. Journal of Basic Microbiology 54 (5), 457-63. https://doi.org/10.1002/jobm.201200639